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# Cumulative exposure to environmental pollutants during early pregnancy and reduced fetal growth: the Project Viva cohort

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## Abstract

**Background:** Reduced fetal growth is associated with perinatal and later morbidity. Prenatal exposure to environmental pollutants is linked to reduced fetal growth at birth, but the impact of concomitant exposure to multiple pollutants is unclear. The purpose of this study was to examine interactions between early pregnancy exposure to cigarette smoke, traffic pollution, and select perfluoroalkyl substances (PFASs) on birth weight-for-gestational age (BW/GA).

**Methods:** Among 1597 Project Viva mother-infant pairs, we assessed maternal cigarette smoking by questionnaire, traffic pollution at residential address by black carbon land use regression model, and plasma concentration of select PFASs in early pregnancy. We calculated sex-specific BW/GA z-scores, an index of fetal growth, from national reference data. We fit covariate-adjusted multi-pollutant linear regression models and examined interactions between exposures, using a likelihood-ratio test to identify a best-fit model.

**Results:** Two hundred six (13%) mothers smoked during pregnancy. Mean [standard deviation (SD)] for black carbon was 0.8 (0.3)  $\mu\text{g}/\text{m}^3$ , perfluorooctane sulfonate (PFOS) was 29.1 (16.5)  $\text{ng}/\text{mL}$ , and BW/GA z-score was 0.19 (0.96). In the best-fit model, BW/GA z-score was lower in infants of mothers exposed to greater black carbon [ $-0.08$  (95% CI:  $-0.15$ ,  $-0.01$ ) per interquartile range (IQR)]. BW/GA z-score (95% CI) was also lower in infants of mothers who smoked [ $-0.09$  ( $-0.23$ ,  $0.06$ )] or were exposed to greater PFOS [ $-0.03$  ( $-0.07$ ,  $0.02$ ) per IQR], although confidence intervals crossed the null. There were no interactions between exposures. In secondary analyses, instead of PFOS, we examined perfluorononanoate (PFNA) [mean (SD): 0.7 (0.4)  $\text{ng}/\text{mL}$ ], a PFAS more closely linked to lower BW/GA in our cohort. The best-fit multi-pollutant model included positive two-way interactions between PFNA and both black carbon and smoking ( $p$ -interactions = 0.03).

**Conclusions:** Concurrent prenatal exposures to maternal smoking, black carbon, and PFOS are additively associated with lower fetal growth, whereas PFNA may attenuate associations of smoking and black carbon with lower fetal growth. It is important to examine interactions between multiple exposures in relation to health outcomes, as effects may not always be additive and may shed light on biological pathways.

**Keywords:** Pregnancy, Birth weight, Epidemiology, Air pollution, Smoking, Perfluoroalkyl substances

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## Background

Fetal growth restriction is a major public health challenge, as it occurs in 5 to 10% of pregnancies and is one of the leading causes of neonatal mortality and morbidity [1]. Smaller weight for gestational age at birth, an indicator of fetal growth restriction, not only influences early-life outcomes, but is associated with adverse cardiometabolic health later in life [2].

Several in utero environmental exposures, many of which are modifiable, are linked to reduced fetal growth. For example, prenatal exposures to cigarette smoke, traffic pollution, and perfluoroalkyl substances (PFASs) have been associated with lower fetal growth in our greater Boston area longitudinal birth cohort, Project Viva [3–5], as well as in other epidemiological studies [6–10]. Cigarette smoke and traffic-related air pollution contain particulate matter (PM), along with hydrocarbons, aldehydes, nitric oxides, and carbon monoxide [11]. Both exposures are hypothesized to lower fetal growth primarily as a result of placental oxidative stress and inflammation [12, 13], which may disturb nutrient and oxygen exchange from mother to fetus [12]. PFASs are synthetic compounds used in the manufacture of a variety of consumer and industrial products, including water- and stain-resistant coatings for clothing, furniture, and carpets; oil-resistant coatings for food packaging; non-stick cookware; fire-fighting foams; and aviation hydraulic fluids. Almost the entire U.S. population (>95%) has detectable serum concentrations of several PFAS [14–16]. PFASs may influence fetal growth by activating the peroxisome proliferator-activated receptor (PPAR) [17], which may inhibit inflammatory gene expression [18, 19] and interfere with lipid and glucose homeostasis [19, 20].

Although studies have reported robust associations of each of these prenatal exposures with reduced fetal growth, few prior studies have evaluated the potential impact of these pollutants combined. This is important because pregnant women are simultaneously exposed to mixtures of environmental toxicants [21].

In the present analysis, we examine additive interactions between early pregnancy exposure to cigarette smoke, black carbon (an indicator of traffic pollution), and select PFASs and their associations with birth weight-for-gestational age (BW/GA) in the Project Viva cohort. Based on our prior work in subsets of this cohort, we expected to find that each exposure would be independently associated with fetal growth and further hypothesized that infants exposed to all three would have the lowest expected fetal growth.

