Background: Pulmonary embolism (PE) is the most serious manifestation of venous thromboembolism and a leading cause of sudden death. Several studies have suggested associations of venous thromboembolism with short-term particulate matter (PM) exposure; evidence on long-term PM and traffic exposure is mixed.

Objectives: We examined the association of long-term exposure to PM$_{2.5}$, PM$_{2.5–10}$, and PM$_{10}$ (PM with diameter of ≤2.5, 2.5–10, and ≤10 μm) and distance to roadways with overall incident PE and with PE subtypes in a cohort of U.S. women.

Methods: The study included 115,745 women from the Nurses’ Health Study, followed from 1992 through 2008. Incident PE cases were self-reported biennially. Nonidiopathic PE were cases for which the medical record revealed an underlying health condition related to PE (i.e., surgery, trauma, or malignancy); idiopathic PE were cases with no such history. We used spatiotemporal models combining spatial smoothing and geographic covariates to quantify exposure at residential addresses, and Cox proportional hazards models to calculate hazard ratios (HR) and 95% confidence intervals (CIs).

Results: PM$_{2.5}$ averaged over 1 month (HR = 1.22; 95% CI: 1.04, 1.44) or 12 months (HR = 1.17; 95% CI: 0.93, 1.48) was associated with incident PE, after adjusting for known risk factors and PM$_{2.5–10}$. Equivalent analyses restricted to PE subtypes showed a positive association for PM$_{2.5}$ with nonidiopathic PE, but not with idiopathic PE. We did not find evidence of an association between distance to roadways and PE risk.

Conclusions: We provide evidence that PM in the prior 1 and 12 months is associated with PE risk. Our results also suggest that women with underlying health conditions may be more susceptible to PE after PM exposure.

Citation: Pun VC, Hart JE, Kabrhel C, Camargo CA Jr, Baccarelli AA, Laden F. 2015. Prospective study of ambient particulate matter exposure and risk of pulmonary embolism in the Nurses’ Health Study cohort. Environ Health Perspect 123:1265–1270; http://dx.doi.org/10.1289/ehp.1408927

Introduction

Exposure to particulate matter (PM) air pollution has been associated with increased short- and long-term morbidity and mortality from heart disease and stroke (Brook et al. 2004, 2010; Dockery et al. 1993; Hoek et al. 2013; Miller et al. 2007; Pope et al. 2002). One of the plausible mechanistic pathways mediating such effects is hypercoagulability and enhanced thrombosis (Baccarelli et al. 2007; Brook et al. 2004; Utell et al. 2002). Procoagulant abnormalities are a hallmark of venous thromboembolism (VTE), which is the third most common cardiovascular disease after acute myocardial infarction and stroke, with annual incidence of 1–2 per 1,000 persons in developed countries (Oger for the EPI-GETBO Study Group 2000; Previtali et al. 2011; Rosendaal 1999). VTE encompasses deep vein thrombosis (DVT) and pulmonary embolism (PE). DVT refers to the formation of one or more blood clots in one of the body’s large veins, most commonly in the lower limbs; PE, the most serious manifestation of VTE, typically occurs when a portion of the blood clot breaks off and travels to the heart, ultimately lodging in the arteries of the lungs, potentially leading to death and disability (Office of the Surgeon General 2008; Piazza and Goldhaber 2006). PE is a leading cause of sudden death (Kürkciyan et al. 2000). Studies have showed that the mortality rate for patients with PE is 2- to 4-fold higher than that for patients with DVT, and 10–20% of PE patients die within 3 months after the event (Goldhaber et al. 1999; Naess et al. 2007).

Until recently, little attention has been given to understanding the association between PM exposure and incident VTE or PE risk compared with the attention given to understanding the association between PM exposure and incident VTE or PE risk. A Section 508–conformant HTML version of this article is available at http://dx.doi.org/10.1289/ehp.1408927.

