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Observed decrease in atmospheric mercury explained by global decline in anthropogenic emissions

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Observations of elemental mercury (Hg^0) at sites in North America and Europe show large decreases ($\sim 1\text{--}2\% \text{ y}^{-1}$) from 1990 to present. Observations in background northern hemisphere air, including Mauna Loa Observatory (Hawaii) and CARIBIC (Civil Aircraft for the Regular Investigation of the atmosphere Based on an Instrument Container) aircraft flights, show weaker decreases ($< 1\% \text{ y}^{-1}$). These decreases are inconsistent with current global emission inventories indicating flat or increasing emissions over that period. However, the inventories have three major flaws: (i) they do not account for the decline in atmospheric release of Hg from commercial products; (ii) they are biased in their estimate of artisanal and small-scale gold mining emissions; and (iii) they do not properly account for the change in $\text{Hg}^0/\text{Hg}^{\text{II}}$ speciation of emissions from coal-fired utilities after implementation of emission controls targeted at SO_2 and NO_x . We construct an improved global emission inventory for the period 1990 to 2010 accounting for the above factors and find a 20% decrease in total Hg emissions and a 30% decrease in anthropogenic Hg^0 emissions, with much larger decreases in North America and Europe offsetting the effect of increasing emissions in Asia. Implementation of our inventory in a global 3D atmospheric Hg simulation [GEOS-Chem (Goddard Earth Observing System-Chemistry)] coupled to land and ocean reservoirs reproduces the observed large-scale trends in atmospheric Hg^0 concentrations and in Hg^{II} wet deposition. The large trends observed in North America and Europe reflect the phase-out of Hg from commercial products as well as the cobenefit from SO_2 and NO_x emission controls on coal-fired utilities.

mercury | trend | emission | atmosphere

Mercury (Hg) is released to the atmosphere by human activities including coal combustion, mining, and manufacturing and discard of commercial products (1, 2). Hg is transported globally as elemental Hg (Hg^0) in the atmosphere, eventually oxidizing to divalent Hg (Hg^{II}) that deposits to the surface, accumulates in ecosystems, and endangers humans and wildlife when converted to the neurotoxin methylmercury (3, 4). Surface air Hg concentrations in the northern hemisphere declined by 30–40% between 1990 and 2010 (5–7), and similar decreases have been observed in Hg wet deposition fluxes across North America and Western Europe (8, 9). By contrast, global inventories suggest flat or increasing Hg emissions over the last two decades (1, 10). Decreasing reemission of Hg from oceans and soils has been speculated (5, 6). Here we show that the declining atmospheric concentrations can be explained by the phase-out of Hg from commercial products and by shifts in the speciation of Hg emissions driven by air pollution control technologies.

Observed Atmospheric Hg Trends Since 1990

Table 1 compiles observed 1990-to-present trends in atmospheric Hg^0 concentrations and Hg^{II} wet deposition fluxes worldwide, including our own analyses. A general decline in the

concentration of Hg^0 is observed at surface sites, continuing to the most recent years. Decreases in atmospheric Hg^0 concentrations range from 1.2 to 2.1% y^{-1} at northern midlatitudes. Trends are weaker and less significant at high northern latitudes above 60° N (-0.9 to $+0.1\% \text{ y}^{-1}$). Preliminary data from an urban and a remote site in China suggest an increasing trend of about $+2\% \text{ y}^{-1}$ over the last decade (17). Wet deposition trends (available only for North America and Western Europe) are similar to trends in atmospheric concentrations.

Observed Hg^0 concentrations in the free troposphere above 2-km altitude show less significant declines. CARIBIC (Civil Aircraft for the Regular Investigation of the atmosphere Based on an Instrument Container) measurements in the northern hemisphere on commercial aircraft over the past decade (www.caribic-atmospheric.com) indicate a weak decline ($-0.6 \pm 0.6\% \text{ y}^{-1}$) that is not statistically significant ($P > 0.05$). Data from Mauna Loa, Hawaii (3.4 km above sea level), similarly indicate a statistically insignificant decline of $-0.9 \pm 0.6\% \text{ y}^{-1}$. We expect these free tropospheric trends to be representative of the tropospheric background, implying trends observed at surface sites are more influenced by regional sources and thus biased for global trend evaluation.

