



# Captive pandas are at risk from environmental toxins

## Citation

Chen, Yi#ping, Lorraine Maltby, Qiang Liu, Yi Song, Ying#juan Zheng, Aaron M. Ellison, Qing#yi Ma, and Xiao#min Wu. "Captive pandas are at risk from environmental toxins." *Frontiers in Ecology and the Environment* 14, no. 7 (2016): 363-367.

## Published Version

10.1002/fee.1310

## Permanent link

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1 **Captive pandas are at risk from environmental toxins**

2

3 Yi-ping Chen<sup>1\*</sup>, Lorraine Maltby<sup>2</sup>, Qiang Liu<sup>1</sup>, Yi Song<sup>1</sup>, Ying-juan Zheng<sup>1</sup>,

4 Aaron M. Ellison<sup>3\*</sup>, Qing-yi Ma<sup>4</sup>, Xiao-min Wu<sup>5</sup>

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6 <sup>1</sup> SKLLQG, Institute of Earth Environment, Chinese Academy of Sciences,  
7 Xi'an, 710075, China.

8 <sup>2</sup> Department of Animal and Plant Sciences, The University of Sheffield,  
9 Sheffield, S10 2TN, UK

10 <sup>3</sup>Harvard University, Harvard Forest, Petersham, Massachusetts, USA 01366

11 <sup>4</sup> Shaanxi Wild Animal Research Center, Zhouzhi, Xi'an, 710402, China

12 <sup>5</sup> Institute of Animal, Shaanxi Academy of Sciences, Xi'an, 710032, China.

13

14 \*Correspondence to: Yi-ping Chen, Institute of Earth Environment, CAS, No.  
15 97, Yanxiang Road, Xi'an, 710061, China; Tel: +86-29-88324766; Fax:

16 +86-29a -88320456; E-mail: chenyp@ieecas.cn or Aaron M. Ellison, Harvard  
17 University, Harvard Forest; email. aellison@fas.harvard.edu

18

19 *Ex situ* conservation efforts are the last resort for many critically endangered  
20 species and captive breeding centers are thought to provide a safe  
21 environment in which to produce individuals for eventual re-introduction to the  
22 wild. The giant panda (*Ailuropoda melanoleuca*) is one of the most  
23 endangered animals in the world, and it is recognized worldwide as a symbol  
24 for conservation. Here, we report that captive pandas of the Sichuan and  
25 Qinling subspecies are exposed to high concentrations of persistent organic  
26 pollutants, including polychlorinated dibenzo-p-dioxins (PCDDs),  
27 dibenzofurans (PDCFs), and biphenyls (PCBs), as well as heavy metals  
28 (arsenic, cadmium, chromium, and lead). Further analysis of the *ex situ*  
29 environment of the Qinling subspecies demonstrated that contaminated food  
30 supplies exposed captive Qinling pandas to high concentrations of PCDD,  
31 PCDFs, PCBs, As, Cd, Cr, and Pb). In the short term, these endangered  
32 animals should be relocated to breeding centers in less contaminated areas.  
33 Their long-term survival, however, depends on reducing emissions of toxic  
34 pollutants throughout China.

35

36 Key words: Endangered species, *ex situ* conservation, Giant panda, heavy  
37 metals, persistent organic pollutants.

38

39 The giant panda (*Ailuropoda melanoleuca*) is one of the most endangered  
40 animals in the world, and it is recognized worldwide as a symbol for  
41 conservation. The panda lineage is at least 11.6 million years old (Abella *et al.*  
42 2012); fossils > 2 million years old and historical records have revealed that  
43 pandas once were distributed in at least 18 of China's 23 provinces (Zhu and  
44 Long 1983). Until the mid-19<sup>th</sup> century, giant pandas still inhabited most of  
45 eastern and southern China (Hunan, Hubei, Sichuan, Shaanxi and Gansu  
46 provinces), but their range has declined in recent years as a result of hunting,  
47 habitat destruction, logging, resource exploitation, and tourism (Zhang *et al.*  
48 2013). Giant pandas now survive only in small, fragmented conservation  
49 zones in the Qinling, Bashan and Qionglai Mountains (Zhang *et al.* 2013) and  
50 in *ex situ* breeding centers including the zoos of Beijing and the breeding  
51 centers of Wolong and Chengdu.

52 It is generally assumed that the conservation areas and the captive  
53 breeding centers protect giant pandas from the adverse impacts of human  
54 activities. However, their presumed safety may be compromised by the  
55 dissemination of widespread pollutants into conservation zones or the  
56 proximity of breeding centers to more heavily-polluted urban areas. For  
57 example, perfluorinated compounds used in consumer and industrial products  
58 as surfactants, surface protectors, and fire-fighting foams have been found in  
59 serum samples taken from giant pandas in the Beijing zoo as well as from red  
60 pandas (*Ailurus fulgens*) in a number of other zoos and wild animal parks in

61 China (Dai et al. 2006). However, the extent to which either wild pandas or  
62 pandas in breeding centers are exposed to persistent organic pollutants (POPs)  
63 and heavy metals that can accumulate in their tissues, compromise their  
64 health, and potentially affect the success of ongoing conservation programs  
65 remains unknown.

66 Here, we present data illustrating that giant pandas in *ex situ* breeding  
67 centers are exposed to much greater concentrations of POPs and heavy  
68 metals than their wild counterparts. Our data suggest that the bamboo fed to  
69 the pandas is the proximate source of these compounds. Consequently, urgent  
70 action is needed to safeguard these conservation icons, both in captivity and in  
71 the wild.

72

### 73 **Materials and Methods**

74 Faecal droppings, which can be used as non-invasive indicators of pollutant  
75 exposure (Christensen *et al.* 2013), were collected from wild pandas in the  
76 Wolong and Foping National Nature Reserves, and from captive pandas  
77 housed in China Conservation and Research Center for the Giant panda  
78 (CCRCGP) and the Shaanxi Wild Animal Research Center (SWARC) (Fig.1).  
79 The CCRCGP is the largest captive panda breeding center for the Sichuan  
80 subspecies of giant panda, and SWARC is the only breeding center for the  
81 Qinling subspecies. Samples of bamboo, the primary food for giant pandas,  
82 were collected in the wild from Foping and from plants grown at SWARC.

83 Mixed feedstuff, fed to pandas as a nutrient supplement, was also sampled  
84 from SWARC. Additional details on sample collection are provided in the  
85 Supplemental Online Material.

86 The faecal droppings, plant tissue, and feedstock samples all were dried  
87 to constant mass, digested, and analyzed using standard methods.  
88 Determination of concentrations of POPs in the samples was done using  
89 (high-resolution mass spectrometry (Liu *et al.* 2006; Li *et al.* 2008) at the  
90 Research Center for Eco-environmental Sciences of the Chinese Academy of  
91 Sciences. Concentrations of heavy metals were determined using atomic  
92 absorption or fluorescence spectrometry at the Institute of Earth Environment  
93 of the Chinese Academy of Sciences. Complete details on analytical methods,  
94 including QA/QC protocols, can be found in the Supplemental Online Material).

