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## Citation

Molina, Ramon M., Siripan Phattanarudee, Jonghan Kim, Khristy Thompson, Marianne Wessling-Resnick, Timothy J. Maher, and Joseph D. Brain. 2011. "Ingestion of Mn and Pb by Rats During and after Pregnancy Alters Iron Metabolism and Behavior in Offspring." NeuroToxicology 32 (4) (August): 413–422. doi:10.1016/j.neuro.2011.03.010.

### **Published Version**

doi:10.1016/j.neuro.2011.03.010

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Author Manuscript

*Neurotoxicology*. Author manuscript; available in PMC 2012 August 1.

Published in final edited form as:

Neurotoxicology. 2011 August ; 32(4): 413-422. doi:10.1016/j.neuro.2011.03.010.

## Ingestion of Mn and Pb by rats during and after pregnancy alters iron metabolism and behavior in offspring

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#### Abstract

Manganese (Mn) and lead (Pb) exposures during developmental period can impair development by direct neurotoxicity or through interaction with iron metabolism. Therefore, we examined the effects of maternal ingestion of Mn or Pb in drinking water during gestation and lactation on iron metabolism as well as behavior in their offspring. Pregnant dams were given distilled water, 4.79 mg/ml Mn, or 2.84 mg/ml Pb in drinking water during gestation and lactation. Pups were studied at time of weaning for <sup>59</sup>Fe absorption from the gut, duodenal Divalent Metal Transporter 1 (DMT1) expression, hematological parameters, and anxiety-related behavior using an Elevated Plus Maze (EPM) test. Metal-exposed pups had lower body weights and elevated blood and brain concentrations of the respective metal. Pb-exposed pups had lower hematocrits and higher blood Zn protoporphyrin levels. In contrast, Mn exposed pups had normal hematological parameters but significantly reduced Zn protoporphyrin. Pharmacokinetic studies using <sup>59</sup>Fe showed that intestinal absorption in metal-exposed pups was not different from controls, nor was it correlated with duodenal DMT1 expression. However, intravenously injected <sup>59</sup>Fe was cleared more slowly in Pb-exposed pups resulting in higher plasma levels. The overall tissue uptake of <sup>59</sup>Fe was lower in Mn-exposed and lower in the brain in Pb-exposed pups. The EPM test demonstrated that Mnexposed, but not Pb-exposed, pups had lower anxiety-related behavior compared to controls. We conclude that gestational and lactational exposures to Mn or Pb differentially alter Fe metabolism and anxiety-related behavior. The data suggest that perturbation in Fe metabolism may contribute to the pathophysiologic consequences of Mn and Pb exposure during early development.

Competing financial interests declaration: The authors declare that there are no competing financial interests.

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#### Keywords

neurotoxicity; biomarker; heavy metals; iron metabolism; anxiety; metal interaction

#### 1. INTRODUCTION

Neurodevelopment begins *in utero* and continues throughout the postnatal period. This period of rapid development is highly susceptible to metal toxicity. Neurocognitive deficits caused by Mn or Pb exposure during developmental period have been reported in humans and animals. Cognitive deficits in Pb-exposed children were demonstrated by lower intelligence quotient (IQ) scores (Al-Saleh et al., 2009, Coscia et al., 2003, Dietrich et al., 1991). Pb exposure may also contribute to attention deficit hyperactivity disorder (ADHD) (Brondum, 2009). Animal studies have reported comparable changes in cognitive and behavioral functions following Pb exposure. Exposure of rats to Pb during gestation (Yang et al., 2003), or during both gestation and lactation (Kuhlmann et al., 1997), led to long-term cognitive deficits in adult offspring. In addition, they found that animals exposed during both gestation and lactation via the mother showed significant behavioral effects even when tested as adults, when tissue Pb levels had returned to control values. In contrast, animals exposed only after weaning failed to demonstrate behavioral alterations despite elevated tissue Pb levels (Kuhlmann, McGlothan, 1997).

Increases in biomarkers of Mn exposure in humans such as hair manganese concentration have been associated with learning disabilities, hyperactivity, and attention deficits (Barlow, 1983, Collipp et al., 1983). It is known that in adults, high Mn exposure produces behavioral changes, such as emotional lability, memory loss, aggressive behavior, impairments in attention, as well as tremor, bradykinesia, postural deficits, and dystonia. In contrast to the adult brain, neonatal and growing brains are more susceptible to Mn toxicity. Absorption of Mn from the gut is generally high in the first weeks of life, followed by a steady decrease with age (Cawte, 1985). Moreover, excretion of Mn is minimal during the first 18 days of life (Cawte, 1985). These factors contribute to the considerably greater risk of neurotoxicity for neonates when exposed to excess Mn. The behavioral effects of Mn during early development are not fully characterized. Tran et. al used passive avoidance and burrowing detour tests to examine the effects of Mn in rats dosed by gavage during the first 21 days. No significant behavioral effects were reported (Tran et al., 2002b). In another study with rat pups similarly exposed to Mn, no effects on the Elevated Plus maze (EPM) and Morris water maze tests even at 10 mg/ml MnCl<sub>2</sub> exposure level were found (Pappas et al., 1997).

Several transport mechanisms for Mn, Pb and Fe as well as other divalent metals have been identified and reviewed recently (Au et al., 2008). Divalent metal transporter-1 (DMT1) is the major transporter for Pb, Mn, and Fe (Garrick et al., 2003). Iron status modulates expression of DMT1 via transcriptional and/or post-transcriptional regulation (Chung and Wessling-Resnick, 2003). While early development constitutes a critical period of susceptibility to Pb or Mn neurotoxicity, there is also evidence and concern that metal uptake can be altered by iron deficiency. Mn uptake into the brain has been shown to be increased by iron deficiency (Aschner and Aschner, 1991, Thompson et al., 2007) while dietary Fe overload diminishes Mn accumulation in the brain (Aschner and Aschner, 1990, Mena et al., 1974). It has also been reported that Pb and Cd exposure increased DMT1 protein expression in the CNS of growing rats (Gu et al., 2009). During development, the uptake of Fe into the rat brain increases sharply after birth, peaks at postnatal day (PND) 15, and decreases thereafter (Taylor and Morgan, 1990). The finding that common pathways for metal absorption such as that mediated by DMT1 exist predicts that the pattern of DMT1's expression could impact the pharmacokinetics and thus the toxic effects of Mn or Pb

exposures. Several studies have examined the influence of iron status on Mn or Pb neurotoxicity. However, the effects of exposure to Mn or Pb on iron metabolism are not fully understood. Maternal exposures of rat dams to high levels of Mn in diet during lactation results in increase in Mn, Cr and Zn and decrease in Fe brain levels associated with higher DMT1 protein expression in various regions of the brain (Garcia et al., 2006). But whether the altered Fe and other essential metal levels result from changed DMT1 activity in the brain or in the gut is unknown. Pb effects on iron metabolism have been reported. Offspring of iron-replete dams exposed to 3% Pb acetate in food and 0.2% in drinking water were intravenously injected with Tf-bound <sup>59</sup>Fe. The results showed no effects on brain and kidney iron uptake and enhanced uptake in the liver (Crowe and Morgan, 1996).