Significance

Anthropogenic mercury poses risks to humans and ecosystems when converted to methylmercury. A longstanding conundrum has been the apparent disconnect between increasing global emissions trends and measured declines in atmospheric mercury in North America and Europe. This work shows that locally deposited mercury close to coal-fired utilities has declined more rapidly than previously anticipated because of shifts in speciation from air pollution control technology targeted at SO_2 and NO_x . Reduced emissions from utilities over the past two decades and the phase-out of mercury in many commercial products has led to lower global anthropogenic emissions and associated deposition to ecosystems. This implies that prior policy assessments underestimated the regional benefits of declines in mercury emissions from coal-fired utilities.

Author contributions: Y.Z., D.J.J., D.P.K., and E.M.S. designed research; Y.Z. performed research; H.M.H., F.S., and V.L.S.L. contributed new reagents/analytic tools; Y.Z., H.M.H., L.C., H.M.A., and F.S. analyzed data; and Y.Z., D.J.J., and E.M.S. wrote the paper.

The authors declare no conflict of interest.

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Data deposition: The GEOS-Chem source code and run directory are available to download at geos-chem.org. The emission inventory and observational data are available at bgc.seas.harvard.edu.

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Table 1. Observed 1990 to present trends in atmospheric Hg⁰ concentrations and Hg^{II} wet deposition fluxes since the 1990s

| Period | Location (network) | Trend, % y ^{-1a} | Source |
|--|---------------------------------|----------------------------------|------------|
| Atmospheric Hg ⁰ concentrations | | | |
| 1995 to 2010 | Canada (CAMNet) | -1.6 ± 0.8 ^{b,c} | Ref. 7 |
| 1996 to 2004 | Cape Point, South Africa | -1.3 ± 0.3 ^{c,d} | Ref. 11 |
| 1990 to 1996 | Wank, Germany | -6.1 ± 1.1 ^{c,d} | Ref. 12 |
| 1996 to 2013 | Mace Head, Ireland | -1.3 ± 0.2 ^{c,e} | Ref. 13 |
| 1990 to 2009 | North Atlantic, cruises | -2.5 ± 0.5 ^{c,d} | Ref. 6 |
| | South Atlantic | Not significant | |
| 2000 to 2009 | Alert, Canada | -0.9 ± 0.5 ^{f,g} | Ref. 28 |
| | Zeppelin, Norway | Not significant | |
| 2008 to 2013 | United States (AMNet) | Not significant ^{c,e} | This study |
| 2005 to 2013 | Experimental Lakes Area, Canada | -2.2 ± 0.6 ^{c,e} | This study |
| 1990 to 2011 | Western Europe (EMEP) | -2.1 ± 0.5 ^{c,e} | This study |
| 1994 to 2012 | North of 60° N | Not significant ^{c,e} | This study |
| 2005 to 2014 | Free troposphere (CARIBIC) | Not significant ^{c,e,h} | This study |
| 2002 to 2013 | Mauna Loa Observatory, Hawaii | Not significant ^{c,d} | Ref. 14 |
| Hg ^{II} wet deposition | | | |
| 1996 to 2008 | North America (MDN) | Not significant ^{c,i} | Ref. 9 |
| | Western Europe (EMEP) | -1.5 ± 0.5 | |
| 1998 to 2005 | Northeast United States (MDN) | -1.7 ± 0.5 ^{c,j} | Ref. 15 |
| | Midwest United States | -3.5 ± 0.7 | |
| | Southeast United States | Not significant | |
| 1996 to 2005 | Northeast United States (MDN) | -2.1 ± 0.9 ^{g,k,l} | Ref. 8 |
| | Midwest United States | -1.8 ± 0.3 | |
| | Southeast United States | -1.3 ± 0.3 | |
| | West United States | -1.4 ± 0.4 | |
| 2002 to 2008 | Northeast United States (MDN) | Not significant ^{g,l} | Ref. 16 |
| | Midwest United States | Not significant | |
| 2004 to 2010 | Northeast United States (MDN) | -4.1 ± 0.5 ^{c,m} | Ref. 26 |
| | Midwest United States | -2.7 ± 0.7 | |
| | Southeast United States | Not significant | |
| | Western United States | Not significant | |
| 1996 to 2013 | North America, MDN | -1.6 ± 0.3 ^{c,j,n} | This study |
| 1990 to 2012 | Western Europe, EMEP | -2.2 ± 0.6 | |

^aAll trends reported are statistically significant ($P < 0.05$). Data from multiple sources are not synchronous; thus we present snapshots of changes over multiple years and regions where data are available.

