Facilitation stabilizes moisture-controlled alpine juniper shrublines in the central Tibetan Plateau

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Facilitation stabilizes moisture-controlled alpine juniper shrublines in the central Tibetan Plateau

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The Tibetan Plateau hosts one of the world’s highest undisturbed alpine juniper shrublines. However, little is known about the dynamics of these shrublines in response to climate warming and shrub-to-shrub interactions. Since growth of shrubline junipers is limited more by moisture availability than by low temperatures, we tested if upslope advancement of alpine juniper shrublines was constrained by warmer temperatures and related recent droughts. We also evaluated whether facilitation among neighboring shrubs, as inferred from spatial analyses, influenced shrubline dynamics. Three rectangular plots crossing the Juniperus pingii var. wilsonii shrubline were sampled at elevations from 4810 to 4917 m a.s.l. near the Nam Co Lake, central Tibetan Plateau. Location of each stem and its diameter at the root collar and age were measured. We reconstructed the spatial and temporal shrubline dynamics during the past 350 years using standard dendrochronological methods. Independent, long-term summer temperature reconstructions also were associated with shrub recruitment. Point-pattern analyses were used to characterize spatial patterns of different size classes of shrubs. The three shrublines showed little long-term changes despite ongoing warming; no upward shift has occurred in the past 100 years. Recruitment was negatively associated with summer temperatures and drought occurrence since the 1920s. Spatial patterns were characterized by clustering at local scales and attraction between the different size classes, suggesting facilitation. We conclude that moisture availability limits the recruitment and elevational advance of junipers in this area of the Tibetan Plateau. Dynamics of alpine shrublines are more
contingent on positive interactions and local environmental factors than on regional climatic variability.

**Keywords:**

Dendroecology; alpine shrubline; climatic warming; conspecific facilitation; drought; Tibetan Plateau.
1. Introduction

A growing body of evidence shows that global climatic warming has been altering the composition, structure, and distribution of ecosystems worldwide (IPCC, 2014). Among these structural changes, the northern or upward expansion of shrublines—the highest latitude or uppermost altitude at which shrubs occur—into arctic tundra or alpine grasslands is often regarded as fingerprint of global climate warming (Sturm et al., 2001; Post et al., 2009). However, little is known about the range shift of alpine shrublines in mid-latitude mountains subjected to predictable seasonal changes in precipitation and moisture availability, even though alpine shrublands are a major community in these treeless regions (Körner, 2003).

Climatic warming is expected to lead to shifts of latitudinal or altitudinal treelines and shrublines (Kullman, 2002; Myers-Smith et al., 2011; Hofgaard et al., 2013) because minimum air and soil temperatures set the frame for the processes at the leading edge of woody plant communities (Fang et al., 2009; Lv and Zhang, 2011; Liu and Yin, 2013). However, in areas subjected to seasonal water deficits, positive effects of warming on local treelines or shrublines could be canceled out without concurrent increases in precipitation that would alleviate heat-induced moisture stress (Daniels and Veblen, 2004; Wahren et al., 2005). For example, where plant growth at alpine treeline or shrubline is limited by moisture availability during the early growing season (Liang et al., 2012, 2014) or the previous winter (Pellizzari et al., 2014), the treelines or shrublines could be at risk of retreat if lack of water leads to growth decline, constrains recruitment, or increases mortality rates.
Local site conditions (e.g., disturbance, topography, biotic interactions) also buffer, nullify, or alter climatic impacts on treelines and shrublines (Callaway et al., 2002; Kikvidze et al., 2005; Case and Duncan, 2014), resulting in heterogeneous and locally contingent responses of nearby ecotones to shared climatic forcing (Harsch et al., 2009). Among those local factors, plant-plant interactions increasingly are thought to be major drivers of treeline and shrubline dynamics (Batllori et al., 2009; Grau et al., 2012). However, surprisingly little information is available about whether such interactions play a role in mediating alpine shrubline dynamics as the climate warms (Dullinger et al., 2011).