95 Data were analyzed using the SPSS software, version 19.0 (IBM SPSS  
96 Inc.). Contaminant concentrations in droppings from wild and captive giant  
97 pandas among and between the two subspecies were compared using *t*-tests.

98

## 99 **Results and discussion**

100 It is generally thought that pandas in captive breeding centers are better  
101 protected from human activities than are wild pandas in nature conservation  
102 zones, primarily because *in situ* conservation zones have become more  
103 fragmented and less suitable for giant pandas (Liu *et al.* 2001). However, *ex*  
104 *situ* breeding centers usually are close to urban areas and there is an

105 increasing concern that *ex situ* conservation efforts may be being  
106 compromised due to environmental pollution associated with urbanization.  
107 With China's rapid industrialization and urbanization, environmental pollution is  
108 increasing in seriousness and following as it follows a trajectory similar to that  
109 previously traversed by developed countries (Seinfeld 2004). This pollution  
110 trajectory is having major impacts on public health, as seen in, for example,  
111 the > 200 "cancer villages" in China (Yang 2013).

112       Among the many pollutants, POPs and heavy metals are of significant  
113 environmental concern because they may be transported over long distances  
114 in air and water (Lohmann *et al.* 2007), are very persistent in the environment,  
115 accumulate readily in fatty tissues, and are highly toxic to humans and other  
116 mammals (*e.g.*, Qiu 2013; Adriano *et al.* 2014; Fernandez-Rodriguez *et al.*  
117 2015; Syed Ali *et al.* 2015). Three classes of POPs – PCDDs (polychlorinated  
118 dibenzo-p-dioxins), PCDFs (polychlorinated dibenzofurans), and PCBs  
119 (polychlorinated biphenyls) were found in much higher concentrations in faecal  
120 droppings of captive giant pandas than in wild pandas (Fig. 2, WebTables1, 2).  
121 POPs were also found at elevated levels in the bamboo fed to captive pandas  
122 and their nutrient-supplement feedstock (WebFigures 1, 2). A variety of forms  
123 ("congeners") of PCDDs and PCDFs are generated as by-products from  
124 various combustion and chemical processes, whereas polychlorinated  
125 biphenyls (PCBs) were widely used as dielectric fluids in transformers and  
126 capacitors, heat exchange fluids, and as additives in pesticides, adhesives,

127 plastics, and paints because of their insulating and nonflammable properties  
128 (Fiedler 2007). Although production of PCBs ceased in 1974, they are still  
129 released from old capacitors and transformers and can still be found in various  
130 environmental components and in human tissues (Mai *et al.* 2005; Imamura *et*  
131 *al.* 2007).

132 Because PCDDs, PCDFs, and PCBs occur as congeners that differ in  
133 toxicity and toxic equivalency factors, the World Health Organization has  
134 defined a single toxic equivalent (WHO-TEQ) that can be calculated to  
135 determine total POP exposure (Van den Berg *et al.* 2006). Both total  
136 concentrations and the WHO-TEQ for PCDDs, PCDFs, and POPs were higher  
137 in droppings collected from captive pandas than they were in wild pandas (Fig.  
138 3). These results are paralleled by total concentrations and WHO-TEQs for the  
139 bamboo fed to the pandas and their nutrient-supplement feedstock  
140 (WebFigure 2).

141 Four heavy metals with known toxicity – arsenic (As), cadmium (Cd),  
142 chromium (Cr), and lead (Pb) (Brahmia *et al.* 2013; Neal and Guilarte 2013;  
143 Uddh-Soderberg *et al.* 2015) – also were found at elevated levels in droppings  
144 of captive pandas relative to wild ones (Fig. 3), as well as in their food and their  
145 nutrient-supplement feedstock (WebFigure 3). Unlike POPs, these heavy  
146 metals occur in the natural environment, but they are readily mobilized by  
147 human activities such as mining, automobile use, and overuse of chemical  
148 fertilizer.

149 Our results provide direct evidence that giant pandas are exposed to  
150 PCDDs, PCDFs, PCBs, and heavy metals in both *ex situ* captive breeding  
151 centers and *in situ* conservation areas, but concentrations of these toxins in  
152 pandas are far greater for pandas in captivity. Previous studies have shown  
153 that PCDDs and PCDFs are associated with developmental toxicity,  
154 immunotoxicity, and reproductive toxicity. PCBs and their breakdown products  
155 are known endocrine disrupters, cause the loss of renal cell viability, and are  
156 associated with increased risk of chloracne, goiter, anemia, and cancer  
157 (Lohmann *et al.* 2007; Qiu 2013; Adriano *et al.* 2014; Fernandez-Rodriguez *et*  
158 *al.* 2015; Gustavson *et al.* 2015; Syed Ali *et al.* 2015). Heavy metal exposure  
159 has been associated with increased incidence of cancer (Cr and As),  
160 nephrotoxicity and bone damage (Cd), and reduced reproductive function (Pb)  
161 (Neal and Guilarte 2013; Brahmia *et al.* 2013; Uddh-Soderberg *et al.* 2015).  
162 We conclude that our results belie the notion that captive breeding centers and  
163 zoos provide a safe haven from human impacts.

164 Our results also illustrate that dietary exposure is the dominant, proximal  
165 pathway through which giant pandas are exposed to POPs and heavy metals  
166 (WebFigures 1-3). Although the food of both captive and wild pandas was  
167 enriched in POPs (WebFigures 1, 2) and heavy metals (WebFigure 3), the  
168 concentrations of both POPs and metals, and WHO-TEQs of POPs were  
169 significantly greater in bamboo eaten by captive pandas (WebFigures 1–3).  
170 We note that the nutrient-supplemented feedstock (baked into steamed bread

171 for the pandas) was enriched only in Cd, Cr, and Pb, but not in As, relative to  
172 fresh bamboo.

173 In sum, our data provide clear evidence that giant pandas both in the wild  
174 and in captivity are exposed to PCDDs, PCDFs, PCBs, and heavy metals  
175 through their diet, and that exposure to these environmental toxins is greater in  
176 *ex situ* breeding centers than in *in situ* nature reserves. Because exposure to  
177 these environmental toxins is likely to impact negatively the health of these  
178 animals, we suggest that urgent action is needed to safeguard these  
179 conservation icons. In the short-term, captive breeding centers should be  
180 relocated to areas less impacted or contaminated by environmental toxins, and  
181 the food provided to captive pandas should be strictly monitored to ensure that  
182 it lacks POPs and heavy metals, and is of consistent high quality. In the long  
183 term, however, a more sustainable solution will rely on improving air quality  
184 through reducing emissions of toxic pollutants.