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Prospective Study of Ambient Particulate Matter Exposure and Risk of Pulmonary Embolism in the Nurses’ Health Study Cohort

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records were obtained and reviewed by study physicians blinded to exposure status. Imaging was considered diagnostic of PE if a ventilation/perfusion lung scan was read by a radiologist as high probability for PE, or if there was a filling defect on contrast-enhanced computed tomography of the pulmonary vasculature or on catheter-based pulmonary angiography. Idiopathic PE was defined by the absence of recent surgery, major trauma, or active malignancy listed in the medical record. A diagnosis of nonidiopathic PE was made based on the presence of either a) surgery or major trauma within 1 month of PE diagnosis as listed in the medical record, or b) history of active malignancy and self-reported physician-diagnosed PE. Given the severity of diagnosis, followed by extensive treatment, therapeutic monitoring, and clinical follow-up required of patients with VTE, PE cases are likely reported with high accuracy in our population of nurses with high health literacy. This is supported by our ongoing medical record review of PE cases, and by a validation study of 101 self-reported cases of DVT for whom medical records were available. We found that 95 (94%) cases were confirmed, 2 (2%) cases were probable, and only 4 (4%) cases were not confirmed. Unlike PE, DVT diagnosis has not been routinely tracked in the NHS; therefore, the present study focused only on PE cases.

Exposure assessment. Monthly predictions of particulate matter ≤ 10 (PM_{10}), ≤ 2.5 (PM_{2.5}), or between 2.5 and 10 μm (PM_{2.5–10}) were available for all residential addresses from nationwide spatiotemporal models. Details of the model development process have been provided elsewhere (Paciorek et al. 2009; Yanosky et al. 2008, 2009, 2014). Briefly, the generalized additive mixed models (GAMM) incorporated monitoring data from the U.S. Environmental Protection Agency (EPA) Air Quality System (AQS) (U.S. EPA 2015a), the Visibility Information Exchange Web System (VIEWS 2010), the Interagency Monitoring of Protected Visual Environments (IMPROVE) network (IMPROVE 2013), Stacked Filter Unit (a predecessor to IMPROVE) networks (U.S. EPA 2013), the ESRI StreetMap Pro 2007 file, and were obtained from the National Oceanic and Atmospheric Administration GTS database. The predicted PM data were then incorporated into a GIS – derived and meteorological covariates (GIS)–derived and meteorological covariates and spatial smoothing to estimate monthly prediction surfaces from 1989 forward. Because PM_{2.5} monitoring data are lacking before 1999, separate GAMM models were constructed for 1988–1998 and 1999–2007 (Yanosky et al. 2014). The pre-1999 model added extinction coefficients and a simpler spatiotemporal structure to estimate the PM_{2.5} to PM_{10} ratio seasonally, whereas the post-1999 PM_{2.5} model included point-source PM_{2.5} emissions. The models were evaluated using cross-validation; they exhibited little bias (slopes of the regression between predicted and measured of 0.87, 0.77, and 0.76 for PM_{2.5}, PM_{10} and PM_{2.5–10}, respectively) and high precision (mean absolute errors of 1.81, 5.44, and 4.73 μg/m³ for PM_{2.5}, PM_{10}, and PM_{2.5–10}, respectively) (Yanosky et al. 2014).

We used distance from each residence to the nearest road as a proxy for traffic-related air pollution exposure. Specifically, the distance (in meters, m) from each residential address to the closest U.S. Census Feature Class Code A1 (roads with limited access, typically interstates), A2 (major, noninterstate roads), and A3 (secondary roads, typically with more than two lanes) road segments from the ESRI StreetMap Pro 2007 file were determined using a GIS (ArcGIS, version 9.2; ESRI, Redlands, CA). Distance to the closest road was categorized as 0–49 m, 50–199 m, 200–499 m, or ≥ 500 m, based on the distribution of distance in the cohort and exposure studies showing exponential decay in a number of exposures with increasing distance from a road (Karner et al. 2010).

Covariates. To assess potential confounding, time-varying data from the biennial questionnaires were used for the following covariates chosen a priori based on previous associations with either PE or air pollution: age (months), calendar year, season, region of residence (Northeast, Midwest, South, West), body mass index (BMI), smoking status (never, former, or current), pack-years of smoking (number of packs/day multiplied by number of years of cigarette smoking), having a high-risk diet according to the Alternative Healthy Eating Index score (Chiuve et al. 2012), total caloric intake, physical activity (≤ 3, 3 to < 9, 9 to 18, 18 to 27, or ≥ 27 metabolic equivalent (MET) hours/week), parity (nulliparous, 1, 2–3, or ≥ 4 children), menopausal status and hormone use, nonaspirin nonsteroidal anti-inflammatory drug use, warfarin use, multivitamin use, hypertension, coronary heart disease (any history of angina, coronary artery stenosis, or myocardial infarction), and rheumatologic disease (any history of systemic lupus erythematosus or rheumatoid arthritis). Race (white or nonwhite) and spouse’s highest educational attainment (high school, college, or graduate school) obtained from the 1992 questionnaire were included. Medication and vitamin use, hypertension, coronary heart disease, and rheumatologic disease were treated as dichotomous variables. Area-level socioeconomic status at the U.S. Census tract level was adjusted for using median household value and median family income (U.S. Census Bureau 2000).