## Methods

### Study population and design

Between 1999 and 2002, we recruited pregnant women into the Project Viva cohort during their initial prenatal

visit (median 9.9 weeks gestation) at Atrius Harvard Vanguard Medical Associates, a multi-specialty group practice with offices throughout eastern Massachusetts. Exclusion criteria included multiple gestation, inability to answer questions in English, gestational age greater than or equal to 22 weeks at recruitment, and plans to move from the area prior to delivery [22]. A total of 2128 women were initially included in the cohort. In the present analysis, we excluded one participant without a recorded birth weight, 21 without information on prenatal cigarette smoking, 37 without data on estimated residential black carbon concentration, one for whom the black carbon regression model provided an implausible data point, and 471 whose mother did not provide a plasma sample or did not have sufficient sample for quantification of PFASs and thus did not have a measurement of maternal perfluorooctane sulfonate (PFOS) plasma concentration. In total, we studied 1597 participants in the present analysis. Compared with those not included ( $n = 531$ ), mothers included in the present analysis were more likely to be white, but other characteristics were similar (Additional file 1: Table S1). We obtained written informed consent from all participants, and the study was approved by Institutional Review Boards of participating institutions. The involvement of the Centers for Disease Control and Prevention (CDC) laboratory did not constitute engagement in human subject research.

### Selection and measurement of early pregnancy environmental exposures

We collected information on maternal cigarette smoking during pregnancy and categorized participants into three groups: 1) prenatal smokers, 2) former smokers, and 3) never smokers. At the first in-person study visit, we asked expectant mothers whether they had ever smoked, and those who smoked less than 100 cigarettes in their lifetime were classified as “never smokers.” We asked smokers whether they had smoked in the three months before learning of their pregnancy, and classified women who quit at least three months prior to pregnancy as “former smokers.” We designated those who smoked during the three months before learning of their pregnancy (i.e., likely smoked in very early pregnancy) or reported smoking on the first or second trimester questionnaire or delivery interview as “prenatal smokers.” The majority (65%) of prenatal smokers quit during the first trimester, and thus were primarily early pregnancy smokers. Therefore, in secondary analyses, we also considered the subset of prenatal smokers who continued smoking beyond the first trimester and termed this group “continued smokers.”

We estimated prenatal exposure to several types of traffic pollution in Project Viva [black carbon,  $PM_{2.5}$

( $PM \leq 2.5 \mu m$  in aerodynamic diameter), neighborhood traffic density, home roadway proximity] [3]. In the present study, we evaluated prenatal exposure to black carbon because this measure of the black particles primarily emitted by traffic has been more strongly associated with health risks such as mortality and childhood asthma, as compared to other traffic-related pollution indicators like PM [23]. In addition, in Project Viva, black carbon was more strongly associated with low fetal growth than other markers of traffic pollution [3]. We estimated daily ambient black carbon exposure at the address each participant provided at the time of enrollment (median 10 weeks gestation). We used a validated spatiotemporal regression model, which included measures of land use, GIS location, daily meteorological factors, day of week and season, and interaction terms of land use measures and meteorological factors as predictors [24, 25]. An “out-of-sample” cross-validation showed a correlation of 0.73 between predicted and monitor-observed daily black carbon concentration [25]. We analyzed first trimester black carbon [average daily exposures from date of the last menstrual period (LMP) to the 93rd day after LMP] for consistency in timing with the other early pregnancy exposures in this analysis.

We also measured maternal plasma concentrations of several PFASs [PFOS, perfluorooctanoate (PFOA), perfluorohexane sulfonate (PFHxS), perfluorononanoate (PFNA), 2-(*N*-ethyl-perfluorooctane sulfonamido) acetate, 2-(*N*-methyl-perfluorooctane sulfonamido) acetate, perfluorodecanoate, and perfluorooctane sulfonamide] in early pregnancy in Project Viva [26]. We evaluated PFOS in our primary analysis because as compared to other measured PFASs, mothers in Project Viva had the highest average concentration of PFOS plasma concentrations. Likewise, PFOS has consistently been the PFAS with the highest mean serum concentrations observed in the U.S. general population since these chemicals were in peak production and detected in the National Health and Nutrition Examination Survey’s participants in 1999–2000 [15, 27]. In an a priori secondary analysis, we examined PFNA because we have previously observed the strongest inverse association between PFNA and fetal growth as compared to other PFASs in our cohort [4]. As compared to PFOS, PFNA has a narrower range of concentrations (i.e., smaller IQR and lower mean concentration), but the coefficients of variation are relatively similar (i.e., 56.9 for PFOS and 55.1 for PFNA). Although PFOA is one of the most common studied PFAS, we found null associations between maternal PFOA plasma concentration and BW/GA in Project Viva and thus did not evaluate PFOA in the present analysis. We collected maternal plasma samples during the first prenatal visit, and samples were shipped to the CDC

(Atlanta, GA) for analysis, as previously described [26]. Briefly, staff at CDC quantified PFAS concentration in maternal plasma using on-line solid-phase extraction coupled to isotope dilution high performance liquid chromatography mass spectrometry [28]. The limit of detection (LOD) was 0.2 ng/mL for PFOS and 0.1 ng/mL for PFNA; we replaced values below the LOD with the  $LOD/\sqrt{2}$ . Two (0.1%) women had PFOS plasma concentrations below the LOD and 22 (1.4%) women had PFNA plasma concentrations below the LOD.