185

## 186 **Acknowledgments**

187 This work was supported by the IEECAS fund. We thank the China  
188 Conservation and Research Center for Giant panda (CCRCGP) for helping  
189 with this research. Here we also thank Professor An for valuable advice during  
190 the course of this study.

191

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- 259

260 **Figure legends**

261 Figure 1. Sites of sample collection (a); typical dropping of wild pandas (b); and  
262 captive pandas (c) at the Shaanxi Wild Animal Research Center (SWARC).

263

264 Figure 2. Concentrations of 12 PCB congeners (top) and 17 CDD/F congeners  
265 (bottom) in the droppings of wild and captive pandas of the Sichuan and  
266 Qinling subspecies of giant pandas. In each of these star plots, the radius is  
267 equal to the maximum observed concentration, and concentrations of each  
268 individual congeners are scaled to the maximum. The conclusion from these  
269 plots is that captive pandas have both more congeners and higher  
270 concentrations of them in their faecal samples than wild pandas. Tabular data  
271 (actual mean concentrations and the standard errors of the means) are given  
272 in WebTables 1 and 2.

273

274 Figure 3. Concentrations of (a) all (summed) PCDDs and PCDFs; (b) all  
275 (summed) PCBs; (c) WHO-TEQs of PCDDs and PCDFs; and (d) WHO-TEQ of  
276 PCBs in faecal samples collected from two subspecies of wild (blue) and  
277 captive (red) giant pandas. Bars (means  $\pm$  1 SE of the mean from  $N = 4$   
278 independent replicates comprising three or four pooled samples) with different  
279 letters between the wild and captive pandas for the same subspecies (A or B),  
280 or between Sichuan and Qinling subspecies (X or Y) are significantly different  
281 ( $P < 0.05$ , *t*-test).

282

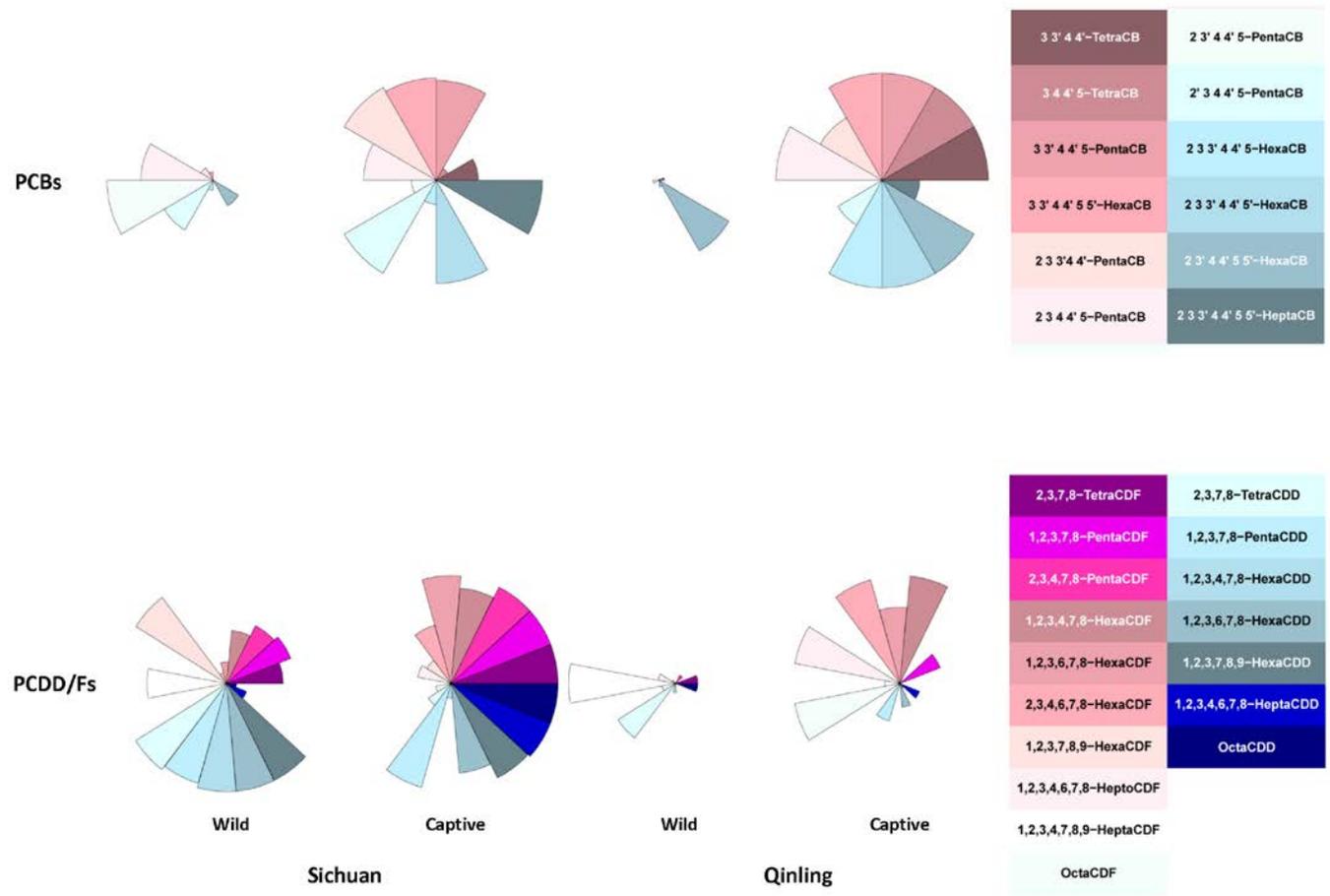
283 Figure 4. Concentrations of heavy metals in faecal samples collected from two  
284 subspecies of wild (blue) and captive (red) giant pandas. (a) Arsenic (As); (b)  
285 Cadmium (Cd); (c) Chromium (Cr); (d) Lead (Pb). Bars (means  $\pm$  1 SE of the  
286 mean from N = 4 independent replicates comprising three or four pooled  
287 samples) with different letters between the wild and captive pandas for the  
288 same subspecies (A or B), or between Sichuan and Qinling subspecies (X or Y)  
289 are significantly different ( $P < 0.05$ , *t*-test).