One of the world’s highest natural alpine shrublines occurs on the Tibetan Plateau (Wu, 1983; Huang and Zhang, 2011). In these cold, dry, and treeless regions, shrublands are the primary locus of nutrient cycling, water flow regulation, carbon sequestration, and biodiversity (Huang and Zhang, 2011). However, little is known about the impacts of climatic warming on the structure and distribution of alpine shrublines. Juniperus pingii var. wilsonii (Rehder) Silba is a widespread alpine shrub forming the upper shrubline across the central Tibetan Plateau at elevations over 4900 m. The extremely harsh climatic conditions experienced at the upper limit of shrubs lead to a relatively narrow ecotone of less than 100-m between the shrubline and other alpine vegetation devoid of taller shrub species.

Here we assess how regional climatic conditions and local-scale biotic interactions (cf. Callaway et al., 2002) affect dynamics at three high-elevation J. pingii var. wilsonii shrublines. Specifically, we: (1) reconstructed the age structure of individual
shrubs; (2) revealed alpine shrubline dynamics on multi-centennial time scales; and (3) inferred underlying processes of shrubline dynamics from spatial point-pattern analyses of different size classes of shrubs. Our earlier research had shown that the radial growth of *J. pingii* var. *wilsonii* is limited by scarce precipitation rather than by low temperature (Liang et al., 2012). Thus, we hypothesized that on the three shrublines we studied that the upslope advancement of the shrubline would be constrained by recent warming-related droughts. Further, because patchy patterns in alpine shrublands across harsh environments can reflect intraspecific facilitation (Choler et al., 2001; Armas and Pugnaire, 2005), we also hypothesized that positive conspecific interactions would be reflected in clustered spatial patterns of individual junipers.

2. Materials and methods

2.1 Study area and climate

The study area is located near the Nam Co Lake (30° 30’ - 30° 55’ N; 90° 16’- 91° 03’ E; 4725 m a.s.l.) on the northern flank of the Gangdise-Nyainqintanglha Mountains (Zhu et al., 2010). Climatically, this area is in the transition zone between semi-arid and sub-humid conditions. Between 2006 and 2008, the annual average temperature at 4730 m a.s.l. on the southeastern shore of the lake (AWS in Fig. 1a), was 0.4°C. January (−8.4 °C) and July (9.5 °C) were the coldest and warmest months, respectively (Zhang et al. 2011). The mean annual precipitation was 415 mm, ≈85% of which fell between July and October (Zhang et al., 2011). However, the climate
was dry in May and June, when mean monthly precipitation was, respectively, < 15 and 35 mm, and mean monthly evaporation was > 130 and 160 mm (Zhang et al., 2011). Average daily wind speed was low (3.6 m s\(^{-1}\); see also Zhang et al. 2011) and southeasterly dry winds dominated during the growing season. At the meteorological station of Baingoin (4700 m a.s.l.), about 100 km north-west of the Nam Co Lake, monthly mean temperature, monthly total precipitation, and evaporation from 1957 to 2013 illustrated the dry and cold climates in central Tibet (Fig. 2). Temperature reconstructions from different regions on the Tibetan Plateau and based on tree-ring or ice-core proxies identified a warming trend in the area over the past 400 years (Bräuning and Mantwill, 2004; Liu et al., 2005; Liu et al., 2009; Yang et al., 2009) but no significant change in reconstructed annual precipitation over the same interval (Yao et al., 2008).

### 2.2 Study species

*J. pingii* var. *wilsonii* (henceforth “juniper”) is the most widely distributed shrub species on the Tibetan Plateau (Huang and Zhang, 2011). In the central Tibetan Plateau and around the Nam Co Lake, junipers grow as prostrate, multi-stemmed shrubs in cushion-forming patches of multiple individuals on south-facing slopes; isolated individuals are rarely observed. Their maximum crown diameter is ≤ 3 m, they rarely exceed 1.5 m in height (Chen and Yang, 2011), and the stem diameter at the root collar is normally < 20 cm.

On the central Tibetan Plateau, junipers grow along an altitudinal gradient from

Human disturbance at the upper elevational limits of juniper usually is negligible because they are remote and lack of edible grass for grazing yaks; cover of edible grass species such as *Stipa purpurea* Griseb. accounts for < 5% cover at the shrubline ecotone. We did not observe any sign of grazing by yaks while we were doing fieldwork in 2013 or 2014, nor did we detect other evidence of human disturbances, such as remains of charcoal, fire scars, browsing damage, or stumps in the study sites.