^bTrends were calculated for sites with >5 y of measurements by using monthly median concentrations with the requirement that 75% of the month had valid data.

^cBased on linear regression.

^dAnnual median concentrations.

^eMonthly median concentrations.

^fDaily averaged concentrations.

^gSeasonal Kendall Test and Sen's slope method.

^hStratospheric data and biomass burning plumes filtered based on potential vorticity and CO and ozone concentrations (details provided in methods).

ⁱTrend were calculated for monthly mean wet deposition fluxes of sites with at least 75% data coverage.

^jAnnual wet deposition fluxes are used.

^kTrends are calculated for sites with >5 y of data and >75% valid data coverage.

^lWeekly wet deposition fluxes were used.

^mTrends are calculated for sites with >75% data coverage. Monthly precipitation volume-weighted mean concentrations were used.

ⁿSites with at least 10 and 7-y data coverage are selected for MDN and EMEP networks, respectively.

Revised Inventory of Hg Emissions

Standard Hg emission inventories used in atmospheric models (1, 10) indicate flat or increasing trends since 1990, seemingly inconsistent with the observed decreases. Horowitz et al. (2) developed an emission inventory that includes a very large missing source from the atmospheric release of Hg in commercial products. The authors showed that this commercial Hg source peaked in 1970 and has been declining rapidly since, driving an overall global decrease in Hg release to the atmosphere over the 1970-to-2000 period. The authors' inventory still shows an uptick in Hg emissions between 2000 and 2010 attributable to Asian coal-fired utilities and to artisanal and small-scale gold mining (ASGM) in developing countries (1, 19).

Recent work suggests weaker growth in Chinese emissions than previously estimated because of improved data on new coal-fired utilities with flue gas desulfurization (FGD) (20, 21). Also, the increasing trend in ASGM emissions appears to be a spurious effect of improved reporting (9).

Here, we revise Hg emissions and speciation from coal combustion to account for FGD and other emission controls in North America, Europe, Japan, and China. In North America and Europe, coal-fired utilities are the largest remaining atmospheric Hg source (1, 10). Combustion releases both Hg⁰, which has a relatively long atmospheric lifetime and is transported globally, and Hg^{II}, which is more likely to deposit regionally. US emissions from coal combustion declined by 75% over 2005 to 2015, mainly

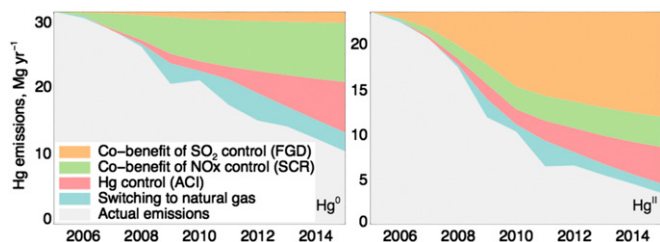


Fig. 1. Major factors driving declines in Hg emission from US coal-fired utilities between 2005 and 2015. Trends were inferred from data on the implementation of different types of emission control technologies.

because of cobenefits from controlling other atmospheric pollutants (Fig. 1). This decrease would be accompanied by a change in the $\text{Hg}^0/\text{Hg}^{\text{II}}$ speciation of emissions that is not recognized in current inventories. Use of FGD to control sulfur dioxide (SO_2) emissions washes out Hg^{II} (22). Use of selective catalytic reduction (SCR) to control nitrogen oxide (NO_x) emissions also oxidizes Hg^0 to Hg^{II} , and application of SCR and FGD in series controls total Hg emissions (23). Activated carbon injection (ACI) to specifically target Hg emissions has also begun to penetrate the energy sector (24). By considering the installation capacity and control efficiency of these devices, we find FGD caused the fraction of total Hg released as Hg^{II} to decline disproportionately to total emissions, from 43% to 24% over the last decade (Fig. 1).