Therefore, the effects of gestational and lactational Mn or Pb exposures on Fe metabolism in young pups are not fully understood. This is a critical time period when Mn and Pb neurotixicity occur as well as an important window for neural development and high iron demand. While studies have examined the effects of these metals on brain development and brain iron transport, their effects on iron transport in the gut and clearance from the blood is unclear. Both of these factors would contribute to the overall amount of metal available for use by the CNS. The relationship between Mn or Pb exposure, iron status, and behavioral effects has not been established. More information is needed on the molecular mechanisms involved. We predict that Mn and Pb will alter iron status. In this study we evaluated the offspring of Mn and Pb exposed dams for 1) disruptions in iron metabolism such as altered iron status, ZPP, <sup>59</sup>Fe kinetics, and intestinal DMT1 regulation, and 2) neurobehavioral outcomes, using anxiety-related responses in an EPM test.)

#### 2. METHODS

#### 2.1 Animals

The Institutional Animal Care and Use Committee at Harvard Medical School and at the Massachusetts College of Pharmacy and Health Sciences (MCPHS) approved protocols used in this study. Sexually mature male (10 wks old, ~300 g) and female (8 wks old, ~220–240 g) Sprague-Dawley rats were purchased from Charles River Laboratories (Wilmington, MA) and housed at the MCPHS animal facilities (68–70°F, 50% humidity, 12-hour light/dark cycle with lights off at 7 PM). The rats were allowed to acclimate to the facility for at least 2 days prior to starting the experiment. Animals had *ad libitum* access to in-house reverse osmosis (RO) deionized water and food (Purina Rodent Lab Chow # 5001). The diet, drinking water solutions and RO water were analyzed for Mn or Pb by inductively-coupled mass spectroscopy (ICP-MS).

#### 2.2 Manganese and lead

Manganese (II) chloride tetrahydrate (molecular weight 197.91;  $98\pm\%$ , A.C.S. reagent) and lead (II) acetate trihydrate (molecular weight 379.33;  $99\pm\%$ , A.C.S. reagent) were obtained from Sigma Aldrich (Allentown, PA). The metal solutions were prepared as needed with reverse-osmosis (RO) deionized water, every 2–3 days.

#### 2.3 Experimental design

Mating of rats (1 male: 2 females) was conducted overnight in standard polycarbonate cages. Females were examined daily after cohabitation for the presence of sperm in the vaginal tract. The observation of a sperm-positive smear was considered evidence of successful mating and the day was designated gestation day 1 (GD 1). The sperm-positive females were then individually housed, and were exposed to metals via the drinking water. Presumed pregnant dams were randomly assigned to the following exposure groups (n=3 dams/group) from GD 1 until postnatal day 24 (PND 24): 1) RO water (control), 2) 4.79 mg/

ml Mn, or 3) 2.84 mg/ml Pb in RO water. Offspring were culled to 12 per dam at PND 2 with as close to an equal male to female ratio as possible. At PND 24, 8 pups/dam (4 males and 4 females) were tested using the EPM apparatus. At PND 25, <sup>59</sup>Fe absorption kinetics, intestinal expression of DMT1, and tissue levels of Mn and Pb and hematological parameters were studied in the same pups and the rest of the litters.

#### 2.4 Elevated Plus Maze (EPM) tests

Anxiety-related responses were evaluated in 24 pups/group at PND 24 using the EPM. The apparatus consisted of 2 opposing open arms (31.5 cm  $\times$  10 cm) and 2 opposing enclosed arms (31.5 cm  $\times$  10 cm  $\times$  15 cm). These arms extended from a central platform of 10 cm<sup>2</sup>. The EPM, which was painted black, was elevated 30 cm above the floor. Pups were placed individually in the center of the apparatus facing an open arm and were allowed to freely navigate the maze for 5 minutes. The EPM was cleaned with 70% alcohol before testing each rat, and all EPM testing was done between 1 and 4 PM. Arm entries were counted when more than half of the rat body entered the arm. A video-tracking system (Ethovision Data Acquisition Software v. 1.95, Noldus Information Technology, Sterling, VA) was used to record the movement of each pup within the maze. Parameters monitored were: latency of the first entry into the open arm (in sec), frequency of open arm entry (number of times in 5 min), duration in open arms (as time and %), and velocity (mm/sec).

#### 2.5 Pharmacokinetics of <sup>59</sup>Fe

At PND 25, after the EPM experiment was performed, the same pups and the rest in the litters were used to determine the effects of maternal exposure to Mn or Pb on Fe metabolism. We sought to determine if iron absorption in the gut, blood clearance and tissue distribution are affected in metal-exposed pups. <sup>59</sup>FeCl<sub>3</sub> was purchased from Perkin Elmer (Boston, MA) and diluted with 1:50 molar excess of ascorbic acid immediately prior to the experiment to reduce  ${}^{3+}$ Fe to  ${}^{2+}$ Fe. A total of 12 pups from each group (4/dam) were dosed by gavage with <sup>59</sup>Fe in this buffer at 1.5 ml/kg volume dose and equivalent radiation dose of 150 µCi <sup>59</sup>Fe/kg. Each pup was anesthetized with up to 4% vaporized isoflurane (Halocarbons Lab, North Augusta, SC) prior to gavage dosing. Blood samples were sequentially obtained from the tail vein over a 24-hour period (5, 13, 30, 60, 90, 120, 240, 480, 1440 m). Plasma and red blood cells were separated for radioisotope analysis. Since the blood levels of <sup>59</sup>Fe represent the amount absorbed from the gut minus the amount cleared from the circulation, another set of 12 pups/group was intravenously injected with the same dose of <sup>59</sup>Fe via the tail vein. Blood samples were similarly obtained from the contra-lateral tail vein over a 24-hour period. At 4 or 24 hours post-dosing with <sup>59</sup>FeCl<sub>3</sub>, 6 pups/group were humanely killed with overdose of isoflurane anesthesia, exsanguinated via the abdominal aorta, and tissue samples collected. The gastrointestinal tract (GIT) was divided into segments after removal of the luminal contents. Radioactivity of samples of tissue, plasma, and red blood cells was measured using a WIZARD 1410 gamma counter (Perkin Elmers, Waltham, MA). Data were analyzed and expressed as tissue concentration ( $\mu$ Ci/g) and % of dose. At the end of the experiment, duodenal scrapings were collected from each pup for DMT1 analyses. Separate blood samples from remaining pups were taken for Zn protoporphyrin (ZPP) measurements.