290

291



Figure 1



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Figure 2

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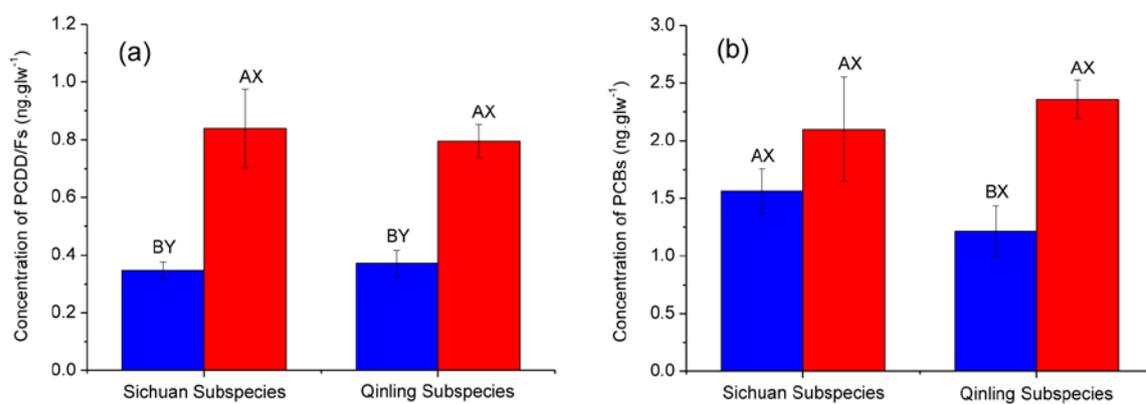
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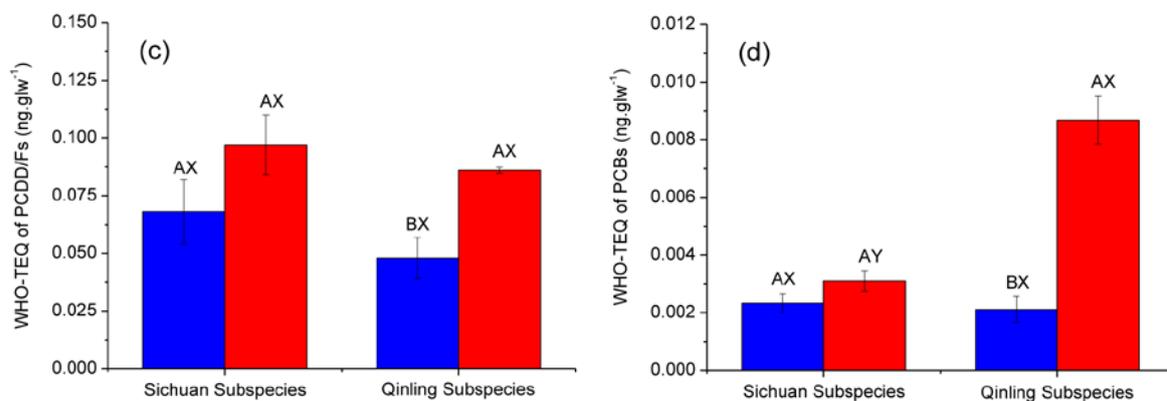
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Figure 3

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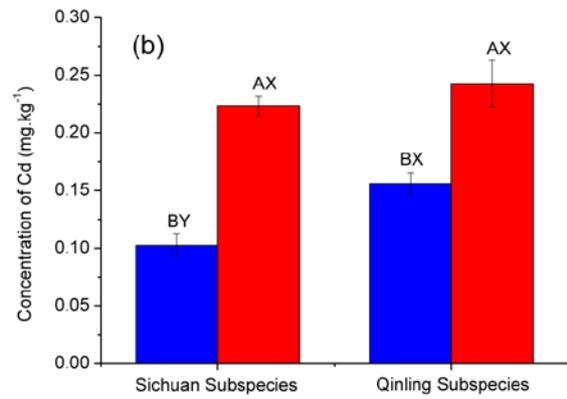
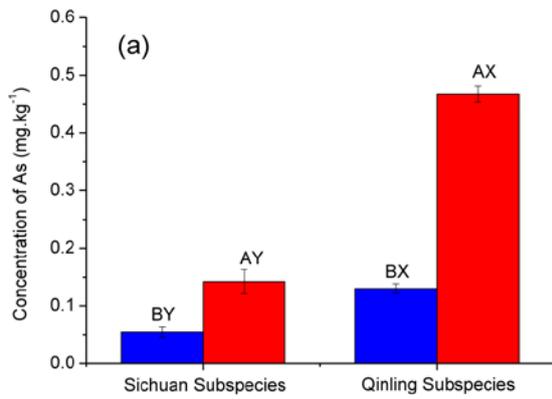
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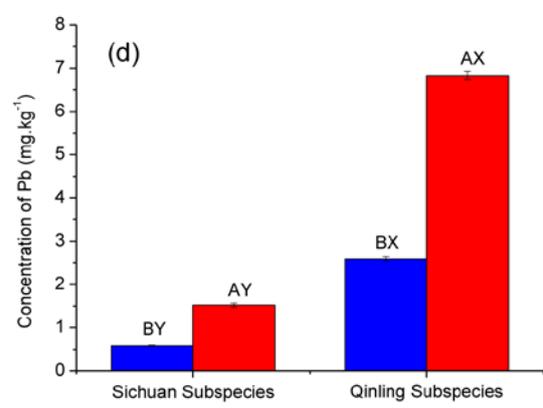
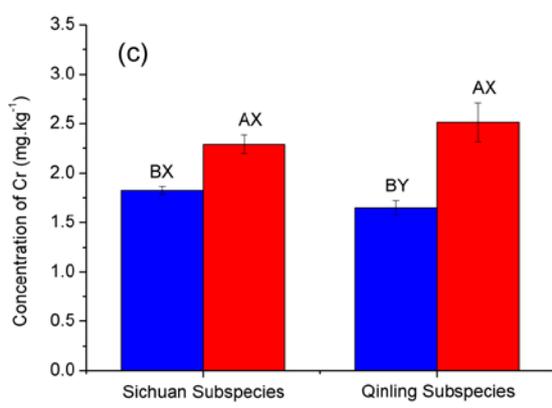
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Figure 4

1 **Captive pandas are at risk from environmental toxins**

2 Yi-ping Chen, Lorraine Maltby, Qiang Liu, Yi Song, Ying-juan Zheng, Aaron M. Ellison,

3 Qing-yi Ma, Xiao-min Wu

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5 **Supplemental Online Material**

6 **Additional Materials and Methods**

7 **Additional References**

8 **WebTables 1, 2**

9 **WebFigures 1 – 3**

10

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11

12 **Additional Materials and Methods**

13

14 ***Sample collection***

15 All faecal, plant, and feedstock samples were collected from the Wolong National

16 Nature Reserve in the Qionglai Mountains (“Wolong NNR”: 30°45’-31°25’N,

17 102°52’-103°25’E), the Foping National Nature Reserve in the Qinling Mountains

18 (“Foping NNR”: 33°33’-33°46’N, 107°40’-55’E), the China Conservation and

19 Research Center for Giant Panda (“CCRCGP”: 30°04’N, 102°59’E) and the Shaanxi

20 Wild Animal Research Center (“SWARC”: 34°06’N, 108°32’ E). The CCRCGP is the

21 largest captive breeding center for the Sichuan subspecies of the giant panda. It was  
22 relocated to its current location in Bifengxia Ya'an city from Wolong after the 2008  
23 Wenchuan earthquake. SWARC is located in Louguantai, Zhouzhi County, Xi'an city.  
24 It was established in 1987 and is the only center for conservation of the Qinling  
25 subspecies of the giant panda.