Statistical analysis. Time-varying Cox proportional hazards models were used to assess the association of PE for each 10-μg/m³ increase in predicted PM_{2.5}, PM_{10}, and PM_{2.5–10} exposures (after assessing linearity with cubic regression splines) and categories of distance to road. All survival models were based on a monthly time scale and were used to estimate hazard ratios (HRs) and 95% confidence intervals (CIs). Person-months of follow-up time were calculated from baseline (30 June 1992) until PE diagnosis, the end of follow-up (30 June 2008), censoring (moving outside the geographic region of interest or loss to follow-up), or death. We estimated the effect of average particulate exposure in the prior 12 months, because previous research in this cohort has shown time period to be the most relevant exposure for cardiovascular outcomes (Puett et al. 2009). In separate models, we also considered average exposure in the prior 1 month, given that short-term exposures may be more relevant for PE (Dales et al. 2010; Martinelli et al. 2012; Office of the Surgeon General 2008). We examined PM_{2.5}, PM_{10}, and PM_{2.5–10} individually in single-pollutant models. In sensitivity analyses, we assessed the robustness of the associations by examining PM_{2.5} and PM_{2.5–10} simultaneously in a multi-pollutant model. Basic Cox models were stratified by age in months and adjusted for year (linear term), region of residence and season. Confounding from covariates was evaluated individually in separate models and also simultaneously in full multivariable models. In addition, we conducted models restricted to after 2000 (when direct monitoring information on PM_{2.5} became available) and models restricted to women who had not moved residence for at least 5 years. All statistical analyses were conducted using SAS version 9.3 software (SAS Institute Inc., Cary, NC).

Results. A total of 115,745 nurses were available for analysis, and selected participant characteristics during follow-up are presented in Table 1. Throughout the follow-up period, most were never or former smokers, white, and had given birth to at least two children. Nearly half of the nurses reported < 9 MET hr/week of physical activity, and were postmenopausal with no or past HRT use. Mean (± SD) levels of PM_{2.5}, PM_{2.5–10}, and PM_{10} exposures in the previous 12 months were 12.6 ± 3.0, 20.8 ± 5.8, and 8.2 ± 4.2 μg/m³, respectively. Most had residential addresses located ≥ 200 m from the nearest road, and only 15% were < 50 m from the nearest road.

Over 1,725,024 person-years of follow-up, we identified 1,091 total cases of incident PE: 323 (29.6%) were confirmed idiopathic, 526 (48.2%) were nonidiopathic (283 in subjects with malignancy, 243 associated with surgery or trauma), and the subtype could not be
Particulate matter and pulmonary embolism risk

The associations for spatiotemporal model-predicted ambient average PM$_{2.5}$, PM$_{2.5-10}$, and PM$_{10}$ in the previous 1 or 12 months with incident PE, and the idiopathic and nonidiopathic subtypes are presented in Table 2. In basic models adjusted for age, calendar time, season, and region of residence, each 10-μg/m$^3$ increase in prior 1-month or 12-month average PM$_{2.5}$, PM$_{10}$, or PM$_{2.5-10}$ was associated with an increased risk of incident PE, and most associations were statistically significant. However, these increased risks were mostly observed within the nonidiopathic subtype for both time periods.

After adding each potential time-varying confounder one at a time to the basic model, the relationships of PM exposures with PE did not change appreciably relative to the basic models (data not shown). Multivariable adjusted models included race, BMI, smoking status, pack-years, diet, total caloric intake, physical activity, husband’s education attainment, parity, menopausal status, hypertension, coronary heart disease, rheumatologic disease, nonaspirin nonsteroidal anti-inflammatory drug use, warfarin use, multivitamin supplement use, median family income in census tract of residence, and median house value in census tract of residence (Table 2). Results from multivariable models were similar to those from basic models (Table 2), as were results from models restricted to after 2000 or to women who had not moved for at least 5 years (data not shown).