Similar changes in Hg emissions from coal combustion can be inferred for other countries that have implemented air pollution controls and fuel switches over the last two decades. The fraction of coal-fired utilities with FGD increased from 20% to 46% between 1990 to 2002 in Western Europe and from 30% to 70% between 1990 and 2005 in Japan, resulting in a total cumulative decrease of 200 Mg Hg^{II} from developed countries including the US (25, 26). In China, the fraction of coal-fired utilities with FGD capacity increased from zero in 2000 to 86% in 2010, resulting in a drop in annual Hg^{II} emissions of 30% and 250 Mg cumulatively between 1990 and 2010 (21). The growth in energy demand in China has led to a rapid increase in coal combustion ($11\% \text{ y}^{-1}$), but total Hg emissions over this period increased less ($5.8\% \text{ y}^{-1}$) because of the implementation of FGD (21). This previously unaccounted shift in speciation implies greater declines in near-field Hg deposition than previously estimated.

Table 2 summarizes our updated global inventory of anthropogenic emissions for 1990 to 2010. The inventory includes revised estimates of emissions from coal combustion as described above, ASGM emissions from Muntean et al. (9), and emissions from commercial products (incineration, volatilization) based on Horowitz et al. (2) (additional details are available in Table S1). Our results indicate a 30% global decline of anthropogenic Hg^0 emissions from 1990 to 2010 ($-1.5\% \text{ y}^{-1}$). These declines are steepest from 1990 to 2000 but continue through 2010. This contrasts the flat or increasing trends in previous inventories (1, 9, 10). Two-thirds of the decline reflects the phase out of Hg in commercial products. Horowitz et al. (2) previously found that this was offset by rising emissions from coal combustion and ASGM, but our revision to the combustion inventory, as well as the Muntean et al. (9) ASGM inventory, removes the offset.

We find a global increase of 9% in Hg^{II} emissions between 1990 and 2010. This increase is attributable to growth in coal combustion in India and China, with FGD mitigating part of the increase in China. By contrast, Streets et al. estimated a 48% global increase in Hg^{II} emission between 1990 and 2008 (1). The authors did not account for growing implementation of FGD in China because of the lack of necessary information.

Large regional differences in Hg emission trends are apparent from Table 2. The decline in emissions from commercial products has been concentrated in developed countries (2). Total Hg emissions declined by a factor of 6.3 in Western Europe, 3.8 in North America, and 2.0 for other regions of Europe over 1990 to 2010, but emissions in Asia increased by a factor of 1.5.

Consistency with Observed Atmospheric Trends

Fig. 2 shows the 1990-to-2010 trends in atmospheric Hg^0 concentrations and Hg^{II} wet deposition fluxes simulated by the GEOS-Chem (Goddard Earth Observing System-Chemistry) global model using our revised anthropogenic emission inventory and the same meteorological year (to isolate the effect of emissions). Observed trends from Table 1 are also shown. Fig. 3 shows the simulated and observed regional trends averaged across the sites of Table 1.

The model successfully reproduces the observed trends. Declines are largest in North America and Western Europe (-1.5 ± 0.18 to $-2.2 \pm 0.15\% \text{ y}^{-1}$), reflecting the particularly large emission decreases in these regions. Shifts in speciation from coal-fired power plants also contribute significantly to the observed decline in wet deposition fluxes. Neglecting this change would result in an underestimate of the trend by a factor of 2 (e.g., ref. 26). The model decline in the free troposphere of the northern hemisphere ($-0.6 \pm 0.037\% \text{ y}^{-1}$) reflects a global decrease in total Hg^0 emission (including anthropogenic, natural, and reemission sources) of $-0.5\% \text{ y}^{-1}$. This decreasing trend is lower than that of anthropogenic Hg^0 emissions ($-1.5\% \text{ y}^{-1}$) because natural and legacy sources are approximately twice the magnitude of anthropogenic sources and are relatively unchanged. This modeled trend is in the range of observations from CARIBIC ($-0.6 \pm 0.57\% \text{ y}^{-1}$) and Mauna Loa Observatory ($-0.9 \pm 0.57\% \text{ y}^{-1}$). Model increases are limited to East Asia, consistent with preliminary observations and previous modeling studies (6, 11, 27).