#### 2.6 Pharmacokinetic analysis

Only five blood samples of approximately 100  $\mu$ L each were taken from each pup to minimize blood loss and changes in hematocrit. All pups were sampled at 5 min post-dosing. In order to generate a continuous blood profile over 24-hour period, each experimental group was divided into two subgroups, one including pups sampled up to 4 h, and the other group sampled up to 24 h. The pups within each subgroup were then ranked, in decreasing order of plasma <sup>59</sup>Fe level at the earliest sampling time (5 min), in both the 4-h

and 24-h subgroups. The ranked pups from each subgroup were combined to generate "sets" that consisted of the 9 sampling points up to 24 h. Pharmacokinetic (PK) analyses were performed for each set to compare PK parameters for <sup>59</sup>Fe among three exposure groups using the WinNonlin Software version 5.2 (Pharsight Corp, Mountain View, CA). PK parameters such as half-life (time during which ½ of <sup>59</sup>Fe dose is cleared from the plasma), total area under the plasma concentration-time curve (AUC) (an estimate of <sup>59</sup>Fe in the plasma), total body clearance (CL) (rate of loss of <sup>59</sup>Fe from the body), and bioavailability (F) (the extent of absorption of <sup>59</sup>Fe from the gut) were calculated. Half-life was calculated as 0.693 divided by the slope of the regression line between 4 and 24 hours from a plot of the natural log of <sup>59</sup>Fe plasma concentration vs. time. A non-compartmental analysis was performed to quantify the degrees of exposure and loss of iron in the pups. AUC was calculated by the trapezoidal rule-extrapolation method. CL was computed by dividing the administered dose by AUC (Kim et al., 1997). F was then calculated as the ratio of AUC post-gavage to AUC post-IV injection.

#### 2.7 Tissue metal analyses

After the <sup>59</sup>Fe kinetics experiments, the whole blood and one half (left side) of the brain samples used for radioactivity measurements were analyzed for Mn and Pb concentrations. Samples were weighed and nitric acid (OmniTrace Ultra High Purity, EM Science, Kansas City, MO) was added to the samples that were then digested overnight in a clean room hood after which they were diluted with deionized water. Analyses were performed by ICP/MS (Inductive Coupled Plasma-Mass Spectrometry: Elan-6100 DRC, Perkin-Elmer, Waltham, MA). Calibrations with external standards were run each day of measurement.

#### 2.8 Western blot analysis of duodenal membrane protein

The proximal duodenal sections from each pup used in the <sup>59</sup>Fe kinetics experiment were rinsed with 2X protease inhibitors (Complete Mini; Roche Diagnostics, Indianapolis, IN) diluted in PBS. Then a duodenal scraping was collected in an Eppendorf tube and snap frozen in liquid nitrogen. The scrapings were homogenized on ice in ice-cold breaking buffer (20 mM HEPES, pH 7.4, 100 mM KCl, 85 mM sucrose, 20 µM EGTA, pH 8.0) containing 2X protease inhibitor. After centrifugation at 3,500 rpm at 4°C for 10 min, supernatant fractions were centrifuged at 95,000 rpm at 4°C for 10 min. The resulting pellet was washed and resuspended in breaking buffer. Triton X-100 was added to a final concentration of 1%, and samples were gently rotated at 4°C for 1 h to solubilize membrane proteins. After further centrifugation at 95,000 rpm at 4°C for 10 min, the detergentsolubilized protein fraction was collected in the supernatant. Twenty ug of solubilized duodenal membrane proteins were electrophoresed on a 10% SDS-polyacrylamide gel and transferred to Immobilon-Fl polyvinylidene difluoride membrane (Millipore, Billerica, MA) in SDS-free transfer buffer (20 mM Tris, 150 mM glycine, 20% methanol). The membrane was blocked in Tris-buffered saline (TBS; 50 mM Tris, 150 mM NaCl, pH 7.4) containing 5% nonfat dry milk and incubated in rabbit anti-DMT1 antibody (a gift of Dr. Philippe Gros, McGill University, Montreal, QC, Canada) diluted 1:1,000 in TBS containing 1% nonfat dry milk. The immunoblot was also incubated with mouse anti-Na/K ATPase (DAKO North America) diluted 1:20,000 in the same buffer as a loading control for the membrane protein preparation. There were no statistical differences in Na/K ATPase levels among the three exposure groups, validating its use in these experiments. Immunoreactive proteins were visualized and quantified using a LI-COR Odyssey Infrared Imaging System with secondary antibodies goat anti-rabbit IR800 and goat anti-mouse IR680 (LI-COR Biosciences, Lincoln, NE). DMT1 levels were normalized to the constitutively-expressed membrane protein Na/K ATPase.

Aliquots of liver and the other half of the brain (right side) were analyzed for non-heme iron (NHI) using the method of Torrance and Bothwell (Torrance and Bothwell, 1968) and calculated as  $\mu$ g Fe/g of wet tissue. ZPP was measured in selected blood samples at the Children's Hospital Department of Laboratory Medicine (Boston, MA) and expressed as  $\mu$ mol/g.

#### 2.10 Statistical analyses

Body weights, blood and brain metal concentrations, and EPM data were analyzed using mixed effects regression models to estimate differences in exposure while accounting for sex effects and dam-to-dam variability in outcomes (Fitzmaurice GM et al., 2004). Specifically, we included exposure group (C, Mn, Pb) and pup sex as fixed effects and dam as a random effect in the model. This approach has the advantage of providing hypothesis tests of exposure effects that account for dam-to-dam variability, or equivalently litter effects, in an outcome. For the continuous outcomes we used linear mixed effects models whereas for the discrete frequency outcome (total frequency in open arms), we used an overdispersed Poisson mixed effects model. We employed a Bonferroni correction to account for multiple pair-wise tests among the three exposure groups. Data analyses were performed with PROC MIXED procedure using SAS statistical analysis software: SAS Institute, Cary, NC). The Western blot, pharmacokinetic parameters for <sup>59</sup>Fe and other hematological parameters were analyzed using analysis of variance (ANOVA) followed by Bonferroni *post hoc* using SPSS (Statistical Package for the Social Sciences, version 17, SPSS Inc., Chicago, IL). In all statistical tests, *P* values < 0.05 were considered significant.

#### 3. RESULTS

#### 3.1 Effects of ingestion of Mn or Pb in drinking water on dams and pups

For exposure studies, Mn (4.79 mg/ml Mn) and Pb (2.84 mg/ml Pb) in drinking water were provided to dams during pregnancy and lactation. These concentrations were selected based on published data (Brockel and Cory-Slechta, 1999a, Dorman et al., 2000, Lasley and Gilbert, 2000, Pappas, Zhang, 1997) and reflected the maximum doses tolerated by the dams. ICP-MS analysis of the LabDiet 5001 Rodent chow showed that it contains 59  $\mu$ g/g Mn and 0.17  $\mu$ g/g Pb. The amount of ingested Mn and Pb from rat chow was estimated as 2.9 mg/kg/day Mn and 0.008 mg/kg/day Pb based on average food consumption (50 g/kg/ day) (Waynforth and Flecknell, 1992). The amounts in the diet were far lower than the calculated Mn and Pb ingested via drinking water and did not to contribute significantly. The estimated amounts of Mn and Pb consumed by the dams each day based on the measured daily intake of metal-containing water is shown in Table 1. There were time-dependent increases in the amounts of metal consumed daily for both Mn and Pb.