#### Measurement of outcome

To compute BW/GA z-score, we abstracted birth weight (in grams) and date of delivery from the hospital medical record. We obtained infant sex from delivery interview. We calculated length of gestation (in days) by subtracting the date of the LMP from the date of delivery, except in cases in which the length of gestation based on the second trimester ultrasound was more than 10 days different from that by the LMP. In these cases, we used the ultrasound result. Using these data, we calculated sex-specific BW/GA z-score based on U.S. national reference data [29].

#### Measurement of covariates

We collected information on maternal age, education, race/ethnicity, parity, and household income using questionnaires at study enrollment. We calculated prepregnancy body mass index (BMI) from self-reported weight and height. As there is potential for the relationship between PFAS concentrations and fetal growth to be confounded by pregnancy physiology [30], we assessed creatinine in the same maternal plasma sample used for PFAS quantification to calculate glomerular filtration rate (GFR) and measured plasma albumin to assess plasma expansion, as described previously [26].

#### Statistical analyses

We first fit unadjusted, followed by covariate-adjusted, linear regression models to examine the association of each prenatal exposure (self-reported maternal prenatal smoking, residential black carbon levels during the first trimester, and PFOS plasma concentration during early pregnancy) individually with BW/GA z-score. Next, we fit a covariate-adjusted additive model that included all three exposures simultaneously. For smoking status, we combined the never and former smoker categories because characteristics of these women were similar and combining the two categories did not appreciably change results. To ease interpretability and comparison, we present both first trimester black carbon and prenatal PFOS plasma concentration associations per interquartile range (IQR). We adjusted for covariates potentially associated with the three prenatal environmental

exposures [3, 26, 31] and/or fetal growth [32]. These covariates included maternal age at enrollment (continuous), maternal pre-pregnancy BMI (continuous), race/ethnicity (white, black, Hispanic, Asian, or other), education (with or without a college degree), parity (nulliparous or multiparous), season of birth (continuous sine and cosine of date), and date of birth (continuous). Inclusion of household income, maternal GFR, plasma albumin, and week of gestation did not considerably change effect estimates (> 10%) and were not significant in adjusted models, and thus we did not include these covariates in final models. We did not adjust for gestational glucose tolerance status or gestational weight gain because these may be on the causal pathway between early pregnancy exposures and fetal growth and therefore may introduce collider bias [33]. Ninety-nine percent of participants ( $n = 1577$ ) had complete covariate information for the multivariate regression models.

Next, we evaluated interactions between the three early pregnancy environmental exposures in the adjusted model. Using a sequential model building approach, we fit covariate-adjusted multi-pollutant models to examine all possible two-way interactions between pollutants (smoking\*black carbon, smoking\*PFOS, and black carbon\*PFOS). Using a likelihood ratio test, we calculated a chi-square  $p$ -value to compare each two-way interaction model separately to the main effects additive model. If the likelihood ratio test was significant ( $p < 0.05$ ), indicating that the more complex model was a better fit, we used this model as the new likelihood ratio test reference as we continued testing other models with one or multiple two-way interactions, and eventually compared it to a final model with all two-way interactions and a three-way interaction (smoking\*black carbon\*PFOS). We thus identified a final best-fit model and present individual and interaction effect estimates.

Although dichotomizing exposures and outcomes can improve interpretability of interaction effect estimates [34], we opted to represent the black carbon and PFAS exposures and the outcome (BW/GA z-score) as continuous variables to maximize power. Then, to improve interpretability, we used the best-fit model to predict the BW/GA z-score of an infant born to a “typical” Project Viva mother [based on mean age (31.8 years) and pre-pregnancy BMI (25.0 kg/m<sup>2</sup>) and modal education (college graduate), race/ethnicity (white), and parity (multiparous) woman] according to select levels of exposure to the three environmental toxicants in early pregnancy. We present extreme high and low exposure scenarios to capture the full range of model-predicted birth size in infants of exposed versus unexposed mothers, although most mothers have more moderate exposures. We considered “high” exposure to be 1) prenatal smoking, 2) first trimester black carbon exposure at the 90th

percentile (1.08  $\mu\text{g}/\text{m}^3$ ), and 3) prenatal PFOS plasma concentration at the 90th percentile (47.0 ng/mL). We considered “low” exposure to be 1) no prenatal smoking, 2) first trimester black carbon exposure at the 10th percentile (0.43  $\mu\text{g}/\text{m}^3$ ), and 3) prenatal PFOS plasma concentration at the 10th percentile (14.0 ng/mL). We predicted infant BW/GA z-score for different mixture combinations of high versus low prenatal exposures. Next, we used birth weight percentiles for a white infant born at term from U.S. national reference data [29] to convert the predicted BW/GA z-scores to predicted birth weight.

In a secondary analysis, we re-ran the model building process but considered prenatal smokers to be the subset of women who smoked throughout pregnancy [continued smokers:  $n$  (%) = 71 (35%)] with all other women (including those who quit smoking during the first trimester) as the reference group. As compared to early pregnancy smoking, continued smoking throughout pregnancy has been more strongly associated with low birth size [6, 7, 35], but we considered continued smokers in the secondary rather than primary analysis due to limited power to detect an interaction based on the relatively small number of continued smokers. We also re-ran the model building process described above considering PFNA rather than PFOS because in Project Viva it was more strongly associated with smaller BW/GA.

We used the *gmodels* package in R 3.3.2 (Vienna, Austria) to calculate predicted BW/GA z-scores. For all other analyses, we used SAS version 9.4 (Cary, NC).