Annual growth rate of juniper is very slow; ring widths average 0.29 ± 0.15 mm/yr. Series-sectioning, i.e., comparison of ring-growth series from several sections taken along one shoot of the same individual (methodological details in Kolishchuk, 1990; Wilmking et al., 2012), confirmed that the cambium along juniper stems remains active. As a result, we were unlikely to miss outer rings in basal stem sections (Liang et al., 2012). Despite juniper’s slow growth, annual ring-width series from different individuals can be cross-dated (Liang et al., 2012; see also the robust cross-dating for other similar shrub species reviewed by Myers-Smith et al., 2015).

Radial growth of junipers is limited primarily by low moisture availability in May-June during the year of ring formation (Liang et al., 2012).
2.3 Field sampling

In July and August, 2013, we sampled three rectangular plots (30 m × 120 m) located near the north-eastern shore of Nam Co Lake (Fig. 1b). Two plots (SE1, SE2) faced southeast (plot SE1: 30.89° N, 90.86° E, 4866 m a.s.l., slope 18°; plot SE2: 30.91° N, 90.80° E, 4917 m, slope 13°), whereas the third (SW1) faced southwest (plot SW1: 30.89° N, 90.87° E, 4810 m, slope 20°). Each rectangular plot was located in a topographically uniform part of the shrubline ecotone. The long side of the plot paralleled the maximum slope and extended above the shrubline (see also Camarero and Gutiérrez, 2004). The relative origin for each plot was situated in its lower left corner. Elevations of lower and upper edges of the plots were measured using a GPS calibrated with a barometric altimeter. Locations (as x, y coordinates ± 0.1 m) of the main stem of each individual shrub within each plot were recorded. Two additional variables for each shrub also were recorded: its maximum height and the diameter at the root collar of the thickest stem of each individual. An electronic caliper was used to measure the stem diameter (± 0.01 mm) and tapes were employed to measure shrub height (± 10 cm). We did not find any dead shrubs within any of the plots.

The age structure of shrubs within the three study plots was obtained using dendrochronological methods (Camarero and Gutiérrez, 2004), as have been used to document shrubline changes in recent years (Myers-Smith et al., 2015). An increasing number of treeline studies have explored recruitment dynamics using static age structures, which reflect the balance between survival and mortality rate (Camarero
and Gutiérrez, 2004; Liang et al., 2011). Due to the dense wood and small size of shrub stems, collecting discs, not increment cores, are used frequently in the Tibetan Plateau and the Arctic to measure their growth rings of shrubs and relate shrub growth to climate (see reviews in Myers-Smith et al., 2015). In fact, we broke several increment borers during field investigation. However, because sampling stem discs destroy the stem, we could not collect discs from every juniper in the plot. Rather, we first randomly collected 22, 36, and 33 discs from individual junipers just outside of plots SE1, SW1 and SE1, respectively. These samples were visually cross-dated using characteristic rings, yielding inter-correlations for samples from sites SE1, SW1, and SW2 of 0.62, 0.58, and 0.61, respectively.

Second, we established relationships between age and root-collar of junipers from 80 randomly-sampled individuals in plot SE2. Wood discs were cut from the thickest stem of each individual shrub, as closely as possible to the root collar; disc diameters ranged from 2.0 to 18.3 cm, and ages ranged from 39 to 361 years. Based on linear relationships between stem diameter at the root collar and juniper age (Fig. 3a), ages of all junipers measured in all three plots were estimated to the nearest decade.

2.4 Characterizing shrubline dynamics

We observed that mature shrubs < 1 m tall may be > 250 years old (Liang et al., 2012). Thus, it made little sense to define the location of the shrubline by shrub height because height growth is very slow and unrelated to age (Fig. 3). Rather, changes in
the upper elevation of the shrubline in the three study plots were reconstructed for the
past 350 years (i.e., 1664 - 2013) in 50-year intervals based on mapping junipers and
estimating their establishment dates to the nearest decade (from the relationship
shown in Fig. 3a). Establishment dates were then related to reconstructed summer
temperatures on the Tibetan Plateau (Thompson et al., 2006).