The model decline at northern high latitudes (including one site in North American sector and four in Western Europe) is $-1.3 \pm 0.11\% \text{ y}^{-1}$, typical of the northern extratropical background (Fig. 24), but observations show a much weaker and statistically insignificant decline ($-0.2 \pm 0.5\% \text{ y}^{-1}$) (28). A similar discrepancy is observed for wet deposition fluxes over high-latitude regions of Western Europe (Fig. 2C). Fisher and coworkers (29, 30) and Zhang et al. (31) previously showed that trends in the Arctic are complicated by influences from riverine Hg discharges and sea ice cover. A GEOS-Chem simulation by Chen et al. (27), which includes long-term warming temperature and shrinking of Arctic sea ice, indicates that decreased oxidation of Hg^0 and deposition from the atmosphere as well as increased evasion of Hg^0 from the Arctic Ocean offsets the effect of the declining atmospheric background, resulting in no significant trend at high latitudes consistent with observations.

There is substantial uncertainty in current anthropogenic Hg emissions estimates (1, 9, 10, 19). Similar to Streets et al. (1), we calculated lower and upper bounds around the central estimate (Table 2 and Table S1) that correspond to an 80% confidence interval (CI) (the probability of emissions being outside this range is less than 20%). We estimate the resulting uncertainty in emissions is -33% to $+60\%$. However, the calculated emissions trend between 1990 to 2010 is much more consistent, ranging between -1.4% to $-0.53\% \text{ y}^{-1}$ and the propagated uncertainty in the simulated atmospheric trend across years is relatively small ($\pm 20\%$; Fig. 3).

Our results show general agreement between modeled and observed trends on the continental scale (Fig. 3), but the model does not reproduce all fine-scale variability in observations. For example, the model underestimates the atmospheric Hg^0 trend for one site in the western United States and cannot capture the observed increases in Hg^{II} wet deposition near the Four Corners Region in Colorado. These discrepancies are largely caused by

Table 2. Anthropogenic Hg emissions by world region

| Region and emissions | Year of emissions | | |
|----------------------------------|--------------------------|--------------------------|--------------------------|
| | 1990, Mg y ⁻¹ | 2000, Mg y ⁻¹ | 2010, Mg y ⁻¹ |
| Western Europe | | | |
| Hg ⁰ | 410 | 121 | 61 |
| ASGM | 0 | 0 | 0 |
| Products | 212 | 77 | 18 |
| Combustion | 198 | 44 | 44 |
| Hg ^{II} | 73 | 22 | 15 |
| Total | 483 (382–498) | 142 (106–228) | 77 (57–129) |
| North America | | | |
| Hg ⁰ | 399 | 174 | 109 |
| ASGM | 0 | 0 | 0 |
| Products | 208 | 88 | 66 |
| Combustion | 189 | 85 | 42 |
| Hg ^{II} | 70 | 42 | 15 |
| Total | 469 (361–635) | 216 (167–295) | 124 (93–175) |
| Asia and Oceania | | | |
| Hg ⁰ | 733 | 812 | 989 |
| ASGM | 81 | 181 | 243 |
| Products | 288 | 325 | 281 |
| Combustion | 363 | 306 | 465 |
| Hg ^{II} | 326 | 358 | 575 |
| Total | 1,060 (774–1,590) | 1,170 (806–1,860) | 1,560 (1,040–2,530) |
| Other Europe | | | |
| Hg ⁰ | 234 | 126 | 98 |
| ASGM | 7 | 11 | 13 |
| Products | 42 | 29 | 11 |
| Combustion | 185 | 86 | 75 |
| Hg ^{II} | 171 | 101 | 102 |
| Total | 405 (318–658) | 227 (173–385) | 200 (146–345) |
| Africa | | | |
| Hg ⁰ | 166 | 136 | 68 |
| ASGM | 78 | 83 | 28 |
| Products | 15 | 6 | 0 |
| Combustion | 73 | 46 | 40 |
| Hg ^{II} | 67 | 55 | 55 |
| Total | 233 (151–395) | 190 (114–335) | 123 (82–216) |
| Central and South America | | | |
| Hg ⁰ | 208 | 174 | 149 |
| ASGM | 110 | 90 | 93 |
| Products | 64 | 51 | 23 |
| Combustion | 35 | 34 | 33 |
| Hg ^{II} | 32 | 39 | 45 |
| Total | 240 (137–406) | 214 (126–366) | 194 (108–340) |
| Global | | | |
| Hg ⁰ | 2,150 | 1,540 | 1,480 |
| ASGM | 278 | 366 | 378 |
| Products | 829 | 576 | 398 |
| Combustion | 1040 | 600 | 699 |
| Hg ^{II} | 739 | 617 | 807 |
| Total | 2,890 (2,120–4,180) | 2,160 (1,490–3,470) | 2,280 (1,520–3,730) |