## Results

Of the 1597 mothers included in the analytic dataset, mean (standard deviation [SD]) maternal age at the time of enrollment was 31.8 (5.2) years. Sixty-five percent of mothers were college graduates, 49% nulliparous, 69% white, and 13% prenatally smoked. Mean (SD) residential black carbon level during the first trimester was 0.8 (0.3)  $\mu\text{g}/\text{m}^3$ , prenatal PFOS plasma concentration was 29.1 (16.5) ng/mL, and BW/GA z-score was 0.19 (0.96) (Table 1).

As compared to non-smokers, mothers who smoked during early pregnancy were less educated, younger, had a higher pre-pregnancy BMI, and were less likely to be white. As compared to prenatal smokers, non-smokers had slightly lower exposure to black carbon (mean 0.80  $\mu\text{g}/\text{m}^3$  for smokers versus 0.75  $\mu\text{g}/\text{m}^3$  for non-smokers;  $p < 0.01$ ), no difference in PFOS plasma concentrations (mean 29.6 ng/mL for smokers versus 29.0 ng/mL for non-smokers;  $p = 0.65$ ), and no difference in PFNA plasma concentrations (mean 0.7 ng/mL for smokers versus 0.8 ng/mL for non-smokers;  $p = 0.06$ ) (Table 1). Mean residential black carbon concentration

**Table 1** Participant characteristics overall ( $n = 1597$ ) and by maternal smoking status, quartiles of first trimester residential black carbon, and quartiles of prenatal perfluorooctane sulfonate (PFOS) plasma concentration

	Prenatal Environmental Exposures										
	Overall	Prenatal smoking		First trimester residential black carbon <sup>a</sup>				PFOS plasma concentration <sup>b</sup>			
			No	Yes	Q1	Q2	Q3	Q4	Q1	Q2	Q3
	$n = 1597$	$n = 1391$	$n = 206$	$n = 399$	$n = 399$	$n = 400$	$n = 399$	$n = 400$	$n = 396$	$n = 403$	$n = 398$
	Mean (SD) or %	Mean (SD) or %	Mean (SD) or %	Mean (SD) or %	Mean (SD) or %	Mean (SD) or %	Mean (SD) or %	Mean (SD) or %	Mean (SD) or %	Mean (SD) or %	Mean (SD) or %
<b>Maternal characteristics</b>											
Age at enrollment (years)	31.8 (5.2)	32.2 (5.0)	29.2 (5.8)	32.9 (4.3)	32.2 (5.2)	31.5 (5.4)	30.7 (5.6)	32.5 (5.1)	31.8 (5)	31.6 (5.3)	31.5 (5.3)
Pre-pregnancy BMI (kg/m <sup>2</sup> )	25.0 (5.6)	24.9 (5.4)	26.1 (6.2)	25.1 (5.5)	25.0 (5.1)	25.1 (5.8)	24.9 (5.7)	24.4 (5.3)	24.8 (5.4)	25.5 (5.8)	25.3 (5.6)
College graduate (%)	65	69	34	75	70	59	56	73	62	64	61
Nulliparous (%)	49	48	55	43	45	50	57	37	45	54	60
<b>Race/ethnicity (%)</b>											
White	69	69	65	86	73	65	51	70	71	66	67
Black	15	16	14	7	13	18	25	12	14	18	18
Hispanic	7	6	11	4	6	7	12	9	7	7	6
Asian	5	5	4	3	6	4	7	5	3	5	6
Other	4	4	6	1	4	7	5	4	5	4	3
<b>Maternal prenatal exposures</b>											
Prenatal smoking (%)	13	0	100	12	10	14	17	11	13	14	14
First trimester black carbon ( $\mu\text{g}/\text{m}^3$ )	0.8 (0.3)	0.7 (0.3)	0.8 (0.3)	0.4 (0.1)	0.7 (0.1)	0.8 (0.0)	1.1 (0.2)	0.7 (0.2)	0.8 (0.3)	0.8 (0.3)	0.8 (0.3)
PFOS plasma concentration (ng/mL)	29.1 (16.5)	29.0 (16.9)	29.6 (14.1)	29.1 (16.8)	28.3 (15.2)	29.1 (16.9)	30.7 (5.6)	14.3 (3.6)	22.3 (1.9)	29.8 (2.7)	50.0 (19.2)
PFNA plasma concentration (ng/mL)	0.7 (0.4)	0.8 (0.4)	0.7 (0.3)	0.7 (0.4)	0.7 (0.4)	0.7 (0.5)	0.8 (0.4)	0.5 (0.3)	0.6 (0.2)	0.8 (0.3)	1.0 (0.5)
<b>Infant characteristics</b>											
Female (%)	48	48	45	47	48	48	47	48	48	49	46
BW/GA z-score	0.19 (0.96)	0.22 (0.94)	0.02 (1.08)	0.36 (0.92)	0.24 (0.96)	0.12 (0.97)	0.03 (0.98)	0.30 (0.98)	0.18 (0.98)	0.17 (0.95)	0.10 (0.94)
Birth weight (g)	3474 (591)	3484 (585)	3403 (626)	3560 (571)	3519 (532)	3424 (644)	3391 (597)	3562 (544)	3457 (652)	3477 (544)	3398 (609)
Gestational age (weeks)	39.4 (1.9)	39.4 (2.0)	39.5 (1.8)	39.5 (1.9)	39.5 (1.6)	39.4 (2.2)	39.4 (2.0)	39.6 (1.5)	39.4 (2.3)	39.5 (1.7)	39.2 (2.2)