2.5 Point pattern analyses

Spatial statistics are efficient tools to analyze vegetation dynamics using space as
a surrogate to test hypotheses and to infer underlying (and usually unmeasured)
processes (Fortin and Dale, 2005; McIntire and Fajardo, 2009; Wang et al., 2010).
Point-pattern analyses were used to characterize spatial patterns of junipers growing
in the three study plots. We used the $O(r)$ statistic (Wiegand and Moloney, 2004;
Wang et al., 2010) to compare observed spatial patterns with those expected under a
null model of complete spatial randomness (CSR). Aggregated or hyperdispersed
(regular) distributions have values of $O(r)$ higher and lower than $\lambda$, respectively,
where $\lambda$ is the value obtained for $O(r)$ for a CSR pattern of the same sample size. Note
that the $O(r)$ statistic is a scale-dependent probability density function related to a
neighbourhood density (Wiegand and Moloney, 2004). Considering that both the
mean radius of shrub patches and the distances between the nearest patches were no
more than 5 m, respectively, we calculated $O(r)$ at a spatial resolution = 1 m and for
spatial scales ranging from 1 to 10 m, which should identify relevant small-scale
patterns in the study plots. Last, because the plots are environmentally heterogeneous
(in, e.g., soil conditions and microtopography), we used the inhomogeneous version
of the $O(r)$ statistic for all point-pattern analyses. The Programita software was used
to perform all point pattern analyses (Wiegand and Moloney, 2004, 2014).

2.5.1 Univariate point pattern analyses

We used the univariate $O_{11}(r)$ statistic to examine spatial patterns of three
different size classes: size 1 (stem diameter at the root collar ≤ 3 cm); size 2 (3 < stem
diameter ≤ 8 cm); and size 3 (stem diameter > 8 cm). These three size classes
correspond to different age classes, because shrub size was found to be related to
shrub age (Fig. 3a). Specifically, size classes 1, 2, and 3 corresponded to the following
age classes: age ≤ 80 yrs, 80 yrs < age ≤ 150 years, and age > 150 years, respectively.
Such classification agreed well with the distribution of shrub ages at the three study
plots (Fig. 3b). We used a heterogeneous Poisson process as the null model to
simulate CSR. Edge correction was used to ensure that the number of points in an
incomplete circle was divided by the proportion of the area of the circle that lay
within the study plot (Wiegand and Moloney, 2004, 2014). Finally, if the calculated
$O_{11}(r)$ statistic was above or below the upper or lower 99% simulation envelopes
based on 999 Monte Carlo simulations of the original data then the pattern was
considered to be significantly aggregated or hyperdispersed (regular), respectively, at
the analyzed spatial scale.

2.5.2 Bivariate point pattern analyses
We used the $O_{12}(r)$ statistic to study spatial associations between pairs of the three different size classes of shrubs. We assumed that larger shrub individuals (size 3 in this study) could influence smaller individuals (size 1 and 2), but not *vice versa.* Hence, we used an antecedent condition as the null model, which only randomized the locations of small individuals (sizes 1 or 2), while keeping fixed the locations of larger adults (size 3) or mid-size juveniles (size 2), respectively (Wiegand and Moloney, 2004, 2014). In this case, the statistic $O_{12}(r) = \lambda_2 g_{12}(r)$, for which $g_{12}(r)$ is the bivariate mark-correlation function, gives the expected number of points of pattern 2 (seedlings or mid-size juveniles) located at distance $r$ from any point of pattern 1 (mid-size juveniles or adults). $O_{12}(r) = \lambda_2$ corresponds to independent patterns, whereas $O_{12}(r) < \lambda_2$ or $O_{12}(r) > \lambda_2$ corresponds to repulsion (negative association) and attraction (positive association), respectively (Wiegand and Moloney, 2004, 2014). If the $O_{12}(r)$ statistic was above or below the 99% upper or lower simulation envelopes, the pattern was considered to show significant positive or negative associations at those spatial scales, respectively.