26 Faecal droppings of wild pandas were collected from 12 sites within the Wolong  
27 NNR and 16 sites within the Foping NNR. Sampling locations were 10 km apart and  
28 samples were pooled to give three samples/replicate from the Wolong NNR and four  
29 samples/replicate from the Foping NNR. Droppings of captive pandas of the Sichuan  
30 subspecies were collected at CCRCGP whereas droppings of captive pandas of the  
31 Qinling pandas were collected at SWARC. Droppings from either 12 individuals  
32 (CCRCGP) or 16 (SWARC) individuals were sampled and pooled into four replicates  
33 each comprising of three or four independent samples.

34

### 35 ***Source of the Environmental Toxins***

36 To investigate the source of the pollutants detected in panda droppings, the Qinling  
37 subspecies was studied in more detail. This subspecies was selected because there  
38 are about 350 individuals left (State Forestry Administration of the People's Republic  
39 of China.2015), so its conservation is much more urgent than that of the Sichuan  
40 subspecies. Further, as noted in the Results, the faecal droppings of the Qinling  
41 pandas contained significantly higher concentrations of As, Cd and Pb than droppings  
42 of the Sichuan subspecies.

43 Fresh leaves of the primary bamboo fed to these pandas (*Fargesia qinlingensis*,  
44 *Bashania fargesii*) and mixed feedstock used to make nutrient-supplements for the  
45 Qingling subspecies were collected from the Foping NNR, from plants cultivated at  
46 SWARC, and from feedstock at SWARC. Twelve samples of each food type were  
47 collected per location and pooled to produce four replicates each consisting of three  
48 samples.

49

### 50 **Heavy metal analysis**

51 All samples were dried to constant mass at 60°C before being homogenized using a  
52 ball mill. Dried samples (500 mg) were placed into Teflon bombs to which were added  
53 5 ml of HNO<sub>3</sub> for digestion with a microwave system (CEM, Mars 6, CEM, USA). After  
54 digestion, samples were diluted to 50 mL with deionized water. Concentrations of  
55 cadmium (Cd), chromium (Cr), and lead (Pb) were measured using a graphite  
56 furnace atomic absorption spectrometer (220-FS; Varian Company, USA.) with a  
57 hollow cathode lamp (Vigorous Instruments Co., Ltd., Beijing, China) (Yu *et al.* 2001).  
58 Concentrations of arsenic (As) were measured using an Atomic Fluorescence  
59 Spectrometer (AF-7500; Beijing Dongxi Instruments Co., Ltd., China) with a hollow  
60 cathode lamp (Vigorous Instruments Co., Ltd., Beijing, China) (Rahman *et al.* 2000).

61

### 62 **Analysis of PCDDs, PCDFs, and PCBs**

63 Samples (dropping, bamboo and feedstuff) were freeze-dried before being spiked  
64 with <sup>13</sup>C-labeled surrogate standards (Environmental Protection Agency [EPA]

65 method 1613B and 1668A) and underwent accelerated solvent extraction with  
66 dichlorinmethene: hexane (1:1). After determining the lipid content of each sample,  
67 the extract was adjusted to 50 ml with hexane; 15 g of acid silica (30% w/w) was  
68 added to remove lipids. The acid silica was stirred for 2 h and the extract was poured  
69 through 5 g of anhydrous sodium sulfite. All of the extracts were concentrated to 2 ml  
70 by rotary evaporation.

71 All solvents were purchased from Fisher (Fairlawn, NJ, USA). Silica gel was  
72 obtained from Merck (silica gel 60; Darmstadt, Germany). Basic alumina was  
73 obtained from Aldrich (Brockmann I, standard grade; Milwaukee, USA). Florisil was  
74 obtained from Riedel-de Haën (60–100 mesh ASTM; Seelze, Germany). Calibration  
75 standard solutions,  $^{13}\text{C}_{12}$ -labeled surrogate standards, and  $^{13}\text{C}_{12}$ -labeled injection  
76 standards were purchased from Wellington Laboratories (Guelph, Canada).

77 PCBs, PCDDs, and PCDFs were analyzed at the POP laboratory of the Research  
78 Center for Eco-environmental Sciences, Chinese Academy of Sciences; all  
79 concentrations were corrected for lipid weight. Sample extraction, cleanup, and  
80 chemical analysis followed established methods with some modifications (Liu *et*  
81 *al.* 2006; Li *et al.* 2008). Twenty-five PCB congeners, including 12 dioxin-like  
82 congeners, were quantified by an isotope dilution method using high-resolution gas  
83 chromatography coupled with high-resolution mass spectrometry (HRGC/HRMS).  
84 Total organic carbon (TOC) concentration was analyzed on a TOC Analyzer (O.I  
85 Analyzer; College Station, TX, USA). A 0.1-g sample was weighed and loaded into  
86 the combustion cup, which was packed with quartz wool. Prior to combustion, the

87 samples were wetted with 5% phosphoric acid and heated to 250°C for 1 min to  
88 purge inorganic carbon. The signal was detected by non-dispersed infrared (NDIR)  
89 detection when flashed at 900 °C for 6 min in the combustion house.

90 The quantification of 17 PCDD/PCDF homologues was done using  
91 HRGC/HRMS on an Agilent 6890 gas chromatograph coupled with an Autospec  
92 Ultima mass spectrometer (Waters Micromass, Manchester, UK) operating in the EI  
93 mode at 35 eV with the trap current was 600 IA. The GC was equipped with a CTC  
94 PAL autosampler. One or two µL samples were injected in splitless mode (splitless  
95 time, 2 min for PCDD/Fs) in a DB-5MS fused silica capillary column (60 m for  
96 PCDD/Fs and PCBs) with helium as carrier gas at a constant flow rate of 1.2 ml/min.  
97 The oven temperature programs were as follows: for PCDD/Fs, start 150°C held for 3  
98 min, 150-230°C at 20°C min<sup>-1</sup> held for 18 min, 230-235°C at 5°C min<sup>-1</sup> held for 10 min,  
99 235-320°C at 4°C min<sup>-1</sup> held for 3 min; for PCBs, start 120°C held for 1 min,  
100 120-150°C at 30 °C min<sup>-1</sup>, 150-300°C at 2.5°C min<sup>-1</sup> held for 1 min.