Taken from Muntean et al. (9). Products are emissions from use and disposal of commercial products from Horowitz et al. (2). Combustion indicates other and includes coal combustion, cement production, and metal smelting from Streets et al. (1). North America includes the United States and Canada. Mexico is included in Central and South America. Numbers in parenthesis are the 80% CIs.

local emission changes and meteorological effects (26), which are missed by the coarser-resolution simulations used here. More detailed, high-resolution emission inventories and models are required to fully resolve such fine-scale variability in observations.

The influence of changing climate and other environmental factors on the reemissions of Hg from soil and ocean remains unclear (32). The concentrations of major oxidants for atmospheric Hg⁰, including OH, O₃, and Br, have remained relatively

steady or slightly decreased since the mid-1990s and are thus not an important driver for the observed decline (5). Decreasing riverine discharges, which was previously speculated to drive the decline in North Atlantic Ocean Hg concentration and subsequent reemission flux (6), are also insufficient for forcing the global atmospheric trend (31, 33).

Our work has shown that revising anthropogenic emissions with the most up-to-date information can explain the observed

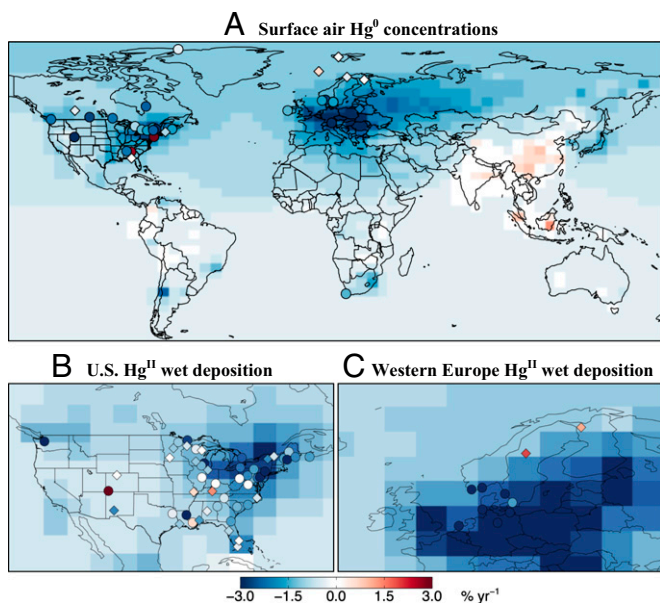


Fig. 2. Trends in atmospheric Hg^0 concentrations (A) and Hg^{II} wet deposition fluxes (B and C) from 1990 to present. Observations from the sites in Table 1 are shown as circles (if trends are statistically significant, $P < 0.05$) and diamonds (not significant). The background shows the trends computed in the GEOS-Chem model driven by our revised 1990 and 2010 anthropogenic emissions inventories from Table 2.

large-scale decline in atmospheric Hg over the past two decades. This finding reinforces the major benefits that have been derived from the phase-out of Hg in many products and emission controls on coal combustion.

Methods

Atmospheric Observations in 1990 to 2010. We include long-term observations (typically >5 y) for atmospheric Hg^0 concentrations and Hg^{II} wet deposition flux at sites worldwide during 1990 to 2014 (Table 1). For all of the measurements,

an ordinary linear regression on the annual means is used to calculate the trend. Because gaseous phase Hg^{II} accounts for less than 1–2% of TGM concentrations in surface air, we do not differentiate between Hg^0 and TGM in ground observations (34). We do not include atmospheric Hg^{II} concentrations because long-term records are few and data quality is uncertain (35).