3. Results

3.1 Size structure of junipers

Stem diameter classes were right-skewed and unimodal in all plots (Fig. 4). The maximum stem diameters at the root collar were 17.2, 20.2, and 25.5 cm in plots SE1, SW1, and SE2, respectively. Most junipers were < 80 cm tall, but the maximum height of individuals within the three plots ranged from 113 to 150 cm (Fig. 3c, d, e).
3.2 Climate and juniper establishment

The oldest junipers in plots SE1, SW1, and SE2, respectively, were estimated to be 351, 409, and 512 years in (Fig. 3b), but there were only one or two individuals older than 350 years in any of the plots. Reconstructed recruitment in each plot appeared to be common before 1950, but rare thereafter (Fig. 5). The relationship between reconstructed recruitment and reconstructed climate also differed before the 1920s, when recruitment and summer temperature were positively and significantly correlated (SE1, \( r = 0.59 \); SW1, \( r = 0.58 \); SE2, \( r = 0.65 \); in all cases \( n = 33 \) and \( P < 0.001 \)), and after the 1920s, when they were negatively, but not significantly, correlated (SE1, \( r = -0.36 \); SW1, \( r = -0.40 \); SE2, \( r = -0.39 \); in all cases \( n = 8 \) and, \( P > 0.05 \)).

3.3 Shrubline dynamics

Overall, the juniper shrublines in plots SE1, SW1, and SE2 have shifted upwards by 8, 4 and 9 m, respectively, in the past 350 years, but such summary values obscure much of the variability. In plot SE1 (Fig. 6a), the juniper shrubline increased 6 m in elevation during the 1714–1763 period relative to the previous 50 years. It then remained unchanged from 1764 to 1863, but shifted upward again by 1 m between 1864 and 1913 and by another 1 m between 1914 and 1963. In the last 50-year period, no change in shrubline elevation was observed in the data. In plot SW1 (Fig. 6b), the juniper shrubline increased in elevation by 2.0 m (1714–1763), was unchanged
(1764–1813), advanced again by 2 m (1814–1863), and then remained unchanged thereafter. Finally, in plot SE2 (Fig. 6c), the juniper shrubline increased in elevation by 4.0 m (1714-1763), ascended another 5.0 m (1764–1813), and then stopped advancing.

3.4 Spatial patterns of shrub size classes

Considering all size-classes together, junipers in all sampled plots were significantly aggregated at spatial scales of 1–10 m (in plots SE1 and SE2) or 1–7 m (plot SW1). Among the three size classes, aggregated distributions in plot SE1 were detected at spatial scales of 1–10 m (size-classes 1 and 2) or 1–3 m (size-class 3) (Fig. 7a). In this same plot, the three analyzed pairings among the size classes (size-class 1 vs. size-class 2, size-class 1 vs. size-class 3, and size-class 2 vs. size-class 3) were positively associated at spatial scales from 1 to 6 m. In plot SW1, size-classes 1 and 2 were significantly aggregated from 1 to 6 m, but size-class 3 was CSR (Fig. 7b). Bivariate analysis revealed that all size classes were positively associated from 1 to 3 m in plot SW1. In plot SE2, all three size classes were significantly aggregated at spatial scales of 1–8 m (size-class 1), 1–3 m (size-class 2) and 1–2 m (size-class 3) (Fig. 7c). In contrast to the other two plots, however, negative spatial associations were found for SE3 between size-classes 1 and 3 at 1–7 m, whereas other pairs (size-class 1 vs. size-class 2, size-class 2 vs. size-class 3) were significantly positively associated at spatial scales from 1 to 10 m. Overall, spatial associations were stronger between the larger size classes (size 2 vs. size 3) than between the smallest and bigger.
4. Discussion and conclusion

Climatic warming is expected to lead to elevational shifts in treelines and shrublines (Myers-Smith et al., 2011; Hofgaard et al., 2013), but our results illustrate that juniper shrublines on the central Tibetan plateau have been unchanged for the last 50–200 years. On the central Tibetan Plateau, the climate is characterized by elevated radiation levels and evaporation rates and low precipitation values (You et al., 2010), hence limited moisture rather than low temperatures constrain growth of junipers at shrubline (Liang et al., 2012). Climatic warming without concurrent increases in precipitation will increase the evaporative demand and adversely affect shrub growth and shrubline upward expansion. Thus, upslope advancement of the Tibetan shrubline cannot occur under the warmer and drier conditions that have affected the Tibetan Plateau over the last century (Thompson et al., 2006; Zhu et al., 2011).