Metal levels in drinking water solutions were determined by of ICP-MS. The mean Mn concentrations in RO water, Mn and Pb solutions were 0.00049, 4790.1, and 15.05 µg/ml, respectively. The mean Pb levels in RO water, Mn and Pb solutions were 0.00003, 7.12, and 2838 µg/ml, respectively. Since there were trace amounts of Mn and Pb in the RO water, control dams were estimated to have ingested 0.000078  $\pm$  0.000003 mg Mn/kg/day during gestation, and 0.000186  $\pm$  0.000012 mg Mn/kg/day during lactation. However, these levels were several orders of magnitude lower than Mn consumed by the Mn-exposed dams during gestation (564.54  $\pm$  36.25 mg Mn/kg/d) and lactation (1255.65  $\pm$  61.21 mg Mn/kg/d) (Table 1). Similarly, the control and Mn-exposed dams ingested far lower amounts of Pb from drinking water. Control dams were estimated to have ingested 0.000007 mg Pb/kg/day during lactation. The Pb-exposed dams during gestation were estimated to have ingested 342.57  $\pm$  28.11 mg Pb/

kg/day, and 744.47  $\pm$  29.24 mg Pb/kg/day during lactation (Table 1). The daily water intake was significantly reduced in the metal-exposed dams. During lactation, the daily intakes of RO water or metal-containing water increased 118–133% compared to during gestation, as a result of the increased fluid needed for adequate milk production.

The average body weights of the dams at the end of gestation or lactation period were not different, except for the Mn-exposed dams at the end of lactation (Table 1). No overt toxicity due to Mn or Pb exposures during gestation was observed in the dams. For example, the gestational term of Mn-exposed ( $22 \pm 0$  days) or Pb-exposed ( $21.3 \pm 0.3$  days) dams was similar to that of the control group ( $21.7 \pm 0.3$  days). Litter size was also not significantly affected by metal exposure. Average litter sizes were  $13 \pm 2$  (control),  $13 \pm 1$ (Mn-exposed), and  $14 \pm 1$  (Pb-exposed). There were no visually overt teratogenic effects in the offspring. However, the weight gain of pups from both Mn- and Pb-exposed dams was significantly lower from birth until the end of the experiments (Figure 1). At PND 25, the body weight of pups exposed to Mn and Pb was 43% and 33% lower than that of the control pups, respectively (P <0.05). The characteristics of pups at PND 25 are summarized in Table 2.

#### 3.2 Effects on blood and brain metal concentrations

Tissue Mn and Pb concentrations were determined by ICP-MS and are shown in Table 2. Statistical analyses using the mixed effects regression models to estimate differences in exposure while accounting for sex effects and dam-to-dam variability showed no significant sex nor dam differences in metal concentrations in blood and brain. It showed that maternal Mn ingestion resulted in significant five-fold increase in blood (14.5  $\pm$  0.5 to 75.6  $\pm$  6.6 ng Mn/g) and a three-fold increase in brain  $(434.9 \pm 5.3 \text{ to } 1278.5 \pm 49.4 \text{ ng Mn/g})$  Mn in the pups. Pb ingestion also resulted in a very large increase in blood ( $5.4 \pm 0.8$  to  $698.1 \pm 78.2$ ng Pb/g) and brain (16.4  $\pm$  5.6 to 1006.2  $\pm$  73.9 ng Pb/g) Pb in the pups. Interestingly, the blood and brain Pb concentrations in Mn-exposed pups were also significantly elevated. Pb was present in the RO water, rat chow, and in the Mn drinking water possibly as contaminant of the Pb acetate. Thus, Mn-exposed dams ingested some Pb from drinking water although at a much lower levels than the Pb-exposed dams (Table 1). Likewise, Mn was also detectable in RO water, chow, and the Pb water. But the Pb brain and blood concentrations in the Pb-exposed pups were not different from controls. Since both blood and brain metal concentrations for each pup were available, these data were analyzed by regression analysis. Blood metal concentrations may serve as a biomarker for brain Mn or Pb retention, and also reflect metal exposure. Blood and brain metal concentrations of pups within each exposure group were analyzed. A significant correlation was found between blood and brain Mn (Figure 2A) but not Pb (Figure 2B). Our data also showed that the brain had a 10-fold higher Mn concentration than the blood.

#### 3.3 Effects on iron status

We measured hematocrits and found significantly lower levels in Pb-exposed but not Mnexposed pups (Table 2). Non-heme iron levels in both the liver and brain were not significantly different in either metal-exposed versus control groups. We found that blood levels of zinc protoporphyrin (ZPP) were elevated in Pb-exposed pups as expected (Labbe et al., 1999). Interestingly, ZPP levels were significantly decreased in the Mn-exposed pups (Table 2).

The effects of gestational and lactational Mn or Pb exposure on iron metabolism were examined by determining the kinetics of <sup>59</sup>Fe absorption from the gut, the disposition of absorbed <sup>59</sup>Fe, as well as the tissue distribution of <sup>59</sup>Fe at 4 and 24 hours after administration. The plasma concentrations of <sup>59</sup>Fe expressed as a percent of administered dose after IV injection or after gavage are shown in Figure 4. The plasma concentrations

of <sup>59</sup>Fe declined bi-exponentially post-IV injection (Figure 4A) with no significant difference in mean half-lives among the three groups (data not shown). Clearance of iv-injected <sup>59</sup>Fe from the blood was significantly decreased in Pb-exposed pups, resulting in greater circulating <sup>59</sup>Fe shown as higher AUC (dose fraction  $\times$  min) than in the control group (Table 3). Mn exposure did not alter iron clearance from the blood. After gavage administration, <sup>59</sup>Fe was rapidly absorbed into the systemic circulation and disappeared exponentially (Figure 4B). These data suggest that Pb-exposure increased plasma levels of <sup>59</sup>Fe regardless of whether <sup>59</sup>Fe was injected or gavaged. Intestinal <sup>59</sup>Fe absorption was not influenced by Pb or Mn exposure.