101

### 102 ***Quality control and quality assurance***

103 All data were subject to quality control and quality assurance. All glassware was  
104 washed two times with distilled water, and then with dichloromethane after use. After  
105 washing, glassware was dried for 6 hours at 400 °C in a muffle furnace.

106 All performance criteria required for the analysis of PCBs and PCDD/PCDFs  
107 followed US EPA methods (1668A and 1613B ).<sup>13</sup>C-labeled surrogated standards

108 (1668A-LCS and 1613-LCS) were spiked in the sample for qualification and  
109 quantification, and <sup>13</sup>C-labeled injection standards (EPA 68A-IS and 1613-IS) were  
110 added for recovery calculation. The recoveries of the surrogate standards ranged  
111 from 76.7±25.2% and 49.2±13.6% for PCB and PCDD/PCDFs, respectively, which  
112 met the requirements of US EPA methods 1668 A and 1613 B. Limit of detection (LOD)  
113 in the sample was defined as a signal to noise (S/N) ratio = 3. The LOD values were  
114 in the range of 0.01–0.82 pg g<sup>-1</sup> for PCBs and 0.04–8.40 pg g<sup>-1</sup> for PCDD/PCDFs.  
115 Laboratory blanks were analyzed with samples quality control at set intervals, and  
116 there was no detection of target compounds in the blanks.

117

118 **Additional References**

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127

128 WebTable 1. Concentrations of PCDD and PCDF congeners in faecal droppings from  
 129 wild and captive pandas of both Sichuan (SS) and Qinling (QS) subspecies. Values  
 130 are means  $\pm$  1 standard error of the mean for  $N = 4$  independent replicates each  
 131 comprising three or four pooled subsamples.

132

Congeners (pg.glw <sup>-1</sup> )	Wild SS	Captive SS	Wild QS	Captive QS
2,3,7,8-TetraCDF	7.802 $\pm$ 2.509	14.278 $\pm$ 4.946	3.402 $\pm$ 1.244	0.536 $\pm$ 0.473
1,2,3,7,8-PentaCDF;	10.098 $\pm$ 2.402	14.536 $\pm$ 5.413	2.074 $\pm$ 1.278	7.117 $\pm$ 7.117
2,3,4,7,8-PentaCDF;	11.585 $\pm$ 2.027	18.100 $\pm$ 3.694	3.214 $\pm$ 1.565	1.816 $\pm$ 1.816
1,2,3,4,7,8-HexaCDF	11.002 $\pm$ 1.477	15.934 $\pm$ 6.152	4.868 $\pm$ 2.764	17.300 $\pm$ 9.700
1,2,3,6,7,8-HexaCDF	8.900 $\pm$ 3.097	19.125 $\pm$ 5.736	6.352 $\pm$ 3.043	15.397 $\pm$ 8.576
2,3,4,6,7,8-HexaCDF;	6.568 $\pm$ 3.119	15.621 $\pm$ 5.895	4.507 $\pm$ 2.000	24.320 $\pm$ 8.676
1,2,3,7,8,9-HexaCDF	6.582 $\pm$ 4.347	2.186 $\pm$ 2.001	0.800 $\pm$ 0.289	0.579 $\pm$ 0.579
1,2,3,4,6,7,8-HeptaCDF	42.796 $\pm$ 14.568	65.054 $\pm$ 13.461	54.471 $\pm$ 17.196	113.048 $\pm$ 37.687
1,2,3,4,7,8,9-HeptaCDF	4.365 $\pm$ 2.732	0.566 $\pm$ 0.382	5.661 $\pm$ 3.843	1.283 $\pm$ 1.222
OctaCDF	34.857 $\pm$ 5.274	72.098 $\pm$ 31.329	67.689 $\pm$ 19.035	291.502 $\pm$ 165.188
2,3,7,8-TetraCDD	2.521 $\pm$ 1.456	0.631 $\pm$ 0.494	1.591 $\pm$ 1.334	☆
1,2,3,7,8-PentaCDD;	2.846 $\pm$ 1.582	2.952 $\pm$ 1.450	0.356 $\pm$ 0.235	1.289 $\pm$ 1.289
1,2,3,4,7,8-HexaCDD	7.514 $\pm$ 3.456	1.051 $\pm$ 0.698	0.680 $\pm$ 0.324	0.053 $\pm$ 0.053
1,2,3,6,7,8-HexaCDD	5.676 $\pm$ 2.729	4.926 $\pm$ 2.046	1.247 $\pm$ 0.428	2.216 $\pm$ 2.174
1,2,3,7,8,9-HexaCDD	6.654 $\pm$ 4.893	6.483 $\pm$ 3.772	☆	0.288 $\pm$ 0.288
1,2,3,4,6,7,8-HeptaCDD	37.442 $\pm$ 5.204	87.879 $\pm$ 24.617	24.736 $\pm$ 8.696	37.214 $\pm$ 22.064
OctaCDD	148.753 $\pm$ 8.656	469.836 $\pm$ 97.490	189.186 $\pm$ 29.664	115.741 $\pm$ 92.914

133 ☆ less than the limit of determination.

