A Satellite-based Biosphere Parameterization for Net Ecosystem CO2 Exchange: Vegetation Photosynthesis and Respiration Model (VPRM)

The Harvard community has made this article openly available. Please share how this access benefits you. Your story matters

Citation

Published Version
http://dx.doi.org/10.1029/2006GB002735

Citable link
http://nrs.harvard.edu/urn-3:HUL.InstRepos:3322246

Terms of Use
This article was downloaded from Harvard University’s DASH repository, and is made available under the terms and conditions applicable to Open Access Policy Articles, as set forth at http://nrs.harvard.edu/urn-3:HUL.InstRepos:dash.current.terms-of-use#OAP
A satellite-based biosphere parameterization for net ecosystem CO₂ exchange: Vegetation Photosynthesis and Respiration Model (VPRM)

Pathmathevan Mahadevan,¹ Steven C. Wofsy,¹ Daniel M. Matross,¹,² Xiangming Xiao,³ Allison L. Dunn,¹ John C. Lin,⁴ Christoph Gerbig,⁵ J. William Munger,¹ Victoria Y. Chow,¹ and Elaine W. Gottlieb¹

Received 4 April 2006; revised 7 August 2007; accepted 12 September 2007; published 12 April 2008.

[¹] We present the Vegetation Photosynthesis and Respiration Model (VPRM), a satellite-based assimilation scheme that estimates hourly values of Net Ecosystem Exchange (NEE) of CO₂ for 12 North American biomes using the Enhanced Vegetation Index (EVI) and Land Surface Water Index (LSWI), derived from reflectance data of the Moderate Resolution Imaging Spectroradiometer (MODIS), plus high-resolution data for sunlight and air temperature. The motivation is to provide reliable, fine-grained first-guess fields of surface CO₂ fluxes for application in inverse models at continental and smaller scales. An extremely simple mathematical structure, with minimal numbers of parameters, facilitates optimization using in situ data, with finesse provided by maximal infusion of observed NEE and environmental data from networks of eddy covariance towers across North America (AmeriFlux and Fluxnet Canada). Cross validation showed that the VPRM has strong prediction ability for hourly to monthly timescales for sites with similar vegetation. The VPRM also provides consistent partitioning of NEE into Gross Ecosystem Exchange (GEE, the light-dependent part of NEE) and ecosystem respiration (R, the light-independent part), half-saturation irradiance of ecosystem photosynthesis, and annual sum of NEE at all eddy flux sites for which it is optimized. The capability to provide reliable patterns of surface flux for fine-scale inversions is presently limited by the number of vegetation classes for which NEE can be constrained by the current network of eddy flux sites and by the accuracy of MODIS data and data for sunlight.


1. Introduction

[²] A primary goal of studying the terrestrial carbon cycle is to determine the magnitude of Net Ecosystem Exchange (NEE) of carbon dioxide between the terrestrial biosphere and the atmosphere and to understand the main drivers for hourly, seasonal, and interannual variations of NEE [Wofsy and Harriss, 2002]. Particular interest attaches to time-resolved measurements of fluxes on regional and continental scales, too small to be reliably resolved by global inverse models, but too large for direct measurement.

[³] Inverse (“top down”) analyses of CO₂ budgets on regional scales utilize measurements of atmospheric CO₂ concentrations on towers and by aircraft within the regions where sources and sinks are most active [Tans, 1980; Fung, 1993; Tans et al., 1993; Bakwin et al., 1998; Lin et al., 2004; Gerbig et al., 2005]. These data are influenced by small-scale, near-field fluxes as well as by continental and global sources and sinks, and the analysis therefore requires fine-scale spatial and temporal resolution for both transport fields and for distributions of surface fluxes [Gerbig et al. 2003a, 2003b; Baker et al., 2006]. Fluxes must be resolved on timescales including hourly, seasonal, and annual and on spatial scales as small as 1–10 km, a difficult challenge because NEE represents the difference between uptake (photosynthesis) and loss (respiration) processes that vary on a wide range of timescales [Goulden et al., 1996; Katul et al., 2001].
Since the inception of inverse modeling of CO\textsubscript{2}, it has been recognized that surface flux submodels must accurately represent relevant spatiotemporal variations of NEE [Fung et al., 1987; Ruimy et al., 1995; Sellers et al., 1996; Goetz and Prince, 1999; Xiao et al., 2002, 2004a, 2004b]. A priori surface flux models must have a low order of parameterization, so that the optimization process is well constrained [Denning et al., 1995; Lin et al., 2004], while retaining the required fine spatial and temporal resolution.

The present paper addresses the need to reliably represent surface fluxes at fine time/space scales with minimal parameters, into which we infuse the maximum information from observations. We use remotely sensed data to define vegetation properties with fine spatial resolution. Unfortunately, temporal resolution is poor, and direct information on NEE is lacking. We use measurements of NEE from eddy flux towers [Baldocchi et al., 2001] for direct flux data at high temporal resolution, capturing ecosystem functional responses to the environment at sites in North, Central, and South America, but, unfortunately, only small spatial scales (1 km\textsuperscript{2}).

The Vegetation Photosynthesis Respiration Model (VPRM) presented here assimilates remote sensing, meteorological, and tower flux data for a large number of sites in order to represent surface fluxes with the highest possible fidelity. Model structure is made very simple to facilitate subsequent inverse analysis. Formulation of the VPRM starts from the Vegetation Photosynthesis Model (VPM) of Xiao et al. [2004a, 2004b], which estimates Gross Ecosystem Exchange (GEE) using satellite-based vegetation indices and environmental data, adding respiration (R) to provide NEE and a nonlinear function to account for the response of GEE to light. The Enhanced Vegetation Index (EVI) [Huete et al., 1997, 2002] estimates of the Fraction of Photosynthetically Active Radiation (PAR) absorbed by photosynthetically active parts of the vegetation (FAPAR\textsubscript{PAR}) [Xiao et al., 2004a, 2004b] and the Land Surface Water Index (LSWI) help capture the effects of water stress and leaf phenology [Xiao et al., 2004a, 2004b], especially for vegetation that becomes dormant in summer (e.g., grasslands).

The VPRM shares many features of earlier models for surface CO\textsubscript{2} fluxes (e.g., NASA-CASA [Potter et al., 1993, 1999], SiB2 [Sellers et al., 1996], and TURC [Lafont et al., 2002]) developed for, and most appropriate to, global-scale inverse analysis, but it returns to the simpler functional representation introduced by Fung et al. [1987]. As summarized schematically in Figure 1, the VPRM systematically incorporates data from eddy flux towers, spanning dominant vegetation types over North America, plus MODIS data and high-resolution meteorological fields, to provide a much finer representation of surface fluxes than in previous simple models. VPRM NEE fields are thus optimally consistent with eddy flux data, and the model is readily exported to potential users and optimized using atmospheric data. Inversion of the VPRM is intended to enable it to capture seasonal and spatial variations of NEE not explicitly represented a priori.