The tissue concentrations of <sup>59</sup>Fe at 4 and 24 hours were determined and were corrected for radioactive decay. At both time points, the tissue levels of <sup>59</sup>Fe in metal-exposed pups were generally lower than control values except for plasma. Tissue concentrations of <sup>59</sup>Fe ( $\mu$ Ci/g) at 4 hours are summarized in Table 4. Compared to controls, <sup>59</sup>Fe concentrations in the brain, lungs, spleen and small intestine of the Mn-exposed pups were significantly lower in the IV-injected group. The lung and spleen levels in Pb-exposed pups were also lower but the plasma level was significantly higher than in controls. After gavage, the kidneys, bone marrow, spleen, and small intestine <sup>59</sup>Fe concentrations of IV-injected <sup>59</sup>Fe in the Mn-exposed pups were generally lower than control values, tissue concentrations of IV-injected <sup>59</sup>Fe in the Mn-exposed pups were generally lower than control (bone marrow, skeletal muscle and spleen). Similarly, after gavage, <sup>59</sup>Fe concentrations in plasma, bone marrow, skeletal muscle, liver, spleen, and large intestine of the Mn-exposed pups were also significantly lower than in controls (Table 4). In contrast, only the plasma level was significantly higher in Pb-exposed pups were also significantly lower than in controls (Table 4). In contrast, only the plasma level was significantly higher in Pb-exposed pups were also significantly lower than in controls (Table 4). In contrast, only the plasma level was significantly higher in Pb-exposed pups after <sup>59</sup>Fe IV injection.

#### 3.4 Effects on anxiety-related behavior

Consequences of gestational and lactational exposure to Mn and Pb were evaluated in the EPM test. Behavioral responses observed included the latency of the first occurrence to the open arm, frequency of open arm entry, and duration in open arms (as time and percentage). These parameters were used as an index of anxiety/anti-anxiety behavior (Pellow et al., 1985). Using the same mixed effects regression models to estimate EPM differences among exposure groups while accounting for sex effects and dam-to-dam variability, we found no significant sex nor dam effect in all EPM parameters. The analyses showed that pups from Mn-exposed but not from Pb-exposed dams had significant decreases in latency to enter an open arm, increased frequency of open arm entries, and increased overall time spent in the open arms (as duration in open arm, time and %). Figure 3 shows combined data for males and female pups.

#### 3.5 Effects on intestinal DMT1 expression

Western blot analyses examined the effects of gestational and lactational exposure to Mn or Pb on the expression of DMT1 in the duodenum. The ratios of DMT1 to Na/K ATPase signals are summarized in Figure 5. We found that pups exposed to Mn had increased duodenal expression of DMT1 protein while Pb exposure did not affect DMT1 levels.

#### 4. DISCUSSION

We conducted this study to evaluate the effects of maternal ingestion of Pb or Mn via drinking water on their offspring at the time of weaning. We hypothesized that these metals may also exert their neurobehavioral effects indirectly through altering iron status, a well known cause of neurobehavioral dysfunction (Lozoff et al., 1991). All analyses were performed at the time of weaning to identify early effects of *in utero* and lactational

exposure to metals. We designed the exposure to include both gestation and lactation since these periods represent critical windows for Mn and Pb neurotoxicity. The concentrations of Pb (2.84 mg/ml) or Mn (4.79 mg/ml) selected were within the ranges reported to increase tissue metal levels and impair behavior in animal studies (Brockel and Cory-Slechta, 1999b, Dorman, Struve, 2000, Lasley and Gilbert, 2000, Pappas, Zhang, 1997). Although these doses were higher than likely human exposure, no overt toxicity was observed in pregnant dams as evidenced by normal gestational term and litter size. No teratogenic effects were noted. However, the weights of Mn-exposed dams were significantly lower at the end of lactation, an indication of some toxic effects of this oral dose of Mn (Table 1). The body weights of Pb-exposed dams were not different from controls despite their similar reduced water consumption (Table 1). More significant effects on body weights were observed in pups of both metal-exposed groups. For Mn, this weight effects may relate to Mn disruptive effects on thyroid hormone homeostasis (Soldin and Aschner, 2007).

Both Pb and Mn translocated to the pups during in utero and lactation periods. Metal concentrations in the blood and brain of pups were all elevated compared to their controls. Although the placental or milk levels of Mn or Pb were not measured, the transfer of Mn across the placenta and the excretion of Mn into milk was previously demonstrated (Dorman et al., 2005a). Dorman *et al.* (Dorman et al., 2005b) showed that exposure to inhaled  $MnSO_4$ (0.5 and 1 mg Mn/m<sup>3</sup>) during gestation (6 hr/day, 7 days/week) increased Mn level in the placenta, and increased Mn levels in the brains and lungs of the dams. Kontur and Fechter (1985) reported elevated forebrain and hindbrain Mn concentrations in newborn rats, whose dams were exposed to high levels of  $MnCl_2$  in their drinking water (5 and 10 mg  $MnCl_2$ / mL) throughout gestation. The excretion of Mn into the milk was evidenced by the increased Mn level in the milk of lactating dams exposed to a high dose  $MnSO_4$  (1 mg Mn/m<sup>3</sup>) by inhalation (Dorman, McElveen, 2005a). Milk collected on PND18 from dams exposed to  $MnSO_4$  (1 mg Mn/m<sup>3</sup>) had Mn concentrations that were significantly higher than controls. The availability of both blood and brain metal concentrations for each pup allowed for additional regression analyses. This is useful since blood metal concentrations may serve as a biomarker for brain Mn or Pb retention, and also metal exposure. A significant correlation was found between blood and brain Mn but not for Pb. The difference perhaps is due to the fact that Mn is a nutrient and is normally regulated, while Pb has no known biological role in metabolism and may be sequestered into depot sites in peripheral tissues (e.g., bone).

*In utero* and lactational transport of Pb has been previously reported in rats (Rodrigues et al., 1996). Pb was undetectable in the brain of day-old pups although present in the blood. By PND 23, Pb accumulated in the blood, kidneys and cerebral cortex. In addition, this study measured the Pb levels in milk from the stomach contents of 1-day-old rats. The Pb levels in the milk of Pb-exposed rats were significantly higher than those in controls confirming that Pb is excreted into the milk. The relative contribution of gestational versus lactational exposure to metal tissue levels could not be determined with our experimental design since pups were exposed during both developmental periods. Unpublished data from our laboratory showed that the percent transport of <sup>54</sup>Mn to the pups relative to maternal dose during lactation (PND 21) was approximately twice of that during gestation (PND 1). Our data suggest that Pb and Mn transport and/or clearance rates were different, since retained Pb (130-fold in blood, 60-fold in brain) was much higher than Mn (5-fold in blood, 3-fold in brain) relative to control. Interestingly, Pb transport appears to be influenced by Mn since pups from Mn-exposed dams had significant elevations of blood (2.5-fold) and brain (1.3fold) Pb levels, suggesting in vivo interaction between Mn and Pb. This may be due to the fact that Mn and Pb share at least one transporter, DMT1 (Garrick, Dolan, 2003). The precise pathways of Pb and Mn transport via the placenta or the uptake and secretion from the mammary glands into the milk are unknown. That DMT1 may be involved is suggested

by the presence of DMT1 in the placenta and in the mammary gland epithelium (Garrick et al., 2006, McArdle et al., 2008).