Atmospheric Hg concentration data are available through the Canadian Atmospheric Mercury Network (CAMNet) (eight sites; <https://www.ec.gc.ca/natchem>), the US Atmospheric Mercury Network (AMNet) (nine sites; nadp.sws.uiuc.edu/amn), and European Monitoring and Evaluation Program (EMEP) (six sites; www.emep.int) networks. We include an analysis of the Hg^0 data measured at a remote and forested site, Experimental Lake Area, Canada (47.9° N, 93.7° W), during 2005 to 2013. Hg^{II} wet deposition fluxes are measured by the Mercury Deposition Network (MDN) (nadp.sws.uiuc.edu/mdn) and the EMEP network over North America and Western Europe, respectively. We select 52 MDN sites with at least 10-y data coverage during 1996 to 2013 and 11 EMEP sites with at least 7-y data since 1990. Atmospheric Hg concentrations have been measured from commercial aircraft by the CARIBIC project since December 2004 (www.caribic-atmospheric.com). We exclude stratospheric data using potential vorticity (European Centre for Medium-Range Weather Forecasts; www.ecmwf.int) and O_3 concentrations (36, 37). We exclude biomass burning plumes if the measured Hg concentration is greater than 2.5 ng m^{-3} . Also included here are the Hg^0 data measured at Mauna Loa Observatory, Hawaii (19.5° N, 155.6° W), during 2002 to 2013, which also samples the free troposphere at an elevation of 3,400 m.

Updated 1990 to 2010 Anthropogenic Emissions. We track the installation of air pollution control devices (FGD, SCR, and ACI) for individual US coal-fired utilities and calculate the associated Hg emission decline and speciation change based on the measured capture efficiencies with different coal types and control device configurations. Fuel type information is obtained from air markets program data by the US Environmental Protection Agency (EPA) (ampd.epa.gov/ampd). The installation time and type of air pollution control devices during 2005 to 2011 are from Mercury and Air Toxics Standards (MATS) Information Collection Request 2011 (www3.epa.gov/ttn/atw/utility/utilitypgp.html), with linear extrapolation to 2015 except for ACI, which is from the Institute of Clean Air Companies (www.icac.com). Hg capture efficiencies for different coal types and configuration of control devices are based on US EPA measurements (22). This speciation change over the US is extrapolated to all other developed countries in North America, Western Europe, and Oceania based on their similar trajectory of control technology (24). For China, we follow the change of speciation derived by Zhang et al. (20).

We develop an integrated emission inventory between 1990 and 2010 with decadal resolution, including improved estimates of Hg^{II} and Hg^0 speciation