2. Model Framework

Monteith [1972] showed that ecosystem production correlates with the Fraction of Absorbed Photosynthetically Active Radiation (FAPAR). FAPAR is often estimated as a linear or nonlinear function of the Normalized Difference Vegetation Index (NDVI) [Prince and Goward, 1995; Running et al., 2000], the normalized ratio between
satellite-derived reflectance in the red ($\rho_{\text{red}}$) and near-infrared ($\rho_{\text{nir}}$) bands [Tucker, 1979],

$$NDVI = \frac{\rho_{\text{nir}} - \rho_{\text{red}}}{\rho_{\text{nir}} + \rho_{\text{red}}},$$

(1)

using NDVI from the Advanced Very High Resolution Radiometer (AVHRR) to compute rates of terrestrial photosynthesis [e.g., Fung et al., 1987; Potter et al., 1993].

Recent studies [Xiao et al., 2004a, 2004b, 2005] showed that MODIS EVI [Huete et al., 1997, 2002] is more closely correlated with photosynthesis [Xiao et al., 2004a, 2004b] across a larger range of leaf area index and more closely follows phenology:

$$EVI = \frac{(\rho_{\text{nir}} - \rho_{\text{red}})}{\rho_{\text{nir}} + (C_1 \times \rho_{\text{red}} - C_2 \times \rho_{\text{blue}}) + L},$$

(2)

where $G = 2.5$, $C_1 = 6$, $C_2 = 7.5$, and $L = 1$. Inclusion of the blue band helps account for atmospheric contamination, and $L$ helps compensate for soil background reflectance. The VPRM also utilizes the LSWI [Xiao et al., 2004a, 2004b] to help capture effects of water stress and phenology on plant photosynthesis:

$$LSWI = \frac{\rho_{\text{nir}} - \rho_{\text{swir}}}{\rho_{\text{nir}} + \rho_{\text{swir}}},$$

(3)

where NIR refers to the 841–876 nm band, and SWIR refers to 1628–1652 nm.

2.1. Gross Ecosystem Exchange

[10] We divide NEE into a light-dependent term, Gross Ecosystem Exchange (GEE), and a light-independent part, ecosystem respiration ($R$), where $\text{NEE} = -\text{GEE} + R$, following the sign convention that uptake of CO$_2$ by plants is a negative flux (removal from the atmosphere). GEE is represented by

$$\text{GEE} = \varepsilon \times \frac{1}{(1 + PAR/\text{PAR}_0)} \times PAR \times \text{FAPAR}_{\text{PFT}},$$

(4)

where FAPAR$_{\text{PFT}}$ is the Fraction of Photosynthetically Active Radiation (PAR, $\mu$mol m$^{-2}$ s$^{-1}$) absorbed by the photosynthetically active portion of the vegetation (PAV), PAR$_0$ is the half-saturation value, and $\varepsilon$ is the light use efficiency ($\mu$mol CO$_2$/$\mu$mol PPFD) at low light levels. We decompose $\varepsilon$ into the product of the maximum quantum yield, $\varepsilon_0$, and factors ranging between 0 and 1 that reduce light use efficiency,

$$\varepsilon = \varepsilon_0 \times T_{\text{scale}} \times W_{\text{scale}} \times P_{\text{scale}}$$

(5)

On average, $\varepsilon_0$ has a value around 1/6 for well-watered, C$_3$ plants at optimal temperatures.

[11] The parameter $T_{\text{scale}}$ in equation (5) represents the temperature sensitivity of photosynthesis, calculated at each time step using the equation developed for the Terrestrial Ecosystem Model [Raich et al., 1991]:

$$T_{\text{scale}} = \frac{(T - T_{\text{min}})(T - T_{\text{max}})}{(T - T_{\text{min}})(T - T_{\text{max}}) - (T - T_{\text{opt}})^2},$$

(6)

where $T_{\text{min}}$, $T_{\text{max}}$, and $T_{\text{opt}}$ are minimum, maximum, and optimal temperatures ($^\circ$C) for photosynthesis, respectively [Aber and Federer, 1992; Raich et al., 1991]. If air temperature falls below $T_{\text{min}}$, $T_{\text{scale}}$ is set to zero [Xiao et al., 2004a, 2004b].

[12] Since temperature and PAR are correlated on a daily basis, inclusion of $T_{\text{scale}}$ in equation (5) modifies values of PAR$_0$ inferred from tower flux data. Moreover, were the parameters $T_{\text{min}}$, $T_{\text{max}}$, and $T_{\text{opt}}$ in equation (6) to be fit to eddy flux data along with the respiration equation (below) and PAR$_0$, parameter values would be unstable because of correlation between the parameters; therefore $T_{\text{min}}$, $T_{\text{max}}$, and $T_{\text{opt}}$ were fixed at literature values. The role of $T_{\text{scale}}$ in the VPRM is explored in a sensitivity analysis below.

[13] The function $P_{\text{scale}}$ accounts for effects of leaf age on canopy photosynthesis, using EVI and LSWI to identify the green-up (leaf expansion) and senescence phases [Xiao et al., 2002, 2004a; Boles et al., 2004]. For evergreen classes, $P_{\text{scale}}$ is assumed to be 1 for the whole year. For deciduous vegetation and grasslands we computed $P_{\text{scale}}$ as a linear function of LSWI from bud burst to leaf full expansion (“phase 1”) by

$$P_{\text{scale}} = \frac{1 + LSWI}{2}.$$  

(7)

[14] After leaf full expansion (phase two), $P_{\text{scale}}$ was set to 1, and equation (7) was adopted again during senescence (phase 3). The dates for the three phases of phenology (bud burst, full canopy, and senescence) were obtained using an EVI seasonal threshold similar to that of the MODIS phenology product MOD12Q2 [Friedl et al., 2003]. Thus for large-scale application of the VPRM across North America, MOD12Q2 dates can be used directly.

[15] The effect of water stress on GEE ($W_{\text{scale}}$) is a complex function of soil moisture and vapor pressure deficit (VPD) [e.g., Field et al., 1995; Running et al., 2000]. Soil moisture is currently used within the VPRM, since it cannot be derived directly from weather or remote sensing data [Pathmathevan et al., 2003]. We explored the use of the soil moisture product from the North American Land Data Assimilation System (NLDAS) [Mitchell et al., 2004], but we found the NLDAS product to be insufficiently accurate in simulating site soil moisture data. VPD could be derived from meteorological data, but it is a relatively minor influence compared to other factors (e.g., soil moisture) for most North American vegetation [cf. Powell et al., 2006; Makela et al., 2006; Cunningham, 2005]. A test run including VPD in the optimization of the VPRM at Harvard Forest produced results indistinguishable from a model where this factor was omitted, in part because of the very strong correlation of VPD with air temperature.

[16] Therefore following Xiao et al. [2004a], we express $W_{\text{scale}}$ as

$$W_{\text{scale}} = \frac{1 + LSWI}{1 + LSWI_{\text{max}}},$$

(8)

where LSWI$_{\text{max}}$ is the maximum LSWI within the plant-growing season for each site (or pixel). LSWI has been shown to capture drought-induced changes in plant canopies for ecosystems that senesce during dry periods, such as grasslands, but not for other vegetation. Hence effects of water stress are not explicitly included in the
VPRM and thus represent a principal source of unaccounted variance to be captured in an inverse analysis via adjustments to the VPRM parameters. This should be possible in inverse studies with regional resolution in space and month resolution in time, because soil moisture and related quantities tend to covary on regional scales and to change relatively slowly with time (~weeks).