Elevated levels of Mn in the brain were associated with behavioral responses indicative of lessened anxiety. In the EPM experiment, observed tasks of shorter latency of first entry into an open arm, an increased frequency of open arm entries, and increased duration of time spent in the open arms indicate anti-anxiety behavior (Walf and Frye, 2007). It has been reported that rats with less anxiety or disinhibition control tended to have shorter latency of first entry into an open arm, and also spend more time in the open arms, perceived as danger zones (Ueno et al., 2002, Walf and Frye, 2007). Several theories have been proposed as underlying mechanisms responsible for this behavior. Frontostriatal systems have been implicated in response inhibition. Alterations within this region are thought to produce impulsivity associated with a number of psychological disorders (Itami and Uno, 2002, Jentsch and Taylor, 1999). Abnormalities in ventral region of the frontal lobes, such as the orbitofrontal cortex have also been observed in ADHD (Hesslinger et al., 2002, Itami and Uno, 2002). Subcortical structures dysfunction within the basal ganglia may account for some of the deficits in impulse control observed in the ADHD children (Vaidya et al., 1998). In addition, there is evidence that dopamine alterations are important in ADHD syndrome such as in dopamine transporter (DAT) (Dougherty et al., 1999), dopamine decarboxylase activity (Ernst et al., 1998), and dopamine D<sub>4</sub> receptor gene polymorphism (LaHoste et al., 1996).

Despite the greater increase in brain Pb concentration over control, Pb-exposed pups exhibited no change in anxiety-related EPM parameters. These negative findings are in agreement with previous reports. The effects of Pb on behavior determined by the EPM have shown mixed results. Trombini *et al.* (Trombini et al., 2001) reported that exposure to 750 ppm of Pb acetate in drinking water during pregnancy and lactation had no effect on the behavior determined by the EPM. In contrast, Moreira *et al.* (Moreira et al., 2001) reported that developmental exposure to Pb at 500 ppm (during pregnancy and lactation) caused higher anxiety-related behavior, as shown by the animals' increased aversion to the open arms. Decreased anxiety in Mn-exposed rats in our results were in agreement with some (Tran et al., 2002a) but not with other reports (Kern et al., 2010, Pappas, Zhang, 1997, Reichel et al., 2006). Several factors may contribute to these discrepancies, such as exposure parameters and age of testing, as well as differences in the testing apparatus used. Both Mn-and Pb-exposed pups displayed slower velocities consistent with effects on locomotor functions reported earlier for Mn (St-Pierre et al., 2001) and Pb (Moreira et al., 2001).

Iron deficiency has been shown to reduce motor activity and produce anxiety-related behaviors (Beard et al., 2002). Since Mn or Pb may disrupt Fe metabolism by molecular mimicry, we examined the influence of gestational and lactational exposures on iron status. Mn and Pb had different effects on iron status. Total brain non-heme iron content was unaffected, although it is possible that specific regions of the brain were affected by metal exposure. Pb exposure did result in low hematocrit accompanied by elevated ZPP. ZPP is a product of iron-deficient erythropoiesis, and is a normal metabolite formed from heme biosynthesis. Since Pb competes with Fe, the stoichiometric balance shifts so that more Zn than Fe is incorporated into the protoporphyrin ring. Hence, increased ZPP levels are often used as a biomarker to detect lead poisoning in children (Labbe, Vreman, 1999). As expected, ZPP was elevated in Pb-exposed pups. Mn exposure did not change most hematological parameters except for a significant reduction in ZPP. This suggests that the presence of excess Mn may exert a similar effect; perhaps MnPP is formed at the expense of ZPP. Ferrochetalase, the enzyme that inserts iron or zinc into the protoporphyrin ring is known to be competitively inhibited by Mn (Dailey, 1987). Moreover, Mn can be non-

catalytically chelated by protoporphyrin (Verger et al., 1983). Further studies to explore blood ZPP as a potential biomarker of Mn exposure are warranted.

Gu et al. (Gu, Chen, 2009) have reported that exposures to Pb and cadmium synergistically increases DMT1 protein expression in the CNS of developing rats. Since DMT1 is a common transporter for Fe, Mn and Pb, we examined whether gestational and lactational exposure to Mn or Pb modulates DMT1 in the duodenum that may alter <sup>59</sup>Fe absorption kinetics. Intestinal DMT1 expression was elevated in Mn-exposed pups but the gut absorption of Fe was not significantly altered. Since Fe and Mn are competitively transported by DMT1, the lack of correlation between intestinal DMT1 expression and Fe absorption in Mn-exposed pups could in part be explained by competitive inhibition of Fe absorption by ingested Mn from the milk despite elevated DMT1 level in the gut.

Pb exposure did not affect DMT1 or influence <sup>59</sup>Fe absorption from the gut and thus most likely does not affect the observed changes in hematocrit via perturbed iron intake. However, <sup>59</sup>Fe injected directly into the circulation was cleared more slowly, and thus was retained to a greater extent in Pb-exposed than in control pups. Although <sup>59</sup>Fe persisted longer in the plasma, its uptake in most tissues of Pb-exposed pups was generally not different from control. These observations are consistent with interference of Fe incorporation into heme by Pb, resulting in slower clearance of <sup>59</sup>Fe from the plasma and the consequential anemia that was observed.

#### 5. CONCLUSIONS

In toto, our data indicate that gestational and lactational exposure to Mn and Pb alters iron metabolism and behavior in offspring in different ways. Mn exposure resulted in reduced plasma and brain <sup>59</sup>Fe levels, and associated decreased locomotor activity and anxietyrelated behavior. It also resulted in decreased ZPP and enhanced duodenal DMT1 expression. These results suggest that excess Mn alters Fe metabolism resulting in an upregulation of DMT1. Mn may also interfere with iron homeotasis by replacing iron in key enzymes or inhibiting enzymes essential in iron metabolism. The resulting increase in brain Mn and decreased iron uptake most likely contributed to the decreased anxiety-related behavior in the developing rat. Pb exposure did not change <sup>59</sup>Fe absorption from the gut nor DMT1 expression. However, <sup>59</sup>Fe stayed longer in the circulation leading to increased plasma, but decreased brain <sup>59</sup>Fe levels. Pb exposure also caused anemia with increased ZPP, and associated decreased locomotor activity. These data suggest that Pb exposure results in decreased iron incorporation into red blood cells. Even with sufficient dietary iron, Pb excess, causes anemia. This concomitant anemia would result in decreased oxygenation of vital organs, especially the brain, and may exacerbate Pb toxicity. Taken together, Mn and Pb exposure each interferes with iron homeostasis that may exacerbate their neurobehavioral effects. Additional studies to elucidate iron homeostatic pathways specifically disrupted by Mn and Pb exposure in neonates are warranted.