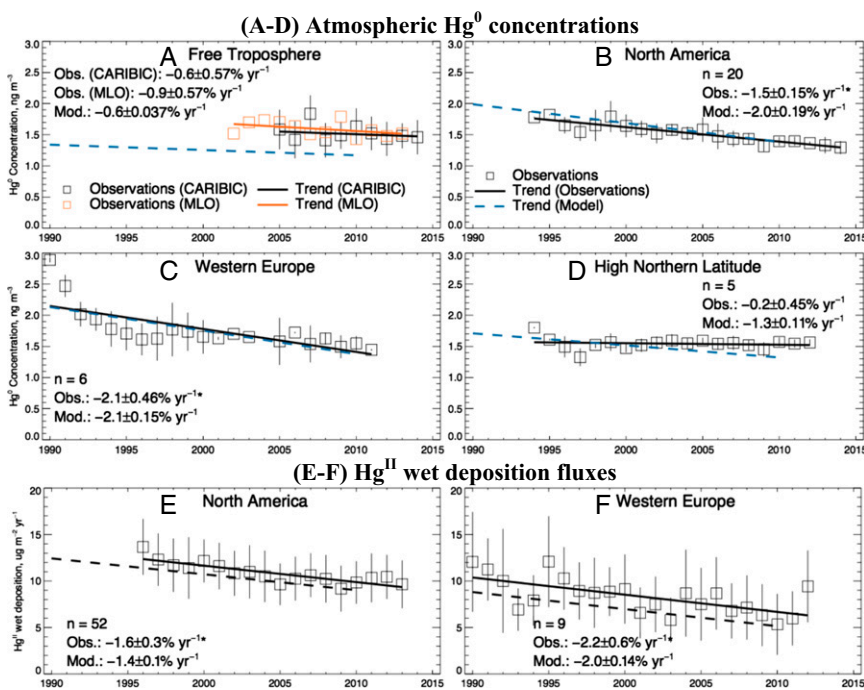


Fig. 3. Regional trends for 1990 to 2013 in atmospheric Hg^0 concentrations (A–D) and Hg^{II} wet deposition (E and F). Observations for individual years are shown as squares with linear regression as solid line. The dashed line is the trend from the GEOS-Chem simulation using our revised anthropogenic emissions inventory for 1990 and 2010 (Table 2). The data are averaged regionally across the sites in Table 1 for the free troposphere (A), North America (B and E), Western Europe (C and F), and high northern latitude regions (D) (vertical bars show the SDs). Regression coefficients (slope \pm SE) and number of sites (n) are given (insets). The SE of modeled trend is calculated based on the uncertainty range of the emission inventory (Table 2). North American atmospheric Hg^0 concentrations are from the CAMNet (<https://www.ec.gc.ca/natchem>) and AMNet (nadp.sws.uiuc.edu/amn) networks and the Experimental Lakes Area, Canada. North American Hg^{II} wet deposition is from the MDN (nadp.sws.uiuc.edu/mdn). Observations in Western Europe are from the EMEP network (www.emep.int). High-latitude sites include Alert, Canada and Zeppelin, Norway, and three sites above 60° N from the EMEP network.

from coal-fired utilities as described above. Default emissions are based on those of Streets et al. (1), who provide byproduct emissions from fossil fuel combustion, metal smelting, and waste incineration for 17 world regions. This inventory is updated using more recent country-specific estimates for China in 1995 to 2010 (20, 38), India in 2001 to 2020 (39), the US in 1990 to 2011 [National Emission Inventory (NEI) inventory: www3.epa.gov/ttnchie1/trends], and Western Europe in 1990 (40). We use global atmospheric releases from the use and disposal of commercial products from Horowitz et al. (2) and distribute to different regions based on Hg consumption (10, 18). Regional total emissions from these inventories are distributed on a $1^\circ \times 1^\circ$ grid following the spatial pattern of the Global Emissions Initiative (GEIA) inventory (41). For ASGM, we use the Emissions Database for Global Atmospheric Research gridded inventory (9).

Modeled Atmospheric Hg Concentrations and Deposition. The GEOS-Chem Hg model (version 9-01-02) is used to calculate the atmospheric Hg trends driven by Hg emission changes. The model includes a 3D global atmosphere coupled to 2D slab ocean and land models. A detailed description and evaluation of this model is available in Holmes et al. (42), Amos et al. (43), and Zhang et al. (44). The horizontal resolution is 4° latitude \times 5° longitude, with 47 vertical layers extending to the mesosphere. Atmospheric transport is driven by assimilated meteorological data from the GEOS-5 of the NASA Global Modeling and Assimilation Office.

The model traces two species: elemental mercury (Hg^0) and divalent mercury (Hg^{II}). Hg^{II} is partitioned thermodynamically between the gas and particle

phase on the basis of local temperatures and total aerosol concentration computed with a GEOS-Chem aerosol simulation (43). The model includes the oxidation of Hg^0 by atomic bromine and the photochemical reduction of Hg^{II} in cloud droplets. We do not include the fast reduction of Hg^{II} in coal-fired power plant plumes previously introduced by Zhang et al. (44) and Amos et al. (43) because more recent studies find that the reduction rate is slower than previously estimated (0–55% of emitted Hg^{II} is reduced with a mean value of 4.9%) (45, 46). Redox chemistry also takes place in the surface ocean and soil reservoirs, which receive atmospheric deposition and reemit to atmosphere through land–air and sea–air exchanges.

We conduct 3-y simulations with different anthropogenic emissions corresponding to years 1990 and 2010 (Table 2). The first 2 y are used for initialization, and the third year is used for analysis. The same meteorological year (2008) is used for all simulations to remove the influence of interannual meteorological variability. The soil and subsurface ocean concentrations are kept constant to isolate the impact of changing atmospheric emissions in the recent two decades.

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