Table 3 shows the sensitivity analyses where we considered exposures to PM$_{2.5}$ and PM$_{2.5-10}$ simultaneously. Most HRs for incident PE were modestly attenuated relative to HRs in the single size fraction models, except for the estimate of 10-μg/m$^3$ increment in PM$_{2.5}$ level averaged over the prior 1 month (HR = 1.22; 95% CI: 1.04, 1.44). Equivalent analyses restricted to specific PE subtypes showed that PM$_{2.5}$ exposure averaged over the prior 1 and 12 months, adjusted for PM$_{2.5-10}$, was associated with increased risk of nonidiopathic PE [1-month HR = 1.30 (95% CI: 1.03, 1.65); 12 month HR = 1.20 (95% CI: 0.86, 1.67)], but not with idiopathic PE. Although elevated incident PE risk was linked to PM$_{2.5-10}$ exposure in single size fraction models, the inclusion of PM$_{2.5}$ led to a decrease in HRs, and the associations were no longer statistically significant.

Last, we did not find evidence of an association between distance to roadways and PE risk (Table 4). For nurses living within 49 m of roads compared with those living further away (≥ 500 m), the multivariable adjusted HRs for incident PE, idiopathic PE, and nonidiopathic PE were 1.02 (95% CI: 0.83, 1.25), 0.92 (95% CI: 0.63, 1.33), and 1.20 (95% CI: 0.90, 1.61), respectively.

**Discussion**

In the NHS cohort of U.S. women, we observed an increased risk of incident PE for each 10-μg/m$^3$ increase in prior 1-month (HR = 1.22; 95% CI: 1.04, 1.44) and 12-month (HR = 1.17; 95% CI: 0.93, 1.48) average exposures to PM$_{2.5}$ after controlling for known risk factors and PM$_{2.5-10}$. Equivalent analyses limited to specific PE subtypes showed a statistically significantly positive association for PM$_{2.5}$ with nonidiopathic PE, but not with idiopathic PE. We did not find evidence of an association between distance to roadways and incident PE risk.
To our knowledge, this is the first study to report positive associations of particulate exposures with incident PE risk in a North American population. To date, only three epidemiologic studies have examined the association of long-term exposure to PM with VTE risk, and there has been little consistency in outcomes or exposures across the studies. In the first studies to examine the impact of pollution on DVT, Baccarelli et al. (2008) observed that each 10-μg/m³ increase in PM₁₀ during the prior year was associated with a 70% elevation in DVT risk in healthy subjects in the Lombardia region of Italy. The risk was smaller in women (40% healthy subjects in the Lombardia region of Italy). The association of long-term exposure to PM with VTE risk, and there has been little consistency in outcomes or exposures across the studies. In the first studies to examine the impact of traffic density and distance to roadways and incident PE generally agrees with the finding of an association between distance to road and risk of DVT (Baccarelli et al. 2009). A possible explanation for the discrepant findings is the different primary disease outcomes across studies. In the ARIC and WHI studies, the main outcome was idiopathic and nonidiopathic combined VTE (Kan et al. 2011; Shih et al. 2011), whereas the main outcome in the Italian studies was first incidence of lower limb DVT with or without symptomatic PE (Baccarelli et al. 2008, 2009). We focused our analysis on incident PE, which is the most serious manifestation of VTE. We found evidence of an association between incident PE and PM₂.₅, but not with distance to roadway, a proxy measure for traffic-related pollution exposure. This suggests that other, non-traffic PM₂.₅ pollution exposures (e.g., industrial emissions) might underlie the association of PE with PM₂.₅.

Table 2. Hazard ratios (95% confidence intervals) for pulmonary embolism (PE) and its subtypes with a 10-μg/m³ increase in averaged predicted PM exposure in the 1 or 12 months before diagnosis (1,725,024 total person-years), single-pollutant models, 1992–2008 (n = 115,745).

<table>
<thead>
<tr>
<th>Pollutants</th>
<th>Incident PE (1,091 cases)</th>
<th>Idiopathic PE (323 cases)</th>
<th>Nonidiopathic PE (526 cases)</th>
</tr>
</thead>
<tbody>
<tr>
<td>PM₂.₅</td>
<td>1.24 (1.06, 1.45)*</td>
<td>1.22 (1.05, 1.43)*</td>
<td>1.08 (0.94, 1.23)</td>
</tr>
<tr>
<td>PM₁₀</td>
<td>1.11 (1.01, 1.22)*</td>
<td>1.10 (1.00, 1.20)*</td>
<td>1.08 (0.94, 1.23)</td>
</tr>
<tr>
<td>PM₂.₅–₁₀</td>
<td>1.27 (1.02, 1.57)*</td>
<td>1.24 (1.00, 1.54)</td>
<td>1.18 (1.04, 1.32)*</td>
</tr>
<tr>
<td>PM₁₀</td>
<td>1.23 (1.03, 1.46)*</td>
<td>1.19 (1.04, 1.42)</td>
<td>1.26 (1.01, 1.58)*</td>
</tr>
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<table>
<thead>
<tr>
<th>Pollutants</th>
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</table>

Regression results for 242 (22.2%) of the 1,091 cases with unknown PE subtype are not shown. *Nonidiopathic PE includes confirmed secondary PE, and PE and previous cancer.