#### Acknowledgments

This work was supported by the National Institute of Environmental Health Sciences (NIEHS) [P01 ES012874, ES000002] and by the US Environmental Protection Agency (EPA) [RD-83172501]. The contents are solely the responsibility of the authors and do not necessarily represent the official views of the NIEHS or EPA. We also thank Brent Coull and Thomas Donaghey for help with statistical analyses, Swati Betharia and Marissa Kakoyiannis for technical assistance, and Melissa Curran for editorial advice.

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#### Figure 1. Weight of pups over time

The body weight of pups exposed to Mn (closed triangles) and Pb (closed circles) was significantly lower compared with control pups (closed squares) at all time points (P < 0.05). At PND 25, the body weight of pups exposed to Mn and Pb was 43% and 33% lower than that of the control pups, respectively. Data are means ± SE. n = 36/group.



#### Figure 2. Correlation of blood to brain metal concentration

Scatter plot of individual brain and blood Mn (A) or Pb (B) levels. Regression analysis showed a significant correlation of brain to blood Mn concentration. \*P < 0.05.



**Figure 3. Behavioral changes in rat pups gestationally- and lactationally-exposed to Mn or Pb** Behavioral test was done on PND 24. Decreased anxiety-related behavior is associated with shortened latency of first open arm entry, frequency of open arm entries and duration of time spent in open arms. Higher average velocity reflects greater locomotor activity. Mn-exposed pups (gray bars) but not Pb-exposed (black bars) displayed significant changes in parameters indicative of lower anxiety than control pups (white bars). Both Mn- and Pb-exposed pups displayed lower overall velocities when compared with control pups. \**P* < 0.05 compared with control. Data are means  $\pm$  SE. n = 24 pups/group.



#### Figure 4. Plasma levels of <sup>59</sup>Fe following IV or gavage administration in pups

Plasma levels of <sup>59</sup>Fe over 24 h period post-intravenous (A) and post-gavage (B) dosing of pups. Figure 4A – Control (closed squares), Mn-exposed (closed triangles), Pb-exposed (closed circles). Figure 4B - Control (open squares), Mn-exposed (open triangles), Pb-exposed (open circles). Pharmacokinetic analyses of both gavage and IV data showed that the bioavailability (extent of absorption) and half-life were unchanged by metal exposure whereas Pb-exposure significantly decreased <sup>59</sup>Fe elimination from the plasma after either IV or gavage administration (See Table 3).



#### Figure 5. Western blot analysis of DMT1 in duodenal membrane fractions

DMT1 was visualized and quantified by using the LI-COR Odyssey Infrared Imaging System and the LI-COR secondary antibodies goat anti-rabbit IR800 and goat anti-mouse IR680 (LI-COR Biosciences, Lincoln, NE). Anti-DMT1 antibody was a gift of Dr. Philippe Gros (McGill University). Na/K ATPase was used as a membrane protein loading control and was detected in all duodenal samples. Shown in Figure 6A is a representative blot. Shown in Figure 6B are means  $\pm$  SE of relative signal intensity ratio (DMT1/Na/K ATPase, n = 12 control, 8 Mn, and 8 Pb). DMT1 was upregulated in Mn-exposed (gray bar) but not in Pb-exposed pups (black bar). \**P* < 0.05 compared with control.

#### Table 1

#### Body weights and estimated Mn and Pb intake of dams from drinking water

	Control (n=3)	Mn-exposed (n=3)	Pb-exposed (n=3)
Weight (g) at	the end of		
Gestation	$418.67\pm29.68$	$354.00{\pm}7.57$	$382.67\pm5.49$
Lactation	$342.33\pm8.57$	$255.00 \pm 11.36^*$	$335.00\pm12.10$
Water consun	nption (ml/kg/day)		
Gestation	$155.75\pm6.40$	$117.85 \pm 13.11 ^{\ast}$	$120.58 \pm 9.91 ^{\ast}$
Lactation	$371.40\pm23.56$	$262.14 \pm 12.78^{*}$	$262.03 \pm 10.30^{\ast}$
Mn intake (m	g/kg/day)		
Gestation	$0.000078 \pm 0.000003$	$564.54\pm36.25$	$1.82\pm0.15$
Lactation	$0.000186 \pm 0.000012$	$1255.65 \pm 61.21$	$3.95\pm0.15$
Pb intake (mg	/kg/day)		
Gestation	$0.000005 \pm 0.0000002$	$0.84\pm0.05$	$342.57\pm28.11$
Lactation	$0.00001 \pm 0.0000007$	$1.87\pm0.09$	$744.47\pm29.24$

Mn and Pb intake were estimated based on measured concentrations in drinking water using ICP-MS and measured water consumption.

 $^{*}P < 0.05$  compared to control, n=3 dams per group.

# Table 2

Characteristics of 25-d-old rat pups exposed gestationally and lactationally to Mn or Pb.

	Control	=	Mn-exposed	=	Pb-exposed	=
Body wt (g)	$66.8\pm1.6$	24	$38.0\pm1.8^{\ast}$	24	$44.6\pm0.08^*$	24
Hematocrit (%)	$33.4\pm0.3$	30	$32.3 \pm 1.0$	24	$27.3\pm0.5^{*}$	30
Liver NHI (µg/g)	$43.4 \pm 3.9$	27	$38.9\pm3.5$	24	$42.6 \pm 3.3$	26
Brain NHI (μg/g)	$7.6 \pm 0.3$	27	$7.1 \pm 0.2$	27	$8.6\pm0.9$	30
Blood Mn (ng/g)	$14.5\pm0.5$	30	$75.6\pm6.6^{*}$	27	$15.8\pm0.5$	30
Brain Mn (ng/g)	$434.9 \pm 5.3$	30	$1278.5 \pm 49.4^{*}$	27	$465.5 \pm 5.6$	30
Blood Pb (ng/g)	$5.4\pm0.8$	30	$13.5\pm3.8^{*}$	27	$698.1 \pm 78.2^{*}$	30
Brain Pb (ng/g)	$16.4\pm5.6$	30	$22.2 \pm 13.7^{*}$	27	$1006.2 \pm 73.9^{*}$	30
Blood ZPP (μmol/g)	$98.2 \pm 2.8$	9	$75.0\pm4.2^{*}$	5	$144.4\pm9.6^{*}$	7
Data are means $\pm$ SE.						

\* P < 0.05 compared to control. NHI = Non-heme iron. ZPP = Zinc protoporphyrin.

#### Table 3

Pharmacokinetic parameters of <sup>59</sup>Fe after intravenous or gavage administration of <sup>59</sup>Fe to rats exposed gestationally and lactationally to Mn or Pb.