[17] The complete expression for GEE in the VPRM is thus given by

$$GEE = \lambda \times T_{\text{scale}} \times P_{\text{scale}} \times W_{\text{scale}} \times EVI \times \frac{1}{(1 + PAR/\text{PAR}_0)} \times PAR$$

(9)

Here \( \lambda \) replaces \( \varepsilon_0 \), in order to aggregate into one parameter empirical adjustments to \( P_{\text{scale}} \), \( T_{\text{scale}} \), and \( W_{\text{scale}} \); \( \lambda \) and \( \text{PAR}_0 \) are the only adjustable parameters for description of the light-dependent part of NEE, with values derived below from tower flux data.

[18] \( PAR \) is measured at all flux tower sites, but not across the continent. At large scales the VPRM will be driven using shortwave (SW) radiation, available for almost all of North America from Geostationary Operational Environmental Satellite (GOES) data [e.g., Diak et al., 2004] and from assimilated meteorological products. SW is very closely correlated with PAR; SW \( \approx 0.505 \times PAR \) (units: SW, Watts/m\(^2\); PAR, \( \mu \text{mole} \text{ m}^{-2} \text{ s}^{-1} \)).

2.2. Ecosystem Respiration

[19] Plant and soil respiration rates generally increase as temperatures rise [Grace and Rayment, 2000; Piovesan and Adams, 2000], and we therefore represented \( R \) as

$$R = \alpha \times T + \beta.$$  

(10)

We set \( T = T_{\text{low}} \) in equation (10) when \( T \leq T_{\text{low}} \), to account for the persistence of soil respiration in winter, when air temperatures are very cold but soils remain warm. Values for \( \alpha, \beta \), and \( T_{\text{low}} \) were derived from tower flux data for each vegetation type (Tables 1 and 2). The intercept in equation (10) can be interpreted biogeochemically as the flux-weighted mean size of the respiring pools of organic matter in the ecosystem. In nature this number is determined by a complex set of antecedent conditions, such as site history and available stocks of necromass, in addition to climate factors. The VPRM adopts the zero-order approximation that these factors are uniform for each vegetation type, because site-specific information is not currently available from remote sensing or land cover databases. This approximation thus represents a second principal source of unaccounted variance, designed to be captured in an inverse analysis via adjustments to the VPRM parameters.

2.3. Net Ecosystem Exchange

[20] The full VPRM model equation is

$$NEE = -\lambda \times T_{\text{scale}} \times P_{\text{scale}} \times W_{\text{scale}} \times \frac{1}{(1 + PAR/\text{PAR}_0)} \times EVI \times PAR + \alpha \times T + \beta$$

(11)

adjusted in an inverse model application to provide an accurate representation for the distribution NEE in space and time across North America. The a priori estimates of these parameters are derived by optimizing the model using flux towers data at sites denoted as “calibration sites”. We assess transferability across the landscape by examining data from sites not used in deriving the prior estimates (“validation sites”).

3. Study Sites and Data

3.1. Vegetation and Tower Flux Data

[21] Tower measurements of NEE and water fluxes are made at numerous sites in North America and worldwide [Baldocchi et al., 2001]. We assembled a large subset of these data to calibrate and test VPRM surface fluxes, classified by vegetation type on the basis of the 1-km International Geosphere Biosphere Programme (IGBP) classification [Belward et al., 1999].

[22] Since tower flux data are not available for each of the 17 IGBP vegetation classes, we grouped North American ecosystems into nine major classes for which eddy flux data are available: evergreen forests, deciduous forest, mixed forest, shrubland (including open and closed shrubland), savannas (savannas and woody savannas), cropland, grassland (grassland, cropland/natural vegetation mosaic, and barren or sparsely vegetated), permanent wetlands, and others (especially the water bodies).

[23] Two of these nine large classes needed to be subdivided to account for major biophysical differences within them. The IGBP class “evergreen needleleaf forests” (~6.751% of land area) is broadly distributed, from boreal boggy black spruce to subtropical slash pine. We combined this class with “evergreen broadleaf forests”, which have negligible occurrence in North America (~0.5%), and then subdivided into four classes (boreal (e.g., black spruce), wet temperate/montane (e.g., Douglas fir, western white pine), dry temperate (e.g., ponderosa pine), and subtropical (e.g., slash pine, with strong summertime droughts)) by climate zone, using Holdridge Life Zone data [Monserud and Leemans, 1992]. Similarly, “cropland” was divided into soy and corn (to be expanded to include wheat when data become available). Fortunately, suitable eddy flux data are available for these subdivisions.

[24] We designated 11 tower sites to calibrate the four parameters for each vegetation class (except water, snow and ice, where fluxes are assumed zero), and identified 11 other sites for testing (“validation”) as listed in Table 1. More details and data for the 22 test sites can be obtained from network Web sites (http://public.ornl.gov/ameriflux/ and http://wwwfluxnet-canada.ca/) and from the original references in Table 1.

[25] The calibration sites for evergreen forests are the Northern Old Black Spruce (NOBS/BOREAS) site in Manitoba (boreal forest), Niwot Ridge in Colorado (subalpine coniferous forest), and Metolius Forest in Oregon.
Table 1. Carbon Flux and MODIS Data From These 22 AmeriFlux and Fluxnet-Canada Sites Used in This Study