**Abbreviations** BMI Body mass index, BW/GA birth weight-for-gestational age, PFOS perfluorooctane sulfonate, PFNA perfluorononanoate, Q1 to 4 Quartiles 1 to 4

Missing in overall cohort: 12 pre-pregnancy BMI, 16 education & race/ethnicity

<sup>a</sup>First trimester black carbon quartile ranges: Q1 (0.13–0.56  $\mu\text{g}/\text{m}^3$ ); Q2 (0.57–0.75  $\mu\text{g}/\text{m}^3$ ); Q3 (0.76–0.92  $\mu\text{g}/\text{m}^3$ ); and Q4 (0.93–1.71  $\mu\text{g}/\text{m}^3$ )

<sup>b</sup>PFOS plasma concentration quartile ranges: Q1 [ $<$ limit of detection (0.1)–18.8 ng/mL]; Q2 (18.9–25.6 ng/mL); Q3 (25.7–34.9 ng/mL); and Q4 (35.0–185.0 ng/mL)

was not correlated with PFAS plasma concentrations [Spearman's  $r = 0.02$  for PFOS ( $p = 0.35$ ) and PFNA ( $p = 0.54$ )].

In covariate-adjusted single-pollutant models, higher black carbon exposure during the first trimester was associated with a 0.08 reduction in BW/GA z-score [95% confidence interval (CI):  $-0.15$ ,  $-0.01$ ] per IQR increment (Table 2). Also, higher maternal PFOS plasma concentration and prenatal smoking were associated with lower fetal growth, although confidence intervals crossed the null [BW/GA z-score of  $-0.03$  (95% CI:  $-0.07$ ,  $0.02$ ) per IQR increment in PFOS and

$-0.09$  (95% CI:  $-0.24$ ,  $0.05$ ) for infants of smokers versus non-smokers] (Table 2).

In the covariate-adjusted multi-pollutant main effects additive model, effect sizes and significance for all three exposures in relation to fetal growth were similar to those in the single-pollutant models (Table 2).

None of the two-way interaction terms were significantly associated with fetal growth, and the likelihood ratio test confirmed that the multi-pollutant additive model without any interaction terms was the best-fit model (Table 2). Additionally, the three-way interaction was not statistically significant [ $0.16$  (95% CI:  $-0.03$ ,  $0.35$ )] (data not shown).

**Table 2** Covariate-adjusted associations<sup>a</sup> of prenatal smoking, first trimester residential black carbon, and prenatal perfluorooctane sulfonate (PFOS) plasma concentration and their interactions with birth weight-for-gestational age (BW/GA) z-score

	Change in BW/GA z-score (95% CI)	Smoking*BC Interaction $\beta$ (95% CI)	Smoking*PFOS	BC*PFOS	LRT $\chi^2$ p-value <sup>b</sup>
Single-pollutant exposure models					
Prenatal smoking (Y/N)	-0.09 (-0.24, 0.05)				
First trimester black carbon (per IQR) <sup>b</sup>	-0.08 (-0.15, -0.01)				
Prenatal PFOS plasma concentration (per IQR) <sup>b</sup>	-0.03 (-0.07, 0.02)				
Multi-pollutant exposure models					
Main effects (Additive model)	Smoking: -0.09 (-0.23, 0.06) BC: -0.08 (-0.15, -0.01) PFOS: -0.03 (-0.08, 0.02)				Reference
Main effects + Smoking*BC	Smoking: 0.04 (-0.39, 0.47) BC: -0.07 (-0.15, 0.00) PFOS: -0.03 (-0.08, 0.02)	-0.06 (-0.24, 0.13)			0.53
Main effects + Smoking*PFOS	Smoking: -0.25 (-0.57, 0.06) BC: -0.08 (-0.15, -0.01) PFOS: -0.04 (-0.09, 0.01)		0.09 (-0.06, 0.24)		0.24
Main effects + BC*PFOS	Smoking: -0.08 (-0.22, 0.06) BC: -0.18 (-0.31, -0.05) PFOS: -0.14 (-0.27, -0.01)			0.05 (-0.00, 0.11)	0.07 <sup>c</sup>

**Abbreviations**  $\chi^2$  chi-square, BC black carbon, BW/GA birth weight-for-gestational age, CI confidence interval, IQR interquartile range, LRT likelihood ratio test, PFOS perfluorooctane sulfonate

Estimates with 95% confidence intervals that do not cross the null are bolded. Likelihood ratio test (LRT) chi-square ( $\chi^2$ ) p-value < 0.05 indicates that the more complex model is a better fit than the reference model

<sup>a</sup>Adjusted for maternal age, race/ethnicity, education, pre-pregnancy body mass index, and parity. The first trimester black carbon single-pollutant exposure model was additionally adjusted for season and date of birth, and these covariates were included in all multi-pollutant exposure models