Adjusted for age in months, region, year, season, race/ethnicity, BMI (continuous), smoking status, pack-years, Alternate Healthy Eating Index score, total caloric intake, physical activity, husband’s education attainment, parity, menopausal status, hypertension, coronary heart disease, rheumatologic disease, nonaspirin nonsteroidal anti-inflammatory drug use, warfarin use, multivitamin supplement use, median family income in census tract of residence, and median house value in census tract of residence. *p < 0.05.

Table 3. Hazard ratios (95% confidence intervals) for pulmonary embolism (PE) and its subtypes with a 10-μg/m³ increase in averaged predicted PM exposure in the 1 or 12 months before diagnosis (1,725,024 total person-years), multi-pollutant models, 1992–2008 (n = 115,745).

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Regression results for 242 (22.2%) of the 1,091 cases with unknown PE subtype are not shown. *Nonidiopathic PE includes confirmed secondary PE, and PE and previous cancer.

Adjusted for age in months, region, year, season, race/ethnicity, BMI (continuous), smoking status, pack-years, Alternate Healthy Eating Index score, total caloric intake, physical activity, husband’s education attainment, parity, menopausal status, hypertension, coronary heart disease, rheumatologic disease, nonaspirin nonsteroidal anti-inflammatory drug use, warfarin use, multivitamin supplement use, median family income in census tract of residence, and median house value in census tract of residence. *p < 0.05.

Table 4. Hazard ratios (95% confidence intervals) for pulmonary embolism (PE) and its subtypes with proximity to nearest road (1,483,727 total person-years), 1992–2008 (n = 108,337).

<table>
<thead>
<tr>
<th>Exposure</th>
<th>Incident PE (954 cases)</th>
<th>Idiopathic PE (293 cases)</th>
<th>Nonidiopathic PE (454 cases)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Distance to road (m)</td>
<td>Person-years</td>
<td>Cases</td>
<td>Basic</td>
</tr>
<tr>
<td>0–49</td>
<td>224,188</td>
<td>156</td>
<td>1.04 (0.85, 1.29)</td>
</tr>
<tr>
<td>50–199</td>
<td>410,044</td>
<td>250</td>
<td>0.93 (0.78, 1.12)</td>
</tr>
<tr>
<td>200–499</td>
<td>476,700</td>
<td>314</td>
<td>1.03 (0.86, 1.21)</td>
</tr>
<tr>
<td>≥ 500</td>
<td>372,796</td>
<td>234</td>
<td>1.00 (referent)</td>
</tr>
</tbody>
</table>

Regression results for 217 (22.7%) of the 954 cases with unknown PE subtype are not shown. *Nonidiopathic PE includes confirmed secondary PE, and PE and previous cancer.

Adjusted for age in months, region, year, season. Adjusted for age in months, region, year, season, race/ethnicity, BMI (continuous), smoking status, pack-years, Alternate Healthy Eating Index score, total caloric intake, physical activity, husband’s education attainment, parity, menopausal status, hypertension, coronary heart disease, rheumatologic disease, nonaspirin nonsteroidal anti-inflammatory drug use, warfarin use, multivitamin supplement use, median family income in census tract of residence, and median house value in census tract of residence.
A strength of our study was the ability to estimate the relative risks of idiopathic and nonidiopathic PE subtypes separately in association with PM. We found that the PM$_{2.5}$ exposure was associated mainly with nonidiopathic PE, and the association between PM$_{2.5-10}$ and idiopathic PE, though consistently positive, was nonsignificant. This is contrary to our hypothesis that air pollution would be more strongly associated with idiopathic PE cases given the general concept that air pollution induces a state of hypercoagulability and enhanced thrombosis. Although the precise biological pathways involved remain to be determined, our findings suggest a more complicated mechanism: that long-term PM exposure may be additive or interact with preexisting health conditions (in this case, active malignancy, recent surgery, or trauma) to promote PE, but may not participate in determining PE among individuals if they are not in the setting of some other risk factors of PE. Such findings may contribute to collective evidence indicating that certain characteristics, such as life stage, genetic polymorphisms, preexisting diseases, and socioeconomic status, may increase the susceptibility of populations to PM-related health effects (Peled 2011; Sacks et al. 2011). Our ability to examine PE subtypes and the fact that we observed different associations by subtype may help explain the conflicting results in previous studies. If the ratio of idiopathic to nonidiopathic VTE varies between studies, heterogeneity in findings may arise. Other explanations for the differential associations observed by PE subtypes include differential random variability, biases between groups, and fewer idiopathic cases than nonidiopathic cases. Further investigation of the effects of air pollution exposures on PE subtypes is warranted.