<table>
<thead>
<tr>
<th>Abbreviation</th>
<th>Site Description</th>
<th>Data Year</th>
<th>LAT(N)</th>
<th>LON(W)</th>
<th>Country Description</th>
<th>Reference</th>
</tr>
</thead>
<tbody>
<tr>
<td>NOBS</td>
<td>NSA old black spruce forest</td>
<td>2000–2003</td>
<td>55.879</td>
<td>98.480</td>
<td>MB, Canada</td>
<td>Goulden et al. [1998], Dunn et al. [2007]</td>
</tr>
<tr>
<td>NIWOT</td>
<td>Niwot Ridge Forest (NWT1)</td>
<td>2000–2003</td>
<td>40.033</td>
<td>105.546</td>
<td>Colorado, United States</td>
<td>Monson et al. [2002], Yi et al. [2004]</td>
</tr>
<tr>
<td>METOLIUS</td>
<td>Metolius-intermediate (69 years) ponderosa pine forest</td>
<td>2002–2004</td>
<td>44.452</td>
<td>121.557</td>
<td>Oregon, United States</td>
<td>Coops et al. [2005]</td>
</tr>
<tr>
<td>DONALDSON</td>
<td>Donaldson midrotation (12 years) slash pine forest</td>
<td>2001–2002</td>
<td>29.755</td>
<td>82.163</td>
<td>Florida, United States</td>
<td>Clark et al. [1999, 2004]</td>
</tr>
<tr>
<td>HARVARD</td>
<td>Harvard Forest (main)</td>
<td>2000–2003</td>
<td>42.538</td>
<td>72.171</td>
<td>Massachusetts, United States</td>
<td>Wofsy et al. [1993], Davidson et al. [2002a]</td>
</tr>
<tr>
<td>HOWLAND</td>
<td>Howland Forest (main)</td>
<td>2000–2003</td>
<td>45.204</td>
<td>68.740</td>
<td>Maine, United States</td>
<td>Wofsy et al. [1993], Davidson et al. [2002a]</td>
</tr>
<tr>
<td>TONZI</td>
<td>Tonzi Range savannas</td>
<td>2002–2004</td>
<td>38.432</td>
<td>120.966</td>
<td>California, United States</td>
<td>Xu and Baldocchi [2004], Baldocchi et al. [2004]</td>
</tr>
<tr>
<td>MEAD-S2</td>
<td>irrigated maize-soybean rotation site (site 2)</td>
<td>soy: 2002 corn: 2003</td>
<td>41.099</td>
<td>96.281</td>
<td>Nebraska, United States</td>
<td>Verna et al. [2005]</td>
</tr>
<tr>
<td>Vaira</td>
<td>Vaira Range grassland</td>
<td>2001–2003</td>
<td>38.407</td>
<td>120.951</td>
<td>California, United States</td>
<td>Xu and Baldocchi [2004], Baldocchi et al. [2004]</td>
</tr>
<tr>
<td>PEATLAND</td>
<td>eastern peatland, permanent wetland</td>
<td>2002</td>
<td>45.409</td>
<td>75.520</td>
<td>Ontario, Canada</td>
<td>Lafleur et al. [2001, 2003]</td>
</tr>
<tr>
<td>ANLGRASS</td>
<td>Walnut River watershed, grassland</td>
<td>2002–2003</td>
<td>37.521</td>
<td>96.855</td>
<td>Kansas, United States</td>
<td>Song and Wescy [2003], Song et al. [2006], Coulter et al. [2006]</td>
</tr>
<tr>
<td>WLEF</td>
<td>Park Falls/WLEF</td>
<td>2000–2001</td>
<td>45.946</td>
<td>90.272</td>
<td>Wisconsin, United States</td>
<td>Davis et al. [2003], Wang et al. [2007b]</td>
</tr>
<tr>
<td>WCREEK</td>
<td>Willow Creek</td>
<td>2000–2004</td>
<td>45.806</td>
<td>90.080</td>
<td>Wisconsin, United States</td>
<td>Desai et al. [2005], Cook et al. [2004]</td>
</tr>
<tr>
<td>LCREEK</td>
<td>Lost Creek</td>
<td>2001–2004</td>
<td>46.083</td>
<td>89.979</td>
<td>Wisconsin, United States</td>
<td>Desai et al. [2008], Wang et al. [2006]</td>
</tr>
<tr>
<td>SOBS</td>
<td>SSA old black spruce forest</td>
<td>2000–2004</td>
<td>53.987</td>
<td>105.117</td>
<td>Saskatchewan, Canada</td>
<td>Turner et al. [2003], Griffis et al. [2003]</td>
</tr>
<tr>
<td>B1850</td>
<td>NSA 1850 burn site</td>
<td>2001–2004</td>
<td>55.880</td>
<td>98.480</td>
<td>Manitoba, Canada</td>
<td>Goulden et al. [2006]</td>
</tr>
<tr>
<td>EOB3</td>
<td>Quebec mature boreal forest</td>
<td>2004</td>
<td>49.693</td>
<td>74.342</td>
<td>PQ, Canada</td>
<td>Bergeron et al. [2007]</td>
</tr>
<tr>
<td>INDIANA</td>
<td>Morgan Monroe State Forest</td>
<td>2000–2003</td>
<td>39.323</td>
<td>86.413</td>
<td>Indiana, United States</td>
<td>Schmidt et al. [2000], Su et al. [2004]</td>
</tr>
<tr>
<td>DUKE_HW</td>
<td>Duke Forest: hardwoods</td>
<td>2001–2004</td>
<td>35.974</td>
<td>79.100</td>
<td>North Carolina, United States</td>
<td>Stoy et al. [2005]</td>
</tr>
</tbody>
</table>
Table 2. Parameters PAR$_{0}$, $\lambda$, $\alpha$, $\beta$, and Their Variances and Light Use Efficiency at Calibration Sites$^a$

<table>
<thead>
<tr>
<th>Site</th>
<th>$T_{\min}$</th>
<th>$T_{\text{avg}}$</th>
<th>$T_{\text{max}}$</th>
<th>$T_{\text{low}}$</th>
<th>PAR$_{0}$</th>
<th>$\lambda$</th>
<th>$\alpha$</th>
<th>$\beta$</th>
<th>$\sigma$-PAR$_{0}$</th>
<th>$\sigma$-$\lambda$</th>
<th>$\sigma$-$\alpha$</th>
<th>$\sigma$-$\beta$</th>
</tr>
</thead>
<tbody>
<tr>
<td>HARVARD</td>
<td>0</td>
<td>20</td>
<td>40</td>
<td>5</td>
<td>570</td>
<td>0.127</td>
<td>0.271</td>
<td>0.25</td>
<td>14</td>
<td>0.002</td>
<td>0.006</td>
<td>0.060</td>
</tr>
<tr>
<td>HOWLAND</td>
<td>0</td>
<td>20</td>
<td>40</td>
<td>2</td>
<td>629</td>
<td>0.123</td>
<td>0.244</td>
<td>-0.24</td>
<td>17</td>
<td>0.002</td>
<td>0.004</td>
<td>0.036</td>
</tr>
<tr>
<td>NIWOT</td>
<td>0</td>
<td>20</td>
<td>40</td>
<td>1</td>
<td>262</td>
<td>0.234</td>
<td>0.244</td>
<td>0.14</td>
<td>5</td>
<td>0.004</td>
<td>0.002</td>
<td>0.015</td>
</tr>
<tr>
<td>METOLIUS</td>
<td>0</td>
<td>20</td>
<td>40</td>
<td>2</td>
<td>446</td>
<td>0.128</td>
<td>0.250</td>
<td>0.17</td>
<td>13</td>
<td>0.003</td>
<td>0.003</td>
<td>0.018</td>
</tr>
<tr>
<td>SOY_MEADS2</td>
<td>5</td>
<td>22</td>
<td>40</td>
<td>1</td>
<td>2051</td>
<td>0.064</td>
<td>0.209</td>
<td>0.20</td>
<td>137</td>
<td>0.002</td>
<td>0.005</td>
<td>0.058</td>
</tr>
<tr>
<td>CORN_MEAD</td>
<td>5</td>
<td>22</td>
<td>40</td>
<td>2</td>
<td>11250</td>
<td>0.075</td>
<td>0.173</td>
<td>0.82</td>
<td>1746</td>
<td>0.002</td>
<td>0.006</td>
<td>0.081</td>
</tr>
<tr>
<td>TONZI</td>
<td>2</td>
<td>20</td>
<td>40</td>
<td>-</td>
<td>3241</td>
<td>0.057</td>
<td>0.012</td>
<td>0.58</td>
<td>293</td>
<td>0.002</td>
<td>0.002</td>
<td>0.036</td>
</tr>
<tr>
<td>VAIRA</td>
<td>2</td>
<td>18</td>
<td>40</td>
<td>-</td>
<td>542</td>
<td>0.213</td>
<td>0.028</td>
<td>0.72</td>
<td>23</td>
<td>0.006</td>
<td>0.002</td>
<td>0.035</td>
</tr>
<tr>
<td>DONALDSON</td>
<td>2</td>
<td>20</td>
<td>40</td>
<td>1</td>
<td>790</td>
<td>0.114</td>
<td>0.153</td>
<td>1.56</td>
<td>18</td>
<td>0.002</td>
<td>0.004</td>
<td>0.076</td>
</tr>
<tr>
<td>LUCKY_HILLS</td>
<td>2</td>
<td>20</td>
<td>40</td>
<td>-</td>
<td>321</td>
<td>0.122</td>
<td>0.028</td>
<td>0.48</td>
<td>14</td>
<td>0.004</td>
<td>0.001</td>
<td>0.019</td>
</tr>
<tr>
<td>PEATLAND</td>
<td>0</td>
<td>20</td>
<td>40</td>
<td>3</td>
<td>558</td>
<td>0.051</td>
<td>0.081</td>
<td>0.24</td>
<td>23</td>
<td>0.002</td>
<td>0.002</td>
<td>0.019</td>
</tr>
</tbody>
</table>