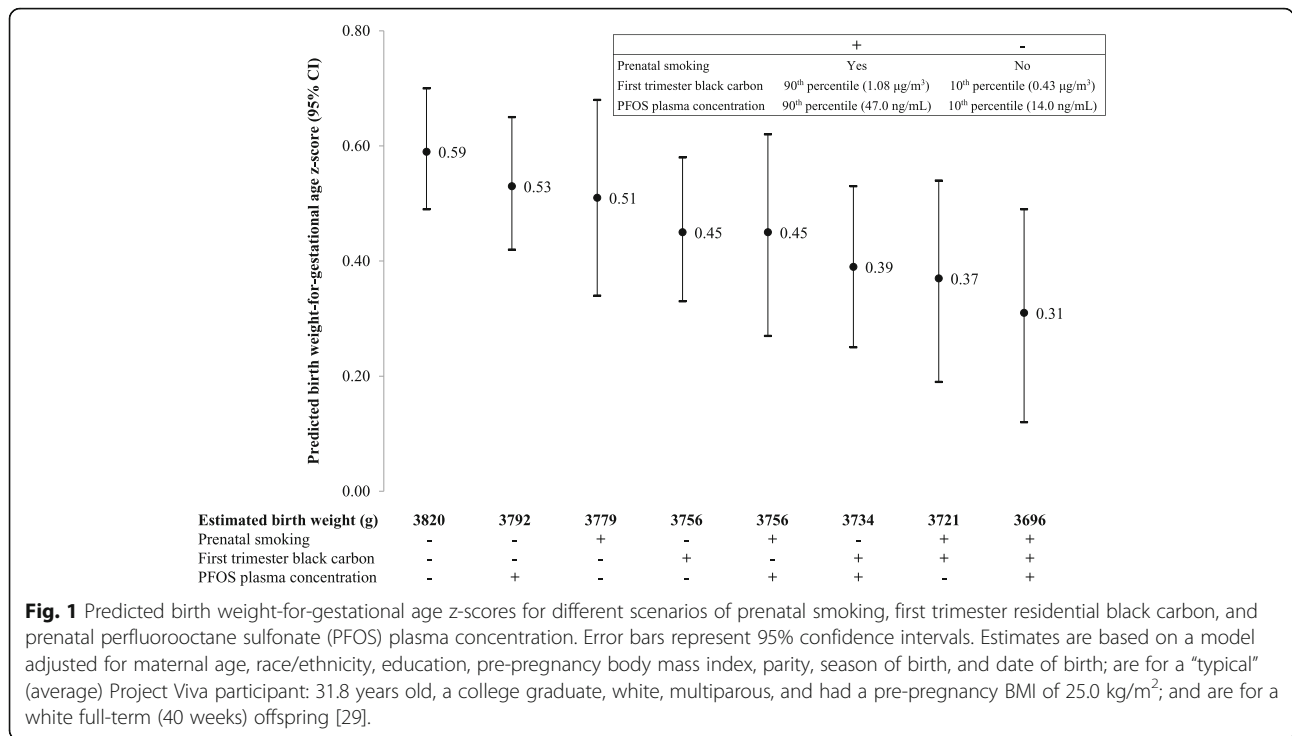
<sup>b</sup>Interquartile range increments: 0.36  $\mu\text{g}/\text{m}^3$  for first trimester black carbon and 16.1 ng/mL for prenatal PFOS plasma concentration

<sup>c</sup>When we subsequently compared models with multiple two-way interactions and the three-way interaction to the best-fit model (main effects additive model), all p-values were > 0.05

To aid in interpretability, we used the best-fit model to calculate predicted BW/GA z-scores for extreme exposure scenarios. As shown in Fig. 1, infants of mothers with low levels of all three exposures (10th percentile of black carbon and PFOS, and non-smokers) had a predicted BW/GA z-score of 0.59 (95% CI: 0.49, 0.70) versus 0.31 (95% CI: 0.12, 0.49) for infants of mothers with high levels of all exposures (90th percentile of black carbon and PFOS, and smokers), corresponding to an estimated birth weight difference of 124 g for a white infant born at term [29]. Infants with mixtures of some high and some low exposures had BW/GA z-scores between 0.31 and 0.59 (Fig. 1).

In secondary analyses, infants of mothers who smoked throughout pregnancy (continued smokers) had lower BW/GA z-scores [-0.40 (95% CI: -0.63, -0.18)] compared to infants of mothers who were non-smokers, former smokers, or early pregnancy smokers, in the best-fit main effects additive model (data not shown).

Maternal plasma concentration of PFNA was more strongly associated with lower BW/GA z-score as compared to PFOS [-0.05 (95% CI: -0.10, -0.01) per IQR increment in PFNA versus -0.03 (95% CI: -0.07, 0.02) per IQR increment in PFOS in the covariate-adjusted single-pollutant model]. The best-fit, multi-pollutant model included positive two-way interactions between PFNA plasma concentration and both prenatal smoking and black carbon exposure (Additional file 1: Table S2). Thus, the BW/GA z-score for infants of mothers with high PFNA concentration and exposed to high black carbon or with high PFNA concentration and exposed to maternal smoking, was not as low as expected based on the individual associations of each of the pollutants with fetal growth. For example, in the extreme exposure scenarios, infants of non-smoking mothers with low black carbon and high PFNA concentration had a predicted BW/GA z-score of 0.41 (95% CI: 0.27, 0.55) and infants of non-smoking mothers with high black carbon and low PFNA concentration had a predicted BW/GA z-



score of 0.43 (95% CI: 0.29, 0.58), but infants of non-smoking mothers with both high black carbon and high PFNA concentration had a predicted BW/GA z-score of 0.39 (95% CI: 0.25, 0.53) (Additional file 1: Figure S1).