In addition to long-term effects of PM exposure, Dales et al. (2010) estimated increased hospital admissions for VTE and PE with daily levels of PM$_{2.5}$ in a time-series analysis in Santiago, Chile, whereas Martinelli et al. (2012) estimated increases in emergency hospital admissions for VTE and PE associated with same day elevations of PM$_{2.5-10}$, but not with PM$_{2.5}$, in Verona, Italy. Although the underlying mechanisms of how particulate pollution may influence VTE pathogenesis (Vircow’s Triad) remain to be determined, it has been suggested that short- and long-term PM exposure enhances thrombotic and coagulation potentials by increasing circulating histamine, inflammatory cytokines, and platelet, plasma levels of coagulation proteins, such as procoagulant factor VIII, von Willebrand factor, and fibrinogen, and shortening prothrombin time (Baccarelli et al. 2007; Brook et al. 2010; Esmon 2003; Nemmar et al. 2006; Rückerl et al. 2006; Uettl et al. 2002).

Several limitations of our study warrant consideration. First, we did not have information on the incidence of overall DVT or VTE, and we were not able to distinguish subtype for 22% of the PE cases. This limits our ability to examine the impact of PM on these other previously examined outcomes, as well as the power to detect statistically significant associations in our subtype analyses. Also, although self-report of PE has been well validated in this population, it is possible that some cases were not reported, leading to an underestimate of the effects of air pollution on PE. Second, our exposure estimates are predictive ambient, not personal, levels of PM. We do not have information on indoor sources, building characteristics that would influence the penetration of ambient exposure, or the amount of time each woman spent at her residence. These factors would all lead to misclassification in exposure, and would likely bias the estimates of association towards the null. Third, because we derived PM$_{2.5-10}$ estimates from those of PM$_{10}$ and PM$_{2.5}$, they are subjected to two sources of random error in measurements, which may lead to lower statistical power. However, Yanosky et al. (2009) demonstrated that long-term average PM$_{2.5-10}$ was reasonably well estimated (cross-validation R$^2 = 0.63$ and 0.65 for long-term post-1999 and pre-1999 PM$_{2.5-10}$, respectively). Third, the GIS-based temporal spatial smoothing of the PM predictive models reduced variability relative to measured concentrations; however, this Berkson measurement error should not cause substantial bias toward the null (Gryparis et al. 2009). Findings from the current study may not be generalizable to nonwhite racial/ethnic groups, to men, to younger women, and to socioeconomically deprived populations. Last but not least, it is possible that our results from this observational study are attributable to confounding. Adjustment for a variety of known confounding variables did not eliminate the observed associations, though there remains the possibility of residual confounding or confounding by unmeasured covariates (e.g., long-distance travel, recent prolonged inactivity) and/or by additional pollutants.

Nonetheless, our study had several strengths, including the ascertainment of incident PE cases and our ability to subtype cases based on detailed medical record review. PE is a reliable diagnosis based on evidence from objective methods such as ventilation/perfusion lung scan; thus the chance of false-positive PE cases is slim. The application of GIS-based spatiotemporal statistical models allowed us to assign exposures specific to biennially updated residential addresses for the entire follow-up period with little bias and high precision relative to other modeling approaches (Yanosky et al. 2008, 2009). Missing exposure data were minimized because of the use of our highly specific GIS-based spatial smoothing model. The availability of updated information on individual-level covariate status (e.g., smoking) throughout the follow-up period, enabled us to take into account possible confounding effects.

Conclusion

We reported evidence of an association between exposure to PM$_{2.5}$ and elevated risk of PE among women in the United States. By contrast, distance to road was not a risk factor for incident PE. Our findings suggest that women with underlying health conditions may be more susceptible to increased PE risk after PM exposure.

REFERENCES