134

135

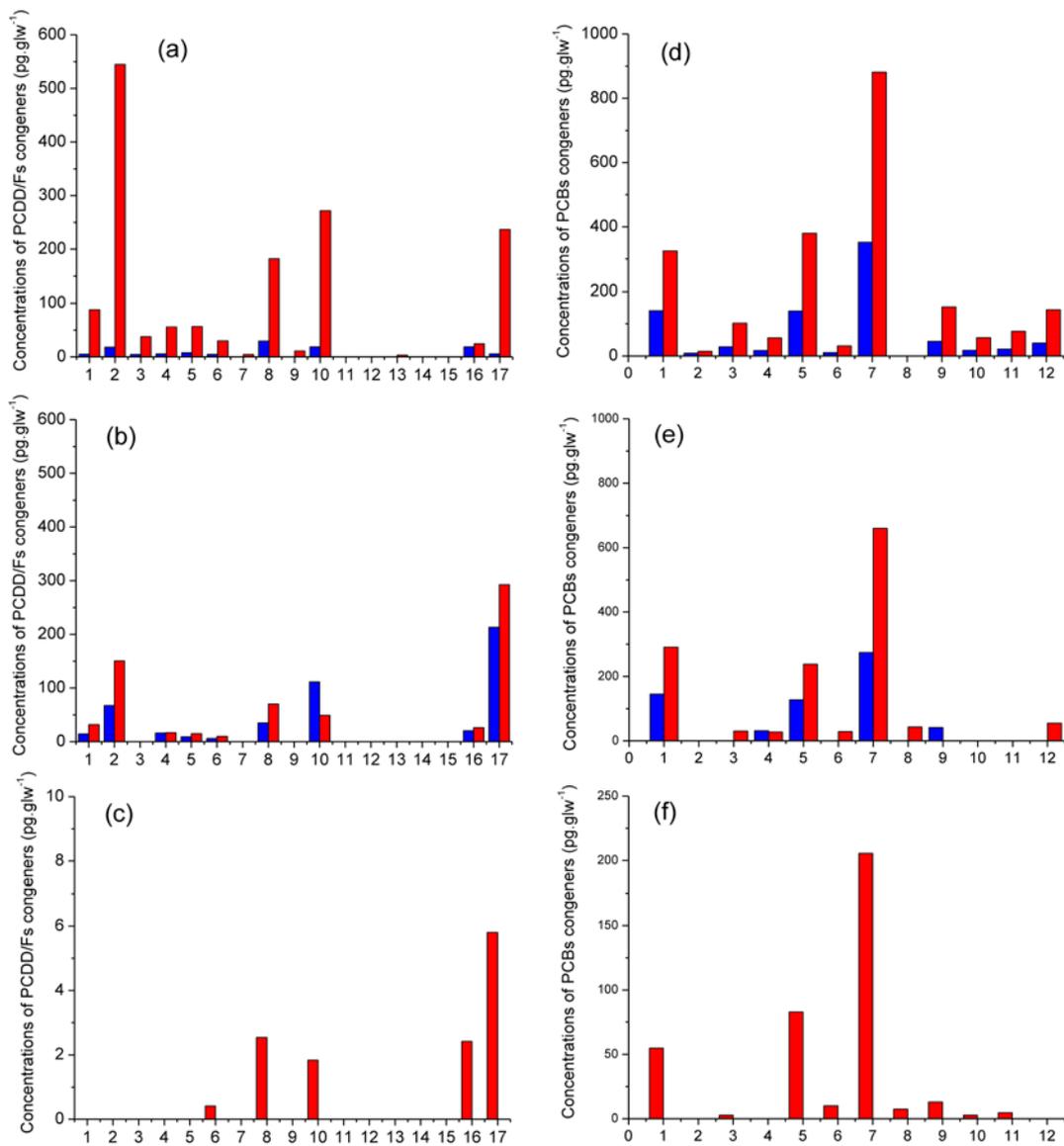
136 WebTable 2. Concentrations of PCB congeners in faecal droppings from wild and  
 137 captive pandas of both Sichuan (SS) and Qinling (QS) subspecies. Values are means  
 138  $\pm$  1 standard error of the mean for  $N = 4$  independent replicates each comprising  
 139 three or four pooled subsamples.

140

Congeners(pg.glw <sup>-1</sup> )	Wild of SS	Captive of SS	Wild of QS	Captive of QS
3,3',4,4'-TetraCB	72.745 $\pm$ 10.721	151.170 $\pm$ 22.013	82.564 $\pm$ 32.295	270.217 $\pm$ 44.520
3,4,4',5'-TetraCB	☆	5.768 $\pm$ 4.476	☆	45.409 $\pm$ 19.320
3,3',4,4',5-PentaCB	7.910 $\pm$ 3.856	155.379 $\pm$ 101.403	9.495 $\pm$ 5.967	165.775 $\pm$ 99.209
3,3',4,4',5,5'-HexaCB	1.721 $\pm$ 1.629	21.159 $\pm$ 14.862	☆	22.145 $\pm$ 22.145
2,3,3',4,4'-PentaCB	295.723 $\pm$ 31.734	716.133 $\pm$ 246.199	230.065 $\pm$ 57.408	556.994 $\pm$ 204.311
2,3,4,4',5-PentaCB	54.729 $\pm$ 6.431	54.830 $\pm$ 15.421	21.003 $\pm$ 8.061	70.455 $\pm$ 9.945
2,3',4,4',5-PentaCB	976.121 $\pm$ 57.739	788.122 $\pm$ 446.010	744.270 $\pm$ 135.106	731.055 $\pm$ 238.280
2',3,4,4',5-PentaCB	54.104 $\pm$ 5.305	79.313 $\pm$ 19.382	25.294 $\pm$ 3.305	51.228 $\pm$ 12.658
2,3,3',4,4',5-HexaCB	64.191 $\pm$ 4.816	68.530 $\pm$ 25.965	61.105 $\pm$ 4.694	94.781 $\pm$ 24.436
2,3,3',4,4',5'-HexaCB	11.179 $\pm$ 2.846	35.653 $\pm$ 13.606	12.693 $\pm$ 9.960	36.634 $\pm$ 6.900
2,3',4,4',5,5'-HexaCB	24.245 $\pm$ 7.676	18.822 $\pm$ 10.687	33.975 $\pm$ 5.332	38.804 $\pm$ 13.105
2,3,3',4,4',5,5'-HeptaCB	☆	4.982 $\pm$ 4.860	☆	1.763 $\pm$ 1.763

141 ☆ less than the limit of determination.

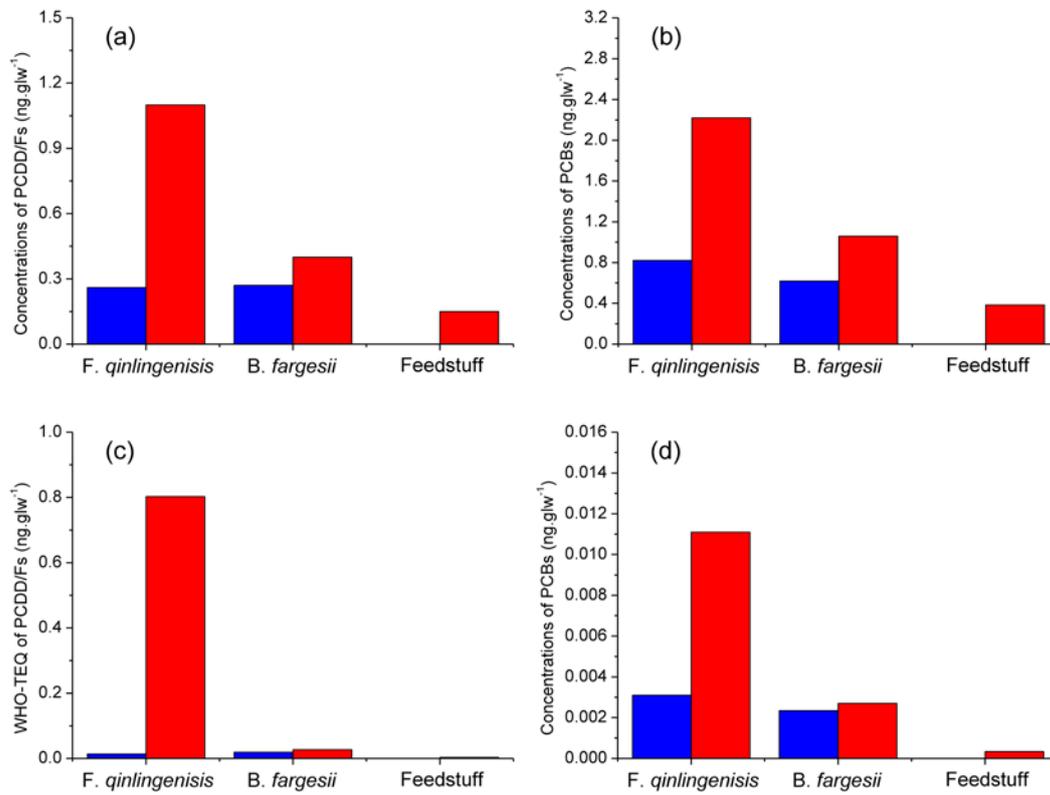
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143