**Discussion**

In our analysis of a large, prospective cohort of mother-infant pairs, we support previously reported findings of lower birth size with higher prenatal maternal exposure to smoking, black carbon, and PFOS in early pregnancy. We additionally found that simultaneous exposure to these compounds had an additive rather than interactive effect on fetal growth. However, infants prenatally exposed to PFNA and either maternal smoking or black carbon had a higher birth size than expected based on the individual associations with fetal growth.

Associations between higher prenatal exposures to maternal smoking, air pollution, and PFASs with lower birth size have been previously reported in the Project Viva cohort [3–5] and others [6–10]. In the analytic cohort of the present study, while all three exposures were associated with lower BW/GA, confidence intervals crossed the null for early pregnancy smoking and PFOS exposure. Other cohorts have shown that maternal smoking in mid to late but not early pregnancy leads to decrements in fetal growth [6, 7, 35], on the order of 175–200 g lower birth weight [6]. Consistent with this, in the Project Viva cohort, prenatal smoking was most strongly associated

with low fetal growth for women who smoked throughout pregnancy (versus those who did not), corresponding to a decrement of 170 g in birth weight for a full term white infant.

An association between higher prenatal air pollution and lower birth weight has been shown in numerous previous studies (e.g., [36–38]), with a pooled-effect of 23.4 g (95% CI: -45.5, -1.4) lower birth weight per 10 µg/m<sup>3</sup> of ambient PM<sub>2.5</sub> (of which black carbon is a component) in a recent meta-analysis [8]. For comparison, in the present analysis, we estimated a 36 g decrement in birth weight (for a white term infant at the 50th percentile of BW/GA z-score) for each IQR increment in first trimester black carbon exposure.

PFASs may reduce fetal growth via activation of PPAR or through structural similarity to free fatty acids which may modulate lipid and glucose homeostasis [20], impair placental transfer of lipids and glucose, and thereby lead to oxidative stress [39]. A meta-analysis of nine studies found a 19 g (95% CI: -30, -8) lower birth weight for each ng/mL increment in cord serum or maternal plasma PFOA concentration [10]. This translates to a 49 g lower birth weight for each IQR increment of PFOA using the IQR from U.S. females in NHANES during 1999–2000 [40]. The present study found more modest associations with PFOS and PFNA. For a white term infant in the 50th percentile of BW/GA z-score, we estimated a 10 g decrement in birth weight for each IQR increment in maternal PFOS plasma concentration, although confidence intervals included the null. We



estimated a larger, 19 g decrement in birth weight for each IQR increment in maternal PFNA plasma concentration.

The present study adds to the existing literature by examining the extent to which early pregnancy exposure to maternal smoking, black carbon, and select PFASs interact to alter fetal growth. We found that while PFOS did not modify the relationships of maternal smoking or black carbon with fetal growth, PFNA attenuated the associations of smoking and black carbon on lower BW/GA. This finding raises the possibility that PFNA may act antagonistically on similar biological pathways as maternal smoking and black carbon. For example, placental inflammation as a result of air pollution exposure during pregnancy may lead to impaired nutrient exchange to the fetus and lower fetal growth [12]. PFAS may counter pollution-induced placental inflammation by inhibiting inflammatory gene expression via activation of PPAR [18, 19], thus leading to a higher birth weight than otherwise expected from the concurrent exposures. Consistent with our finding of an interaction with PFNA but not PFOS, in mouse and human models, PFNA activated PPAR- $\alpha$  at lower doses and with higher potency than shorter chain carboxylates, such as PFOA, or sulfonates, like PFOS [41]. Nevertheless, our finding requires replication in other cohorts.

We found that although maternal smoking and black carbon independently lowered fetal growth, they did not act synergistically. The prior literature on the association between traffic pollution and birth outcomes for women prenatally exposed versus not exposed to cigarette smoke is conflicting and limited. Specifically, in a study of over 200,000 Canadian births, maternal smoking attenuated the association between air pollution exposure and low birth weight [42], whereas a study of over 600,000 U.S. births found the association between high prenatal traffic-related carbon monoxide exposure and low birth weight to be larger among smoking versus non-smoking mothers [43]. In a U.S. urban cohort of over 200 non-smoking pregnant women, passive smoke exposure and a blood marker of ambient air pollution (benzo[a]pyrene-DNA adducts) synergistically lowered birth weight, although there were not independent associations between either exposure with birth weight [44]. Even when we considered maternal smoking throughout pregnancy, which was more strongly associated with low fetal growth than early pregnancy smoking in the Project Viva cohort, we observed no interaction with black carbon on fetal growth. Although cigarette smoking impairs maternal immune function [45] and induces fetal hypoxia [13], which could potentially cause a fetus to be more vulnerable to the effects of other pollutants like black carbon, mothers in our cohort may not have been heavy enough smokers to significantly weaken responses

to these other toxicants. Alternatively, the fact that cigarette smoke and black carbon exposure impact fetal growth through similar mechanisms of action may explain their additive rather than synergistic or antagonistic mechanism of action. Furthermore, we may not have had the power to detect an interaction between maternal smoking and black carbon exposure on fetal growth because only a small proportion of the cohort smoked during pregnancy and variability of ambient black carbon concentrations in the Boston area is relatively low. Power to detect an interaction between these two exposures may have been further limited because of exposure misclassification. Social desirability bias may have introduced some misclassification into our questionnaire measure of maternal smoking, and we did not have data on work location or time at home which could have improved the accuracy of black carbon exposure estimates [46]. In contrast, PFAS exposure was less likely to be misclassified, as these compounds have long half-lives, and we measured concentrations in plasma.