Parameters (Unit)	Control	Mn-exposed	Pb-exposed
$AUC_{gavage}\text{, (dose fraction}\times min)$	$41.0\pm1.1$	$27.7\pm3.8^{*}$	$52.5\pm5.1$
$AUC_{iv}\text{,}$ (dose fraction $\times$ min)	$61.3\pm4.1$	$50.3\pm 6.5$	$82.6\pm 3.8^\ast$
F, (%)	$68.3\pm3.6$	$55.6\pm5.7$	$62.8\pm3.7$

Derived from the IV and gavage data were AUC (area under the <sup>59</sup>Fe plasma concentration-time curve). F (bioavailability, the extent of absorption from the gut, AUC<sub>gavage</sub>/AUC<sub>iv</sub>) was computed using both IV and gavage data. Data are means  $\pm$  SE (n = 4–6 pups/group). ANOVA,

\*P < 0.05 compared with control.

# Table 4

Tissue concentrations of  ${}^{59}$ Fe at 4 hours post-administration.

		Intravenous			Gavage	
Tissues	Control	Mn	Pb	Control	Mn	Pb
Plasma	$0.114\pm0.012$	$0.127\pm0.025$	$0.193\pm0.019^{*}$	$0.217\pm0.032$	$0.141\pm0.031$	$0.265\pm0.030$
RBC	$0.999\pm 0.087$	$1.263 \pm 0.221$	$0.832 \pm 0.113$	$0.614 \pm 0.054$	$0.698 \pm 0.217$	$0.468 \pm 0.123$
Brain	$0.014 \pm 0.004$	$0.013\pm0.0007^{*}$	$0.014\pm0.0006$	$0.009\pm0.001$	$0.007 \pm 0.001$	$0.008 \pm 0.001$
Lungs	$0.183 \pm 0.009$	$0.152{\pm}0.004^{*}$	$0.151\pm0.011^{*}$	$0.131 \pm 0.013$	$0.117\pm0.029$	$0.107 \pm 0.023$
Kidneys	$0.133 \pm 0.012$	$0.128 \pm 0.008$	$0.148 \pm 0.015$	$0.094\pm0.009$	$0.067{\pm}0.009^{*}$	$0.087 \pm 0.013$
Bone Marrow	$0.432 \pm 0.012$	$0.394 \pm 0.027$	$0.451 \pm 0.035$	$0.260 \pm 0.023$	$0.136{\pm}0.018^{*}$	$0.240\pm0.042$
Sk Muscle	$0.036\pm0.001$	$0.0401 \pm 0.004$	$0.035 \pm 0.002$	$0.0228 \pm 0.002$	$0.0182 \pm 0.003$	$0.019 \pm 0.003$
Liver	$0.249\pm0.021$	$0.228\pm0.0234$	$0.287{\pm}0.039$	$0.316 \pm 0.083$	$0.283 \pm 0.094$	$0.316 \pm 0.055$
Spleen	$2.554\pm0.153$	$1.371 {\pm} 0.076^{*}$	$1.648 \pm 0.199^{*}$	$1.510\pm0.250$	$0.571 \pm 0.135^{*}$	$1.127 \pm 0.271$
Stomach	$0.056\pm0.009$	$0.053 \pm 0.002$	$0.050 \pm 0.018$	$0.317 \pm 0.214$	$0.447\pm0.125$	$0.422 \pm 0.116$
S Intestine	$0.089 \pm 0.004$	$0.088 \pm 0.007$	$0.097 \pm 0.005$	$0.127\pm0.013$	$0.067{\pm}0.009^{*}$	$0.114 \pm 0.011$
L Intestine	$0.076 \pm 0.005$	$0.071 \pm 0.004$	$0.080 \pm 0.006$	$0.156\pm0.039$	$0.082 \pm 0.019$	$0.086 \pm 0.007$

Data are means  $\pm$  SE of  $\mu Ci/g~59 Fe,~n=6/group.$ 

\* P<0.05 compared with control.

# Table 5

Tissue concentrations of  $^{59}$ Fe at 24 hours post-administration.

		Intravenous			Gavage	
Tissues	Control	Mn	Pb	Control	Mn	Pb
Plasma	$0.007 \pm 0.001$	$0.007 \pm 0.001$	$0.016\pm0.003^{*}$	$0.009 \pm 0.002$	$0.004{\pm}0.001^{*}$	$0.010\pm0.001$
RBC	$2.871\pm0.104$	$2.637 \pm 0.340$	$2.891 \pm 0.142$	$2.048\pm0.189$	$1.818 \pm 0.260$	$1.493 \pm 0.185$
Brain	$0.019 \pm 0.0004$	$0.015\pm0.002$	$0.018\pm0.0009$	$0.014\pm0.001$	$0.010\pm0.002$	$0.010\pm0.0009^{*}$
Lungs	$0.253\pm0.011$	$0.196 \pm 0.027$	$0.228 \pm 0.007$	$0.350 \pm 0.129$	$0.140 \pm 0.018$	$0.122 \pm 0.013$
Kidneys	$0.125\pm0.006$	$0.119 \pm 0.016$	$0.140 \pm 0.005$	$0.089 \pm 0.005$	$0.082 \pm 0.009$	$0.075 \pm 0.007$
Bone Marrow	$0.131 \pm 0.009$	$0.081\pm0.012^{*}$	$0.129\pm0.009$	$0.118\pm0.012$	$0.054{\pm}0.004^{*}$	$0.083 \pm 0.012$
Sk Muscle	$0.043 \pm 0.002$	$0.031\pm0.005^{*}$	$0.046\pm0.004$	$0.029\pm0.002$	$0.016\pm0.002^{*}$	$0.024\pm0.003$
Liver	$0.226 \pm 0.025$	$0.207\pm0.037$	$0.220\pm0.010$	$0.393 \pm 0.067$	$0.210{\pm}0.039^{*}$	$0.279\pm0.101$
Spleen	$0.745 \pm 0.065$	$0.446\pm0.073^{*}$	$0.763 \pm 0.075$	$0.651 \pm 0.101$	$0.338{\pm}0.026^{*}$	$0.446\pm0.050$
Stomach	$0.050 \pm 0.003$	$0.038 \pm 0.006$	$0.047 \pm 0.007$	$0.032 \pm 0.002$	$0.038 \pm 0.008$	$0.028 \pm 0.004$
S Intestine	$0.075\pm0.003$	$0.060 \pm 0.009$	$0.087 \pm 0.007$	$0.067 \pm 0.004$	$0.052 \pm 0.007$	$0.058 \pm 0.006$
L Intestine	$0.056 \pm 0.003$	$0.044 \pm 0.006$	$0.061 \pm 0.003$	$0.057 \pm 0.005$	$0.042{\pm}0.002^{*}$	$0.050 \pm 0.005$
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Data are means  $\pm$  SE of  $\mu Ci/g$   $^{59}Fe,$  n=6/group.

\* P<0.05 compared with control.