Shrubline dynamics also depend on the establishment of new individuals (Myers-Smith et al., 2011), and our data illustrate that little or no new recruitment has occurred in our study plots since the 1920s. Future warming (IPCC, 2014) could reduce juniper recruitment even further. Static age structures within treeline ecotones indicate tradeoffs between survival and mortality (Camarero and Gutiérrez, 2004; Wang et al., 2006; Liang et al., 2011), and have also illustrated reduced recruitment at treelines since the 1950s, coincident with warmer and drier climates, in northwest
We suggest that age structures of juniper shrub plots also can provide insight into recruitment conditions on the central Tibetan Plateau. We acknowledge that the inferences on shrubline dynamics focused on recruitment and rates of advance would be more robust if complemented with growth and mortality data because mortality of shrubs can bias recruitment estimates through time. However, dead shrubs were not found within our shrub plots, making it impossible to estimate mortality rates. Further studies on shrub death and decomposition of dead woody material would help to provide estimates of juniper mortality. Since published growth data revealed the dominant role of moisture availability in juniper wood formation (Liang et al., 2012), the growth data in this study do not promote a deeper understanding of shrubline dynamics in this region.

The slow pace of juniper shrubline change is different from the rapid and extensive shrub expansion observed in Arctic sites where radial growth of shrubs is constrained primarily by low temperatures (Myers-Smith et al., 2011; Naito and Cairns, 2011). However, slow expansion of shrubs also has been reported in the relatively xeric environments of central and eastern Siberia, the interior Alaskan tundra, and continental Mediterranean mountains that receive only seasonal moisture (Wahren et al., 2005; García-Cervigón et al., 2012; Frost and Epstein, 2014). It seems reasonable to suggest, therefore, that warming can exert either negative or positive influences on shrub growth and expansion depending on simultaneous moisture availability (Wahren et al., 2005).
In central Tibet, we hypothesize that recent warming has negatively influenced shrubline dynamics by worsening drought stress, because no significant change in precipitation was observed on the central Tibetan Plateau over the past four centuries and severe droughts were frequent from the 1970s through the 1990s (Yao et al., 2012). In fact, if the frequency of drought increases, the juniper shrubline is at increasing risk of down-slope range shifts. Additional climatic changes also could contribute to exacerbate drought stress: warmer winters could lead to a rapid melting of the snowpack and a decrease in the snowpack thickness could induce drought stress in early spring. Finally, other global drivers such as rising atmospheric concentrations of CO$_2$ could increase drought tolerance or growth by increasing water-use efficiency of plants (Ainsworth and Long, 2005). However, neither the growth trends (Liang et al., 2012) nor the recruitment patterns we report here match with those expectations.

Climate plays an important role in determining the uppermost elevational limits of arboreal and woody vegetation distribution at global, continental, and regional scales (Körner, 2003), but specific dynamics and related patterns are contingent on local scale processes (Callaway et al., 2002; Holtmeier, 2009). Mounting evidence has shown that the magnitude of shrub expansion in cold biomes such as Arctic tundra is highly reliant on microsite facilitation (e.g., small-scale geomorphic settings, disturbance regimes, or positive plant-environment feedbacks; Tape et al., 2006; Hallinger et al., 2010; Myers-Smith et al., 2011; Naito and Cairns, 2011; Frost and Epstein, 2014; Hagedorn et al., 2014). The presence of neighboring conspecifics exerts considerable control on further shrub establishment and encroachment.
irrespective of the existence of additional constraining abiotic drivers (Stueve et al., 2011). For example, local microclimatic warming resulting from shrub encroachment is similar in magnitude to regional warming observed over the past century (He et al., 2014). Spatial analysis of junipers in our plots revealed highly clustered distributions and aggregation among different size classes at relatively small scales (<10 m), consistent with the patchy patterns of shrub distribution observed in the landscape around Nam Co Lake (Liang et al., 2012). We infer that such aggregated spatial patterns are produced by conspecific facilitation as has been found in some other treelines (Camarero et al., 2000). Clumping patterns also may stabilize shrubline dynamics by retarding the advance of the ecotone and leading to shrubline stasis (Harsch et al., 2009).