Although our current analysis uniquely studied how maternal smoking, black carbon, and PFASs may interact to influence birth size, prior studies have focused on how mixtures of other pollutants may impact birth weight. One study of about 1000 European mother-infant pairs observed a stronger negative cross-sectional association between cord blood dioxin-like activity and birth weight among infants of women who smoked during pregnancy versus non-smokers, but the interaction was not significant [47]. Another study of over 1000 mother-infant pairs from Greenland, Poland, and Ukraine observed greater birth weight reductions for infants of smokers concomitantly prenatally exposed to PFOA and di(2-ethylhexyl) phthalate (DEHP) versus non-smokers exposed to these two chemicals [48]. Additionally, a study of about 250 Flemish mother-infant pairs found a larger decrement in birth weight in infants simultaneously exposed to arsenic, PFOA, lead, cadmium, and DEHP as compared to individual exposures, although they did not test for interactions between compounds [49].

The multi-pollutant model approach allows for evaluation of joint effects and is generally expected to minimize co-exposure confounding [50]. However, as is the case with single-pollutant models, misspecification due to adjustment for irrelevant or collider variables, or lack of adjustment for confounder variables could lead to bias [51, 52]. Although collinearity between exposures in a multi-pollutant model could decrease precision of effect estimates [50], this was not a limitation of the present analysis, as the exposures we studied were not tightly correlated.

Generalizability is a limitation of our cohort, as participants are primarily of moderately high socioeconomic

status and all resided in Massachusetts at enrollment. In addition, although we limited our primary analysis to early pregnancy exposures for which we have the most robust data, timing of exposure could alter independent and combined effects of these pollutants on fetal growth. Also, to target our analysis and limit multiple comparisons, we focused on one air pollutant (black carbon) and one PFAS (PFOS) in our primary analyses, although pregnant women are exposed to several types of air pollutants and PFAS concurrently. Our observation that PFNA had stronger independent and combined effects on fetal growth than PFOS suggests that specific PFASs may act differently. However, PFNA plasma concentrations have a relatively narrow range in our cohort, and thus we cannot generalize our findings to women with higher PFNA concentrations.

Strengths of our study include use of a large prospective cohort with measurements of multiple prenatal environmental exposures known to influence fetal growth. Additionally, most environmental epidemiological studies examine the effect of a single pollutant on health outcomes, despite the fact that individuals are exposed to multiple pollutants concurrently [21]. We present an approach, of examining multi-pollutant models with two-way and three-way interactions, to investigate health effects related to joint exposure to multiple toxicants.

## Conclusion

We observed that concurrent early pregnancy exposures to maternal smoking, residential black carbon, and PFOS were associated with additive reductions in fetal growth, while exposure to PFNA attenuated the associations of smoking and black carbon exposure on low birth size, in a Boston-area cohort. Because a developing fetus is exposed to mixtures of chemicals, it is critical to continue to evaluate the role of multiple pollutant exposures in future studies. Further investigation of the effect of mixtures of in utero exposures on health outcomes may provide insight into potential biological pathways of action.

## Additional file

**Additional file 1:** Supplemental Material. (DOCX 37 kb)

## Abbreviations

BMI: Body mass index; BW/GA: Birth weight-for-gestational age; CDC: Centers for Disease Control and Prevention; CI: Confidence interval; DEHP: di(2-ethylhexyl); GFR: Glomerular filtration rate; IQR: Interquartile range; LMP: Last menstrual period; LOD: Limit of detection; PFAS: Perfluoroalkyl substances; PFHxS: Perfluorohexane sulfonate; PFNA: Perfluorononanoate; PFOA: Perfluorooctanoate; PFOS: Perfluorooctane sulfonate; PM: Particulate matter; PPAR: Peroxisome proliferator-activated receptor; SD: Standard deviation

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## Availability of data and materials

The datasets used and/or analyzed during the current study are available from the corresponding author or the Project Viva study team (project\_viva@hphc.org) upon reasonable request.

## Authors' contributions

LBR and AFF conceived the study. LBR performed the statistical analyses and drafted the manuscript. SLR checked the statistical analyses. AG created the black carbon spatiotemporal regression model. AMC and XY analyzed plasma PFAS concentrations. LBR, SLR, BAC, AC, AMC, XY, AG, JS, SKS, DRG, EO, and AFF contributed to the interpretation of data and provided critical revision of the manuscript for important intellectual content. All authors read and approved the final manuscript.

## Ethics approval and consent to participate

All women provided written consent, and the Institutional Review Board of Harvard Pilgrim Health Care approved the study.

## Consent for publication

Not applicable.

## Competing interests

The authors declare that they have no competing interests.

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