$^a$Units are as follows: PAR$_{0}$: $\mu$mole m$^{-2}$ s$^{-1}$; $\lambda$: $\mu$mole CO$_2$ m$^{-2}$ s$^{-1}$/mole PAR m$^{-2}$ s$^{-1}$; $\alpha$: $\mu$mole CO$_2$ m$^{-2}$ s$^{-1}$; $\beta$: $\mu$mole CO$_2$ m$^{-2}$ s$^{-1}$. Light use efficiency: $\lambda$.

(ponderosa pine) for wet and dry temperate evergreen, respectively, and Donaldson (Florida slash pine) for subtropical dry evergreen forest. We would like additional evergreen classes for cool nonmontane pines (e.g., white pine) and for hemlock, but flux data are unavailable.

[26] Harvard Forest was the calibration site for deciduous broadleaf forests ($\sim$1.976%) and closed broadleaf deciduous (e.g., larch) which do not occur extensively in North America. IGMP mixed forest ($\sim$7.29%) was calibrated using Howland (Maine). IGMP closed ($\sim$0.54%) and open shrubland ($\sim$8.6%) were combined into “shrubland,” and calibrated using Lucky Hills. IGMP woody savannas ($\sim$1.3%) and savannas ($\sim$0.14%) were combined (“savannas”) and calibrated using Tonzi Ranch. The IGMP class of croplands ($\sim$3.77%) was adopted as is and calibrated using data from Mead-S2 (Nebraska) for both irrigated maize and soybeans, planted in rotation.

[27] IGMP “grasslands” ($\sim$3.3%), “crop/natural vegetation mosaic” ($\sim$3.9%), and “barren or sparsely vegetated lands” ($\sim$1.8%) were combined into VPRM “grasslands” and calibrated at the Vaira range site. This class may be affected by significant representation errors when the grassland calibration is applied to crop/natural mosaics, which in the northern tier are often dairy farms interspersed with woodlands. However, there are no data to allow subdivision of these categories. The IGMP “permanent wetlands” ($\sim$0.7%) was calibrated at the eastern peatland site in Canada. IGMP classes for water bodies ($\sim$59%), urban and built-up ($\sim$0.18%), and snow and ice ($\sim$0.27%) were combined into our last class, for which vegetation-derived fluxes are assigned as zeros.

[28] Tower data sets provide several versions of NEE: with and without filtering by turbulent intensity ($u^*$) and with or without gap filling. Some sites also provide $G_{EE}$ and $T_{\text{net}}$, separated using various approaches. To avoid possible biases and inconsistencies from filtering or separating $G_{EE}$ and $T_{\text{net}}$, VPRM parameters were optimized against unfilled tower NEE, with a $u^*$ filter applied to eliminate unrepresentative observations.

[29] The current VPRM is intended to cover vegetation from 11°N to 65°N and 50°W to 145°W, including the continental United States, Mexico, and most of Canada. For large-scale applications the 1-km IGMP vegetation data were classified into these types and regridded at 10 × 10 km, or 1/4° × 1/6°, retaining information on the fractional coverage for each vegetation type. These data are provided to the public with the VPRM distribution.

3.2. Satellite Data

[30] We analyzed a multiyear satellite images from the MODIS sensor aboard the Terra satellite (2000–2003/2004), crossing the equator at 1030 local time. MODIS views the entire surface of the Earth every 1–2 d measuring 36 spectral bands at 250 or 500 m resolution between 0.405 and 14.385 $\mu$m.

[31] We acquired 8-d mean MODIS surface reflectances (MOD09A1) for our calibration and validation sites from the Oak Ridge Distributed Active Archive Center (http://modis.gsfc.nasa.gov/modis/index.cfm), which provides time series data for most flux towers in ASCII format. We had to process MODIS subsets directly (Hierarchical Data Format (HDF); http://landval.gsfc.nasa.gov) for sites where the MODIS ASCII subsets were unavailable (e.g., Lucky Hill).

[32] The MOD09A1 products give data for nine MODIS pixels covering 1.5 km × 1.5 km, centered on each flux tower. We averaged the 8-d mean surface reflectance data for red (620–670 nm), NIR (841–876 nm), blue (459–479 nm), and SWIR (1628–1652 nm) to calculate EVI and LSWI, then applied a low-order smoothing algorithm (“lowess”, locally weighted least squares) [Cleveland, 1981] to the time series for each to reduce noise associated with imperfect atmospheric corrections in MOD09A1 data.

4. Results

[33] We optimized model parameters ($\lambda$, PAR$_{0}$, $\alpha$, and $\beta$; see Table 2) via nonlinear least squares (Newton-Raphson, tangent linear approximation) and estimated confidence intervals assuming Gaussian errors for both model and tower data. For each calibration site we generated hourly data from the smoothed time series of vegetation indices (EVI and LSWI) and obtained measurements of air temperature and PAR from the tower sites.

[34] Examples of observed and modeled NEE are shown in Figure 2a. The VPRM provides consistent partitioning of
tower NEE data into light-dependent and light-independent parts for all calibration sites, and it thus provides an independent tool for filling missing data (see Figure 2a). (Note that $T_{\text{scale}}$ is assumed to define the temperature dependence of photosynthesis.) VPRM has the advantage of incorporating satellite data into the process, and it can be applied to any tower site. It yields consistent, independent estimates of annual net exchange for all sites where the optimization procedure is run.

When driven by high-resolution data sets, the VPRM equations are able to reproduce 1 to 4 years of data with remarkable fidelity, including both diurnal cycles (Figure 2b) and aggregation to monthly timescales (Figure 3), despite their ultrasimple form. Inputs of accurate solar irradiance and temperatures allow the VPRM to closely track hourly variations; inputs from remote sensing data enable the VPRM to also track the seasonal course of NEE. The model even captures a significant amount of interannual variability, driven by variations in $T$, PAR, and EVI (Figure 2a, right).