144 WebFigure 1. The concentrations of congeners of (a, b, c) PCDDs and PCDFs; and  
 145 (d, e, f) PCBs in the bamboos *Fargesia qinlingensis* and *Bashania fargesii*, and  
 146 feedstuff of wild (blue) and captive (red) pandas. Numbers on the x-axis of panels (a),  
 147 (b) and (c) denote different congeners of PCDD/F. 1=2,3,7,8-TetraCDF;  
 148 2=1,2,3,7,8-PentaCDF; 3=2,3,4,7,8-PentaCDF; 4=1,2,3,4,7,8-HexaCDF;  
 149 5=1,2,3,6,7,8-HexaCDF; 6=2,3,4,6,7,8-HexaCDF; 7=1,2,3,7,8,9-HexaCDF;  
 150 8=1,2,3,4,6,7,8-HeptoCDF; 9=1,2,3,4,7,8,9-HeptaCDF; 10=OctaCDF;

151 11=2,3,7,8-TetraCDD; 12=1,2,3,7,8-PentaCDD; 13=1,2,3,4,7,8-HexaCDD;  
152 14=1,2,3,6,7,8-HexaCDD; 15=1,2,3,7,8,9-HexaCDD; 16=1,2,3,4,6,7,8-HeptaCDD;  
153 17=OctaCDD; Numbers on the x-axis of panels (d), (e) and (f) denote different  
154 congeners of PCBs. 1=3,3',4,4'-TetraCB; 2=3,4,4',5-TetraCB; 3=3,3',4,4',5-PentaCB;  
155 4=3,3',4,4',5,5'-HexaCB; 5=2,3,3',4,4'-PentaCB; 6=2,3,4,4',5-PentaCB;  
156 7=2,3',4,4',5-PentaCB; 8=2',3,4,4',5-PentaCB; 9=2,3,3',4,4',5-HexaCB;  
157 10=2,3,3',4,4',5'-HexaCB; 11=2,3',4,4',5,5'-HexaCB; 12= 2,3,3',4,4',5,5'-HeptaCB.  
158



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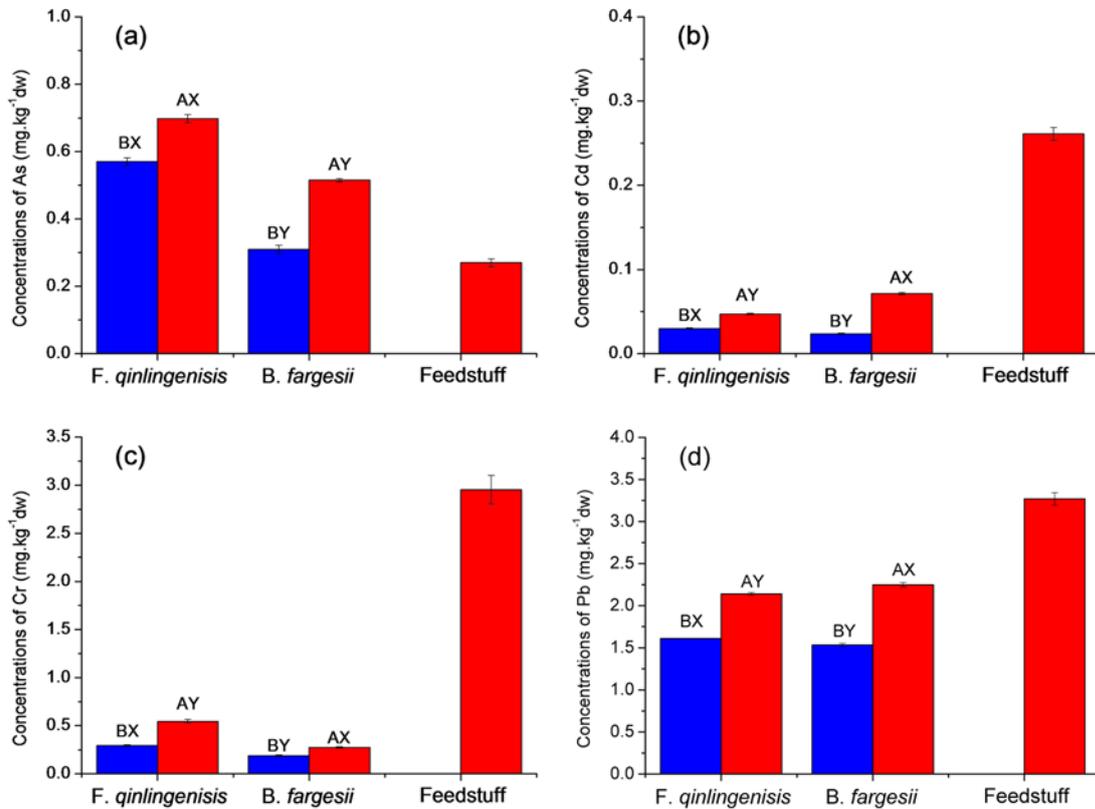
160 WebFigure 2. Concentrations of (a)  $\Sigma$ PCDDs and PCDFs, (b)  $\Sigma$ PCBs, (c) WHO-TEQ  
 161 of PCDDs and PCDFs, and (d) WHO-TEQ of PCBs in the bamboos *Fargesia*  
 162 *qinlingensis* and *Bashania fargesii*, and from panda feedstuff. Bars are the value from  
 163 a single replicate comprising five pooled samples.

164

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169

170 WebFigure 3. Concentrations of heavy metals in Qinling subspecies food of wild (blue)  
 171 and captive (red) from bamboo species (*Fargesia qinlingensis* and *Bashania fargesii*)  
 172 and feedstuff. (a) Arsenic (As); (b) Cadmium (Cd); (c) Chromium (Cr); (d) Lead (Pb).  
 173 Bars (means  $\pm$  1 SE of the mean from N = 4 independent replicates from three pooled  
 174 samples) with different letters between the two bamboo species (*Fargesia*  
 175 *qinlingensis* and *Bashania fargesii*) (A and B) or between the bamboos fed to wild and  
 176 captive (X or Y) are significantly different ( $P < 0.05$ , t-test).