Positive interactions among woody plants in similar harsh and dry environments result from enhanced establishment and growth and reduced evaporation that buffer the extreme temperature ranges and improve soil fertility (Callaway et al., 2002; Kikvidze et al., 2005). Thus, it is not surprising that conspecific facilitation at local scales dominates in juniper shrublines. Positive interactions related either to structural or to growth-form effects may allow shrub populations to persist for hundreds of years by preserving specific regeneration-niche features, even under extremely adverse environmental conditions (Holtmeier, 2009). Enhanced longevity linked to slow growth rates and aggregated shrub patterns could act as stabilizing factors (Crawford, 2008). If juniper shrublands mainly respond to changes in water availability, climatic warming could induce drought stress and lead to increased mortality of formerly
fast-growing individuals (Nobis and Schweingruber, 2013). In central Tibet, such positive interactions could alleviate drought stress and create suitable regeneration niches, leading to enhanced recruitment and upslope shrubline shifts when climatic conditions are favorable. The relatively large amount of recruitment observed prior to the 1920s matches this scenario and supports our hypothesis that local-scale facilitation drives shrubline dynamics.

Acknowledgements

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**Figure legends**

**Figure 1.** Locations of sampling plots (SW1, SE1, and SE2 plots) and the automatic weather station (AWS) around the Nam Co Lake in the central Tibetan Plateau (a). The upper left inset shows the spatial location of the study area on the Tibetan Plateau. Landscape view of the alpine shrubline (plot SW1) located at an elevation of 4795 m a.s.l. (photograph taken by Y.F. Wang) (b).

**Figure 2.** Monthly mean air temperature, monthly total precipitation, and monthly evaporation from 1957 to 2013 in Baingoin (4700 m a.s.l.), about 100 km north-west of the Nam Co Lake.

**Figure 3.** The linear relationship found between age of *Juniperus pingii* var. *wilsonii* shrubs and stem diameter measured at the root collar (a). These data correspond to individuals sampled in plot SE2 (a). The right part in the first row (b) shows the age distribution of shrubs in 50-year classes for the three study plots. The second and third rows (c, d, e) show scatter plots between height and age of shrub individuals in the three study plots (SE1, SW1, and SE2).

**Figure 4.** Distributions of stem diameters (2.5 cm size-class) at the root collar of sampled *Juniperus pingii* var. *wilsonii* shrubs in the three study plots (SE1, SW1, and SE2).
Figure 5. Decadal dynamics of *Juniperus pingii* var. *wilsonii* establishment estimated for three shrublines (SE1, white bars; SW1, gray bars; SE2, dark bars) and reconstructed summer temperature variability presented as Z-scores (see more details in Thompson et al., 2006). Positive and negative Z-scores correspond to warm and cool summer conditions, respectively.

Figure 6. Spatio-temporal dynamics of the *Juniperus pingii* var. *wilsonii* shrubline in plots SE1 (a), SW1 (b), and SE2 (c). Each solid symbol corresponds to an individual shrub that germinated during the period shown at the top and open symbols indicate individuals established during the previous period. Different symbols represent different establishment periods. Note that smaller symbols are used in the last two periods, otherwise they are not visible.

Figure 7. Spatial point patterns and related univariate and bivariate point-pattern analyses of shrubs located in the three study plots (SE1, SW1 and SE2). Three size classes of *Juniperus pingii* var. *wilsonii* individuals (size-class 1, stem diameter at the root collar ≤ 3 cm; size-class 2, stem diameter ≤ 8 cm; size-class 3, stem diameter > 8 cm) were used in the analysis. Figures in the first row show the spatial positions of the three shrub size classes (note that the plot axes are not drawn at the same scale), whereas figures in the second and third rows show the univariate ($O_{11}(r)$ statistic) and bivariate ($O_{12}(r)$ statistic) point-pattern analyses, respectively. Lines with symbols
represent the $O_{11}(r)$ or $O_{12}(r)$ statistics, whereas thin lines correspond to the upper and lower 99% bounds of the simulation envelopes.
Figure 1
Figure 2
Figure 3
Figure 4
Figure 5
(a) SE1

(b) SW1
Figure 6
(a) SE1

- Size 1
- Size 2
- Size 3

- Y (m)

- X (m)

- Size 1-size 2
- Size 2-size 3

- O_{11}(r)
- O_{12}(r)

- Scale (m)

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(b) SW1

![Graphs of Size 1, Size 2, and Size 3 distributions with Size1-size2 and Size1-size3 comparisons.](image)

**Scale (m)**

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Figure 7