Values of $\lambda$ for forests and crops range from 0.17 to 0.27 (Table 2), consistent with the expectation that optimum light use efficiency at low light should be $\sim 1.6$ ($\lambda = 0.17$) for a dense vegetation canopy. Values are lower for semiarid grasslands and shrublands, again as expected. Values of $r^2$
Figure 2b. A comparison between the observed (solid squares) and VPRM (open circles) mean diurnal variation of NEE (μmole m$^{-2}$ s$^{-1}$) during the peak growing season at calibration sites.
Figure 3
range from 0.6 to 0.9 for calibration sites; correlations are almost as good at many validation sites. Note the high value of \( \text{PAR}_0 \) (Table 2) for corn, apparently capturing the high LUE [Gower et al., 1999]. \( \text{PAR}_0 \) values in Table 2 are higher at cropland xeric sites than would be found in conventional analysis of a light curve, where NEE is fit to a hyperbolic function of PAR. Midday summer temperatures often exceed \( T_{\text{opt}} \), and hence the VPRM infers high \( R \) and low GEE, attributing the decrease in photosynthetic efficiency to excessive heat rather than to light saturation.

[37] Figure 3, top, shows the relationship between the seasonal dynamics of NEE and VPRM photosynthesis factors. As expected, croplands and grasslands respond strongly to phenology \( (P_{\text{scale}}) \) and the amount of photosynthetically active vegetation \( (\text{EVI}) \). Likewise, variations in \( P_{\text{scale}} \) and \( \text{EVI} \), as well as light \( (\text{PAR}) \), strongly modulate the uptake of \( \text{CO}_2 \) at deciduous and mixed sites (Harvard and Howland), whereas the temperature dependence of photosynthesis \( (T_{\text{scale}}) \) is the primary factor limiting uptake of \( \text{CO}_2 \) by well-watered evergreen forests (NOBS, Metolius/Oregon, and Niwot). Intrasessional trends in the VPRM sometimes were able to capture water stress and changes in EVI.

[38] There are a few surprises. Harvard and Howland forests both include significant evergreen conifers, as typical for “deciduous” and “mixed” forests, and \( T_{\text{scale}} \) is thus also critically important in limiting uptake at these sites in winter. Donaldson is warm and evergreen, but in winter it is not actually very green at all, and the very low values of \( \text{EVI} \) limit uptake. The notably poor fit at Donaldson in summer may be particular to the 2001–2002 interval used for calibration; this was the end of a severe, extended drought, and remotely sensed indices might not have captured the associated aftereffects.

[39] Other discrepancies appear to be associated with the inability of remotely sensed data to detect water stress and/or conductance limitations during summer at sites with strong coniferous representation (Donaldson, Metolius, and Howland). Thus the VPRM overpredicts uptake at these sites in middle and late summer, when photosynthesis rates decline steeply but \( \text{EVI} \) and \( \text{LSWI} \) change only modestly. At some sites the model does a surprisingly good job in capturing declines in net uptake due to increased respiration in middle and late summer, for example, NOBS/BOREAS [Dunn et al., 2007].

[40] The shrubland site (Lucky Hill) had the worst fit. Carbon dioxide exchanges at this site derive from both organic and soil inorganic pools [Emmerich, 2003]; the latter is beyond the scope of a model like the VPRM.

[41] We carried out VPRM simulations for 11 different validation sites (SOBS, B1850, EOBS, DUKE-PP, INDIANA, DUKE-HW, WCREEK, LCREEK, WLEF, BOND, and ANLGRASS) using derived model parameters from calibration sites in the same or similar vegetation classes, without any adjustment. SOBS, B1850 and EOBS were classified as old growth evergreen boreal forests, and model parameters were taken from NOBS. DUKE-PP was classified as evergreen dry temperate forest, and model parameters were taken from Metolius. INDIANA and DUKE-HW were classified as deciduous forest and model parameters were taken from Harvard Forest. WCREEK, LCREEK, and WLEF were classified as mixed forest, and model parameters were taken from Howland Forest, although this classification is not completely accurate for WCREEK (a young maple stand) and LCREEK (a major fraction is wetland). Soy and corn at BOND and ANLGRASS were validated using Mead-S2 soy and corn and VAIRA model parameters, respectively. We were not able to test the VPRM independently for other vegetation classes because of lack of tower data.

[42] Most validation simulations were very successful. Figure 4 shows that the diurnal variation of NEE was slightly underestimated at B1850, DUKE-HW, and ANLGRASS and slightly overestimated at BOND-soy and LCREEK. ANLGRASS nighttime respiration was notably underestimated. Figure 5 shows the seasonal variation of NEE at validation sites, and associated VPRM functions, as in Figure 3. Seasonal peaks of NEE were slightly overestimated at DUKE-PP, INDIANA, ANLGRASS, and BOND-SOY.

[43] Overall, when model parameters from calibrated sites were applied to similar ecosystems for validation (Table 3), \( r^2 \) values were almost as high as at calibration sites, demonstrating strong predictive ability for sites with similar vegetation. WLEF was an outlier. Several studies have noted [Desai et al., 2008; Wang et al., 2006] the sharp differences between WLEF fluxes versus WCREEK and LCREEK, which are not as similar to each other, or to Howland, as one would like. Mackay et al. [2002] compared WLEF stand types to IGBP classes and suggested that four distinct stand types are needed to characterize the region’s evapotranspiration fluxes. Possibly the great tower height affects resolution of surface fluxes, extends the area influencing the tower, or introduces measurement artifacts.

[44] The VPRM provides excellent prediction of monthly NEE for most calibration and validation sites (Figure 6), excluding WLEF. Since the optimization exclusively used hourly data, the excellent agreement between VPRM and observations at the monthly timescale (Table 3), representing aggregation by a factor of ~600 in time, indicates successful elimination of bias in the nonlinear optimized functions. Only one calibration site (Donaldson/slash pine) and two of the validation sites (ANL-grassland and Duke Ponderosa pine) fail to scale up in time. These are the sites

**Figure 3.** (top) Seasonal dynamics of the prefactors \( T_{\text{scale}}, P_{\text{scale}}, \) and \( W_{\text{scale}} \) in the VPRM equation (equation (11)), driven by satellite and meteorological data. All factors are significant, at various times and places. (bottom) Comparison of the seasonal dynamics between observed (black) and VPRM (red) monthly mean NEE (\( \mu \text{mole m}^{-2} \text{s}^{-1} \)) at calibration sites, monthly means, averaged over all years. Values in the title bar give the fraction of the total variance of the mean seasonal cycle of NEE captured by the model, \( [1 - \text{var(NEEobs} - \text{NEEVPRM)}]/\text{var(NEEobs)}] \) (in parenthesis, the same quantity for the time series of individual monthly means). Note that net uptake of CO\(_2\) corresponds to negative values of NEE.
Figure 4. Comparison between the observed (solid squares) and predicted (open circles) diurnal mean NEE ($\mu$ mole $m^{-2} s^{-1}$) over the peak photosynthetically active period, as in Figure 2b, but at validation sites without adjustment of parameters.
Figure 5. A comparison of the seasonal dynamics of model and observed NEE, as in Figure 3, but at validation sites (no adjustment of VPRM parameters).
are likely affected by water stress, which we already noted may not be accurately captured in the VPRM.

The VPRM validations did not capture the seasonal cycle as well at boreal evergreen forests (SOBS, B1850, and EOBS; see Table 3) as at other sites. These biomes exhibit an especially strong seasonal cycle of ecosystem respiration, controlled by subsurface processes such as slow thawing and draining of snowmelt-saturated soils [Dunn et al., 2007] that are not remotely sensible, as well as prior site history. Thus the VPRM cannot distinguish the late summer trends at these sites on the basis of the fit to the NOBS data.

Table 3. Correlation Coefficients for Monthly and Hourly NEE and Means for All Seasons and for the Growing Season Hourly Data of Tower Flux and VPRM Calculations at Calibration and Cross-Validation Sites

<table>
<thead>
<tr>
<th>Site</th>
<th>Calibration Site</th>
<th>Years</th>
<th>(r^2) Monthly</th>
<th>(r^2) Hourly</th>
<th>Mean NEE-All ((\mu) mole m(^{-2}) s(^{-1}))</th>
<th>Growing Seas ((\mu) mole m(^{-2}) s(^{-1}))</th>
</tr>
</thead>
<tbody>
<tr>
<td>HARVARD</td>
<td>-</td>
<td>4</td>
<td>0.96</td>
<td>0.83</td>
<td>-1.64 -1.70</td>
<td>-7.24 -7.50</td>
</tr>
<tr>
<td>HOWLAND</td>
<td>-</td>
<td>4</td>
<td>0.33</td>
<td>0.65</td>
<td>-0.59 -0.59</td>
<td>-1.32 -2.61</td>
</tr>
<tr>
<td>NOBS</td>
<td>-</td>
<td>4</td>
<td>0.83</td>
<td>0.72</td>
<td>-0.54 -0.59</td>
<td>-1.75 -1.96</td>
</tr>
<tr>
<td>NIWOT</td>
<td>-</td>
<td>4</td>
<td>0.25</td>
<td>0.56</td>
<td>-0.19 -0.19</td>
<td>-0.85 -1.13</td>
</tr>
<tr>
<td>METOLIUS</td>
<td>-</td>
<td>3</td>
<td>0.55</td>
<td>0.63</td>
<td>-0.99 -0.99</td>
<td>-1.66 -1.31</td>
</tr>
<tr>
<td>SOY_MEADS2</td>
<td>-</td>
<td>1</td>
<td>0.61</td>
<td>0.66</td>
<td>0.08 0.05</td>
<td>-2.32 -2.05</td>
</tr>
<tr>
<td>CORN_MEADS2</td>
<td>-</td>
<td>1</td>
<td>0.94</td>
<td>0.83</td>
<td>-1.54 -1.58</td>
<td>-8.13 -9.42</td>
</tr>
<tr>
<td>TONZI</td>
<td>-</td>
<td>3</td>
<td>0.57</td>
<td>0.43</td>
<td>-0.59 -0.59</td>
<td>-1.22 -0.81</td>
</tr>
<tr>
<td>VAIRA</td>
<td>-</td>
<td>3</td>
<td>0.44</td>
<td>0.55</td>
<td>-0.42 -0.43</td>
<td>-1.33 -2.43</td>
</tr>
<tr>
<td>DONALDSON</td>
<td>-</td>
<td>2</td>
<td>-1.04</td>
<td>0.82</td>
<td>-1.49 -1.52</td>
<td>-1.15 -2.12</td>
</tr>
<tr>
<td>LUCKY-HILLS</td>
<td>-</td>
<td>4</td>
<td>0.36</td>
<td>0.46</td>
<td>0.02 0.01</td>
<td>0.31 0.74</td>
</tr>
<tr>
<td>PEATLAND</td>
<td>-</td>
<td>1</td>
<td>0.50</td>
<td>0.71</td>
<td>-0.04 -0.04</td>
<td>-0.77 -1.06</td>
</tr>
<tr>
<td>SOBS</td>
<td>NOBS</td>
<td>5</td>
<td>0.81</td>
<td>0.69</td>
<td>-0.88 -1.23</td>
<td>-2.04 -2.83</td>
</tr>
<tr>
<td>EOBSS</td>
<td>NOBS</td>
<td>1</td>
<td>0.88</td>
<td>0.74</td>
<td>-0.51 -0.29</td>
<td>-1.91 -1.70</td>
</tr>
<tr>
<td>B1850</td>
<td>NOBS</td>
<td>4</td>
<td>0.84</td>
<td>0.62</td>
<td>-0.66 -0.57</td>
<td>-1.97 -1.69</td>
</tr>
<tr>
<td>DUKE_PP</td>
<td>METOLIUS</td>
<td>4</td>
<td>-0.43</td>
<td>0.58</td>
<td>-1.01 -2.63</td>
<td>-1.66 -4.89</td>
</tr>
<tr>
<td>DUKE_HW</td>
<td>HARVARD</td>
<td>4</td>
<td>0.64</td>
<td>0.58</td>
<td>-1.00 0.47</td>
<td>-3.59 -1.80</td>
</tr>
<tr>
<td>INDIANA</td>
<td>HARVARD</td>
<td>4</td>
<td>0.59</td>
<td>0.65</td>
<td>-0.70 -0.40</td>
<td>-4.35 -4.94</td>
</tr>
<tr>
<td>WCREEK</td>
<td>HOWLAND</td>
<td>5</td>
<td>0.77</td>
<td>0.77</td>
<td>-0.87 -1.78</td>
<td>-5.36 -7.95</td>
</tr>
<tr>
<td>LCREEK</td>
<td>HOWLAND</td>
<td>4</td>
<td>0.53</td>
<td>0.66</td>
<td>-0.27 -0.18</td>
<td>-2.12 -2.68</td>
</tr>
<tr>
<td>WLEF</td>
<td>HOWLAND</td>
<td>2</td>
<td>-1.10</td>
<td>0.46</td>
<td>0.26 0.37</td>
<td>-0.52 -3.94</td>
</tr>
<tr>
<td>SOY_BOND</td>
<td>SOY_MEAD</td>
<td>1</td>
<td>0.80</td>
<td>0.72</td>
<td>0.31 0.85</td>
<td>-2.62 -3.02</td>
</tr>
<tr>
<td>CORN_BOND</td>
<td>CORN_MEAD</td>
<td>1</td>
<td>0.76</td>
<td>0.63</td>
<td>-1.27 -0.53</td>
<td>-7.02 -9.03</td>
</tr>
<tr>
<td>ANLGRASS</td>
<td>VAIRA</td>
<td>2</td>
<td>-0.04</td>
<td>0.57</td>
<td>-0.34 -0.80</td>
<td>-0.52 -1.32</td>
</tr>
</tbody>
</table>

*a* Only intersection data (available in both observation and model columns) were used. Growing season hourly data: April to June for VAIRA and ANLGRASS; June to August for all other sites. Correlation coefficients: \(r^2\).

Figure 6. (left) Observed and predicted monthly mean NEE (\(\mu\) mole m\(^{-2}\) s\(^{-1}\)) for calibration sites (solid symbols) and validation sites (open symbols) excluding WLEF. Regression line for all sites (dotted line) is very similar to the regression for validation sites only (dashed line). (right) Mean NEE by site (except WLEF) for the growing season. Line labeled (0,1) has zero intercept and slope = 1 ("1:1 line"). Regression lines are labeled similarly.
We quantified the role of satellite vegetation indices and of the temperature function for photosynthesis ($T_{scale}$) using a series of reduced models. Each was optimized independently using NEE data for Harvard and NOBS, then compared to the VPRM:

$$NEE_{model-2} = -\lambda' \times T_{scale} \times \frac{1}{(1 + PAR/\overline{PAR}_0)} \times EVI \times PAR$$

$$+ \alpha' \times T + \beta'$$

(13)

$$NEE_{model-1} = -\lambda'' \times T_{scale} \times \frac{1}{(1 + PAR/\overline{PAR}_0)} \times PAR$$

$$+ \alpha'' \times T + \beta''$$

(14)

$$NEE_{model-0} = -\lambda''' \times \frac{1}{(1 + PAR/\overline{PAR}_0)} \times PAR + \alpha''' \times T + \beta'''$$

(15)

Model-2 deletes the water and phenology scaling factors using LSWI, Model-1 deletes all satellite information (LSWI and EVI), and Model-0 deletes these and also drops $T_{scale}$.

Figure 7 compares GEE (µmole m$^{-2}$ s$^{-1}$) obtained by fitting VPRM, MODEL-2, MODEL-1, and MODEL-0 to tower data from 2000 to 2003 at Harvard Forest (a) and NOBS/BOREAS (b). VPRM incorporates EVI, $P_{scale}$, and $W_{scale}$ (driven by satellite data), plus $T_{scale}$ driven by meteorological data. MODEL-2 drops LSWI factors ($P_{scale}$ and $W_{scale}$), MODEL-1 drops all satellite $T_{scale}$.

5. Discussion

This paper develops and validates the VPRM, a satellite-based vegetation photosynthesis and respiration model, intended to provide NEE over North America with fine temporal and spatial resolution. The model has very simple structure and few adjustable parameters. It was tested using observations from all across the AmeriFlux and Fluxnet-Canada networks. When combined with maps of vegetation type, meteorological data for temperature, and satellite-derived shortwave radiation, it provides an excellent a priori representation of surface CO$_2$ fluxes, with
hourly time resolution and spatial resolution equal to that of the vegetation data (1 km for the IGBP).

[50] There are many process-based biogeochemical models (e.g., SiB2 or Biome-BGC) that simulate the storage and fluxes of water, carbon, and nitrogen by vegetation, litter, and soil. They can provide estimates of net primary production (NPP) or gross primary production (GPP), and in some cases, NEE, with hourly resolution. However, these models require complex parameter specification. For example, 47 parameters were spatially interpolated for regional simulations of SiB2 [Wang et al., 2007a]. In many cases, model parameters need frequent recalibration within short time periods, and the models may incur significant computational effort.

[51] The data-driven approach of the VPRM is capable of reproducing spatial and temporal variations of NEE using simple equations plus a compact database derived from MODIS. There are only four parameters per vegetation type that persist for the whole annual cycle, with spatial and temporal variations rendered by high-resolution meteorological and remote sensing data.

[52] Statistical uncertainties in the VPRM are given in Tables 2 and 3. Important additional systematic errors arise in part from the model structure. The lack of a soil moisture component and inability to remotely sense water stress are discussed above. Errors also arise because of limited resolution in the vegetation classification. Calibration and validation sites do not have identical vegetation assemblages, and the landscape includes assemblages not represented at all in present networks (e.g., northern white pine forests and loblolly pine plantations). Differences in vegetation functional responses are associated with climate, soil properties and soil moisture, canopy structure, necromass, and tree ages and distribution, none of which can currently be resolved using tower site data. Related errors arise from misclassification by the IGBP (e.g., at LCREEK).

[53] Noise in MODIS data also introduces significant errors in EVI and LSWI, and the noisy time series of MODIS data leads to errors in phenology. Notably large errors in model NEE accrue because of deficiencies in the driver data (sunlight and temperature), affecting CO2 flux predictions from all surface flux models. Detailed studies of errors in driver data will be described in a subsequent paper.

6. Conclusions

[54] The VPRM assimilates large amounts of data from remote sensing, meteorology, and flux towers and compresses the acquired knowledge into just four parameters in each vegetation class. Vegetation indices (EVI and LSWI) from the MODIS sensor, representation of the temperature dependence of photosynthesis, and accurate driver data are all that is required to describe the hourly and seasonal dynamics of NEE across the landscape. When coupled to accurate data sets for these factors, the VPRM partitions NEE into GEE (light-dependent) and R (light-independent) without complex algorithms or submodels, and with a minimal resort to arbitrary assumptions. The four parameters of the VPRM drive a model with very simple structure that demonstrates strong predictive ability for NEE from hourly to monthly timescales.

[55] The selected calibration and validation sites provide a minimal representation of the vegetation of North America. At present, over 200 eddy flux tower sites make up a global FLUXNET network (http://www.daac.ornl.gov/FLUXNET). Data for CO2, H2O, and energy flux for numerous ecosystem types have been accumulated; but availability of quality-assured data has not kept pace. Once multiyear data from more eddy flux tower sites are available, the VPRM can be refined and extended across a wider range of ecosystem and climate and soil conditions, and to other continents. Enhanced vegetation classification, including stand ages, would provide the basis for further improvements while conserving the simple structure of the VPRM.

[56] The VPRM can be applied at the scale of North America, providing a detailed representation of the spatio-temporal variation of CO2 fluxes across the landscape, with a low-dimensional parameter space for optimization in an inverse model framework. The calibrated model coefficients (λ, PAR0, α, and β) represent a priori parameter estimates. We envision the principal application of the VPRM to be reoptimization of the parameters at local, regional, or continental scales in top down analyses of carbon fluxes. The model and underlying databases are publicly available at (http://www-as.harvard.edu/data/).

[57] Acknowledgments. We would like to thank the flux site investigators for providing their data through AmeriFlux and Fluxnet-Canada programs. This study was supported at Harvard University by a grant from the National Science Foundation Bio-complexity in the Environment Program (ATM-0221850) and by grants from the U.S. Department of Energy in the Terrestrial Carbon Program, grant DE-FG02-98ER62695, and the Northeast Regional Center (NERC) of the National Institute for Global Environmental Change (NIGEC) under cooperative agreement DE-FC02-03ER63613, also by NASA grants NAG5-11154 and NNG05GA76F from the Terrestrial Ecological Program. We also thank William E. Emmerich for providing Lucky Hills flux data from the USDA-ARS Agriflux Carbon project. The research at Duke Forest was supported by the Office of Science (BER), U.S. Department of Energy, grant DE-FG02-00ER63015 (Hardwood Forest), and through its Southeast Regional Center (SERC) of the NIGEC under cooperative agreement DE-FC02-03ER63613 (Pine Plantation). The research at the Metolius ponderosa pine site was supported by the Office of Science (BER), U.S. Department of Energy, grant DE-FG0203ER63653.

References
